# The Lancaster **Environment Centre**

Annual Report for 2011 on the UK Toxic Organic Micro-pollutants (TOMPs) Air **Monitoring and Analysis Network** 

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Report to the Department for Environment, Food and Rural Affairs, the Northern Ireland Department of Environment, the Scottish Government and the Welsh Assembly



# **Executive Summary.**

This report contains the 2011 quarter 1(Q1), quarter 2 (Q2) quarter 3 (Q3) and quarter 4 (Q4) ambient air concentration data for polychlorinated biphenyls (PCBs), polychlorinated-pdioxins (PCDDs), polychlorinated dibenzofurans (PCDFs) and polybrominated diphenyl ethers (PBDEs) from the Toxic Organic Micro-pollutants Monitoring Network (TOMPs) which is funded by the Department for Environment, Food and Rural Affairs (Defra) and the devolved administrations.

In 2011, 37 PCBs congeners, 4 co-planar PCB congeners, 22 PBDE congeners, 10 furan congeners and 7 dioxin congeners were measured in each sample. The TOMPs network includes sites in London, Manchester, Hazelrigg (Lancashire), High Muffles (North Yorkshire), Auchencorth Moss (Midlothian) and Weybourne (Norfolk). The aim of the TOMPs network, which has operated since 1991, is to provide information on the ambient levels of organic pollutants in the UK through monitoring of air concentrations at six sites. The results and other related scientific work are used to inform policy development on exposure to persistent organic pollutants (POPs).

The TOMPs network provides data to inform the public of air quality, and information to support the development of policy to protect the environment. The specific aims of the TOMPs programme are:

- To identify sources of a range of POPs in the UKs atmosphere.
- To quantify sources that are regarded as potentially significant.
- To measure concentrations of TOMPs in ambient air in UK cities, in order to assess both human exposure and the relationship between source emissions and levels in the ambient atmosphere.

There are a number of international instruments aimed at reducing releases into the environment, such as the 1998 UN/ECE Protocol on Persistent Organic Pollutants made under the Convention on Long-Range Transboundary Air Pollution, and the Stockholm Convention (SC) on POPs. The TOMPs network provides valuable evidence for effectiveness

of such agreements on the concentrations of a range of POPs in UK urban and rural ambie air.	nt

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#### 1. Introduction

Lancaster University (LU) has been involved in the TOMPs programme since its inception in 1990. LU currently manages the programme on behalf of Defra and the devolved administrations which operates six sites, three urban, two rural and one semi-rural. The current contract commenced in October 2010 and will run until the end of September 2013. Atmospheric sampling is carried out at each site, collecting a biweekly sample which is bulked to provide quarterly data. These data are reported to Defra and published on the air quality data website <a href="http://uk-air.defra.gov.uk">http://uk-air.defra.gov.uk</a>. Two sites are currently maintained via subcontracts; Auchencorth Moss by the Centre for Ecology and Hydrology (CEH) and the Weybourne Observatory by the University of East Anglia. The analytes quantified at Lancaster University are PCDD/Fs ('dioxins and furans'), PBDEs and PCBs. Polycyclic aromatic hydrocarbons (PAHs) are quantified and reported separately under another monitoring programme currently operated by the National Physical Laboratory.

This annual report for polychlorinated biphenyls (PCBs), polychlorinated-p-dioxins (PCDDs), polychlorinated dibenzofurans (PCDFs) and polybrominated diphenyl ethers (PBDEs) (Q4 only) includes:

- Information on PCBs, PCDDs, PCDFs, and PBDEs
- A summary of network operations including details of monitoring sites, equipment employed, details of site installations/removals, site calibration visits and equipment servicing and breakdowns
- A summary of the analytical procedure used to detect PCBs, PCDDs/PCDFs and PBDEs.
- Trends in estimated sources of PCBs, PCDDs and PCDFs in the UK
- Review of annual mean and quarterly concentration.

# 2. Background to PCBs, PCDDs and PCDFs and PBDEs.

**PCBs** were first synthesized in 1881 by Schmidt and Schulz but their commercial production only began in 1929 in USA (*Danse et al.*, 1997). They were marked as mixed products under various trade names depending on the country where they were produced such as Aroclor (Monosanto, USA), Phenochlor and Clophen (Bayer, EU). Because of high chemical and thermal stability, electrical resistance, low or no flammability, PCBs had extensive

applications. They have been used as dielectric fluids in capacitors and transformers, in plasticizers, adhesives, inks, sealants and surface coatings (Eduljee, 1988; de Voogt and Brinkman, 1989; Harrad et al., 1994). Their basic structure is a biphenyl backbone with one to ten chlorine substituents and a general structure of  $C_{12}H_{10-n}Cl_n$  (n=1-10).

$$Cl_n$$
  $Cl_m$ 

There are 209 different congeners with one to ten chlorines atoms attached. The International Council for the Exploration of the Seas (ICES) 7 PCB congeners generally reported in environmental samples are PCB 28 (2,4,4'-triPCB), PCB 52 (2,2',5,5'-tetraCB), PCB 101 (2,2',4,5,5'-pentaCB), PCB 118 (2,3',4,4',5-heptaCB), PCB 138 (2,2',3,4,4',5-heptaCB), PCB 153 (2,2',4,4',5,5'-heptaCB), PCB 180 (2,2',3,4,4',5,5'-heptaCB), although several dozen different congeners can be found in the environment.

Production of PCBs peaked in the 1960s in Europe and USA and terminated in the mid 1970s, when they where ultimately banned in the late 1970s/early 1980s (de Voogt and Brinkman, 1989). The most recent inventory of PCB production estimates the cumulative global production of PCBs at 1.3 million tonnes (Breivik et al., 2002). Approximately 97% of this has been used in the Northern Hemisphere, mostly between 30 °N and 60 °N (Breivik et al., 2002). Before the ban, PCBs had entered the environment through point and diffusive sources such as landfill sites, accidental releases/spillages via leaking during commercial use of electrical equipment and transformer and capacitor fires, incineration of PCB waste etc. (de Voogt and Brinkman, 1989; Danse et al., 1997). Current atmospheric levels of PCBs in the environment are can be accounted by on-going primary anthropogenic emissions (e.g. accidental release of products or materials containing PCBs), volatilization from environmental reservoirs which have previously received PCBs (e.g. sea and soil) or incidental formation of some congeners during combustion processes (Breivik et al., 2002). The National Atmospheric Emission inventory estimates that the emission of PCBs to the UK atmospheric was 760kg, the majority emitted from electrical equipment such as capacitors and transformers.

PCDD/Fs. The term Dioxin is commonly used to refer to a family of toxic chemicals that all share a similar chemical structure and a common mechanism of toxic action. This family includes seven of the polychlorinated dibenzo dioxins (PCDDs), ten of the polychlorinated dibenzo furans (PCDFs) and twelve of the polychlorinated biphenyls (PCBs). PCDDs and PCDFs are not commercial chemical products but are trace level unintentional byproducts of most forms of combustion and several industrial chemical processes. PCBs were produced commercially in large quantities until production was stopped in 1977. Dioxin levels in the environment have been declining since the early seventies and have been the subject of a number of federal and state regulations and clean-up actions; however, current exposures levels still remain a concern.

**Polybrominated Diphenyl Ethers (PBDEs)** have been widely used as additive flame retardants in products such as furniture, cars, textiles, paints, electronic equipment and plastics to reduce fire risk. They are referred as additive flame retardants, because they are simply blended with the product. This makes them more prone to volatilize into the atmosphere during the product lifetime. They reduce fire hazards by interfering with the combustion of the polymeric materials (BSEF, 2000; Commission of the European Communities, 2000). Their general structure is C<sub>12</sub>H<sub>10-n</sub>Br<sub>n</sub>O (n=1-10). Therefore, there are 209 possible PBDE congeners, depending on the position of the bromine atoms on the phenyls rings. Three different types of commercial PBDE formulation have been produced with different degrees of bromination namely penta-, octa- and deca-BDE products. The

penta-BDE product contains a range from tetra to hexa-BDE congeners, the octa-BDE contains a mixture of hexa- to deca-BDE and the deca contains predominantly the deca-BDE congener and is currently the most widely PBDE flame retardant product in use.

$$Br_m$$
— $Br_n$ 

The global demand for PBDEs has previously been very substantial with a peak estimation of 70,000 tonnes for the year 2003 (Hites *et al.*, 2004). Of these technical mixtures, the commercial pentabromodiphenyl ether (PeBDE) and commercial octabromodiphenyl ether (OctaBDE) mixtures have been banned in the EU and Japan and are currently being phased out in the rest of the world (after being banned in some states of the USA). In the UK there has been previously high use of PeBDE as a result of particularly stringent fire retardancy regulations for furniture. Lower brominated PBDEs can also be formed from the degradation of higher brominated BDEs although the environmental importance of this process is still unclear. For the first time a range of PBDE congeners have been included in the TOMPs methodology. The congeners that have been analysed are: PBDEs 28 (tri), 47 (tetra), 49 (tetra), 99 (penta), 100 (penta), 153 (hexa), 154 (hexa), 183 (hepta). Congeners BDE-47 and BDE-99 account for approximately 72% of the composition of the penta commercial mixture (pentaBDE). Atmospheric emission estimates for 2008 for the tetra and penta PBDEs in the UK were 2500kg and detailed in a recent Defra report (AEAT/ENV/R2767/WP2 ED47664).

# 3. TOMPs sites operating in 2011

In 2011 the TOMPs programme operated 6 sites:

London (LON)	urban site established in 1991
Manchester (MAN)	urban site established in 1991
Hazelrigg (HR)	semi-rural site established in 1992
High Muffles (North Yorkshire) (HM)	rural site established in 1999
Auchencorth Moss (AC)	rural site established in 2008
Weybourne (WE)	rural coastal site established at the end of 2008

The sites consist of two urban locations in London (LON), Manchester (MAN), three rural sites at High Muffles (HM, North Yorkshire), Auchencorth Moss (AC, Mid Lothian) and Weybourne (Norfolk), one semirural site at Hazelrigg (HR, Lancashire). At the rural and semirural sites, samplers are located away from major roads, whereas at the urban sites samplers are located in the city centre on the roof of a building. The locations of the current samplers in the network are shown in Figure 1.



Figure 1. Location map of the current TOMPs sites.

#### 4. Network sampling operations

The sampling modules for the Andersen GPS-1 sampler are prepared just prior to deployment which involves disassembling, inspecting and cleaning the modules. Modules are stored frozen in sealed bags prior to deployment. All parts of the modules that come into contact with the glass fibre filter (GF/A Whatman) and polyurethane plugs (PUFs, Klaus Ziemer GmbH Langerwehe, Germany) are routinely solvent cleaned between each sample. In addition, the modules are fully disassembled and all parts thoroughly cleaned in solvent. The GF/As are pre-cleaned by baking out in a muffle furnace at 450 °C for 24 hours. They are then transferred to aluminium foil packages (the aluminium foil has also been baked out) and stored sealed until they are used. PUFs and GF/As filters are regularly sent to CEH in Edinburgh who manage the Auchencorth Moss site and University of East Anglia who manage the Weybourne site. The PUFs are prepared for all the sites from the same batches, by a rigorous pre-extraction procedure. This involved a soxhlet extraction in acetone/hexane (1:1), with subsequent solvent removal in a solvent cleaned desiccator, maintained under

vacuum. PUFs are also prepared to serve as field and laboratory blanks. The GF/As and PUFs are placed in the sampling modules using solvent cleaned stainless steel tongs and are exposed to the laboratory environment for the minimum amount of time possible.

The modules are changed every 14 days at all sites. In addition, sample information and temperature data are recorded, airflows adjusted, data loggers exchanged and preventative maintenance carried out when necessary. The time during which the sampler operates is recorded with a timer, and the flow rate determined using the flow venturi and MagnaHelic gauge. Each sampler is also fitted with a pressure transducer and a data logger that records the pressure drop during the sampling period, so that the sampling rate can be accurately determined. Log books are used to record sampling data at each site, but sampling data are also available electronically. The following are recorded routinely for each sample at each of the sites: start time, date, counter reading, MagnaHelic reading; stop time, date, counter reading, MagnaHelic reading; maximum, minimum and actual temperature (°C). Crosschecks are possible between the manually calculated air volume and the electronically calculated air volume. During each visit, the sampler, sampler platform and auxiliary pieces of equipment are checked for corrosion or breakages. A number of spare parts are routinely taken to each site and preventative or remedial maintenance carried out when necessary. Long life brushless motors are used to minimise samples lost through motor failures. A sampler calibration is performed once a year at each site.

### 5. Extraction and clean-up procedures

Preparation of the samples takes place in a laboratory with restricted access. All glassware is thoroughly solvent cleaned prior to use and where necessary baked out at 450 °C overnight following established procedures. Each sample (gas + particle) is spiked with a recovery standard of <sup>13</sup>C<sub>12</sub>-labeled PCB congeners (<sup>13</sup>C<sub>12</sub> PCB 28, 52, 101, 138, 153, 180, 209) and PBDE congeners BDE 51, BDE 128, and BDE 190, and an isotope dilution/recovery standard containing 21 <sup>13</sup>C<sub>12</sub>-labelled PCDD/Fs and coplanar PCBs. Samples are individually extracted in a Buchi extraction unit for 18 hours with hexane and 6 hours with toluene. PAHs, PCBs, PBDEs and tri, tetra and penta PCDD/Fs are extracted in the hexane fraction. The remaining PCDD/Fs are extracted in the toluene fraction. The extracts are concentrated using rotary-evaporation and nitrogen-evaporation. The hexane and toluene fraction are combined for each sample and extracts pooled before purification to obtain quarterly data

(Jan-March (Q1), April-June (Q2), July-Sept (Q3), and Oct-Dec (Q4)). The 6 or 7 hexane fractions (depending on the length of each quarter) of each quarter are then bulked together. The samples are transferred into a 250ml round bottom flask using hexane. If necessary this can then be rotary evaporated to 2ml for splitting. The toluene fractions are then bulked in the same way using hexane. Each quarter will consist of 6-7 two week samples, representing approximately 4500 m<sup>3</sup> of air. The hexane fraction (topped up to 50 ml with hexane) is then split: 10% (5mL) is used for the PAHs analysis, 40% (20 mL) for the analysis and 50% (25 mL) is archived. The toluene fraction (also topped up to 50 mL using hexane) is also split: 10% (5mL) is discarded, 40% (20 mL) is analyzed and 50% (25 mL) is archived. (The toluene and hexane fractions for archive are combined in the same vials). The same is done for the fractions (40%) that will be analyzed. The extracts are then eluted through a multilayer 20 mm inner diameter (id) acid silica column containing a small layer of sodium sulphate, 1 g activated silica (Merck Silica 60), 2 g of basic silica (Merck Silica 60), 1 g of activated silica (Merck Silica 60, 4 g of acid silica (Merck Silica 60), 1 g activated silica and a small layer of sodium sulphate (silica and sodium sulphate baked at 450°C overnight) followed by two times acid digestion using concentrated H<sub>2</sub>SO<sub>4</sub> and a second multicolumn. The extracts are eluted through gel permeation columns containing 6 g of Biobeads SX 3 and concentrated to 500 µL. Each sample is then fractionated with a basic alumina column to obtain three fractions. Fraction 1 contains PCBs and PBDEs, Fraction 2 contains co-planar PCBs and Fraction 3 contains PCDD/Fs. Fraction 1 containing PCBs and PBDEs is solvent exchanged to 160 mL of dodecane (for urban site) and 80 mL of dodecane (for the more remote sites) containing PCB 30 [13C<sub>12</sub>], PCB 141, [13C<sub>12</sub>] PCB 208, BDE 69, and BDE 181 as internal standards. The PCB and PBDE fractions are analyzed by gas-chromatography mass spectrometry (GC-MS) with an EI+ source operating in selected ion mode (SIM). Details of the instruments, temperature programme and monitored ions are given elsewhere (Thomas et al., 1998 and Gouin et al., 2002). Thirty-seven PCB congeners and 22 PBDE congeners are constantly measured in all samples, but only the following congeners are reported: PCBs 28, 52, 90/101, 118, 138, 153/132 and 180. Some congeners co-elute and are hence reported as a pair, for example, 153/132. Fractions 2 and 3 are solvent exchanged to 15 mL of nonane containing an injection standard of <sup>37</sup>Cl-labeled 2,3,7,8-TCDD. Analysis is performed on a Micromass Autospec Ultima high resolution-mass spectrometry (HR-MS) operated at a resolution of at least 10,000. Dioxins, furans and co-planar PCBs are generally found in mixtures containing several kinds of dioxins and dioxin-like compounds, each

having its own degree of toxicity. To express the overall toxicity of such a mixture as a single number, the concept of "Toxic Equivalents" (TEQ) has been developed. The concentration of co-planar PCBs and PCDD/Fs are expressed in units of fgTEQm<sup>-3</sup>. The concentration in fgm<sup>-3</sup> is multiplied by the WHO Toxic equivalency factors (TEF) to obtain the final concentration in fgTEQm<sup>-3</sup>. The WHO TEF scheme used for the data conversion was developed in 1998, and although the scheme was updated in 2005, the original scheme is still used to ensure consistency within the dataset. From Q4 2010 the following PBDE congeners have been reported; BDE 17, 28, 32, 35, 37, 47, 49, 66, 71, 75, 77, 85, 99, 100, 119, 138, 153, 154, 166, 183, 196.

QA/QC A number of steps are taken to obtain data that would allow an assessment of the accuracy and reliability of the data. PCB and PBDE recoveries are monitored by quantifying 10 \(^{13}\text{C}\_{12}\)-labelled PCB and PBDE standards and they ranged between 61-101%. PCDD/F and coplanar PCB values are corrected using 21 \(^{13}\text{C}\_{12}\)-labelled PCDD/F and coPCB isotope dilution standards, using the injection standard as an internal standard. The criteria for the quantification of analytes are a retention time found within 2s of the standard, isotope ratio found within 20% of standard and a signal to noise ratio of at least 3. Analytical blanks, consisting of solvent are included at a rate of one blank for every 12 samples. The method detection limit was calculated as 3 times the standard deviation of the concentrations found in the analytical blanks. If the concentrations in the blanks are below the instrumental detection limit, then the method detection limit is defined as equal to the instrumental detection limit. All results are blank corrected using the concentration of the field blanks. Field blanks are produced for each site and each quarter and they are used to calculate method detection limits (MDLs). When compounds are not detected in the field blanks, laboratory blanks produced for each quarter are used to estimate MDLs.

# 6. Data storage.

The data are reported to Defra and published on the UK air quality website (<a href="http://uk-air.defra.gov.uk/">http://uk-air.defra.gov.uk/</a>). Archived samples for each year (50% of the samples) are stored in the freezer in the laboratory at Lancaster University.

#### 7. RESULTS FOR YEAR 2011

7.1 Network Operations: Table 1 contains information on the samples collected, including, bulked air volume (in m³) and the number of samples bulked for each site for each quarters in 2011. The bulked air volume is obtained by summing the volume (in m³) obtained from each sample taken during the quarter (usually 6-7 samples depending on the sampling schedule). Total volume per quarter (bulked volume) under normal operating conditions ranged from 3900-5500 m³, with the exception of Q1 at Weybourne (see Table 1).

London and Auchencorth operated normally with 100% data capture over the year. The sampler at Manchester experienced motor/equipment failures during Q1/Q2 with and overall operating efficiency of 90%. The coastal site at Weybourne also suffered from a series of power supply and motor failures over the early part of 2011. This resulted in a reduced efficiency during this period with data capture rates of 10% and 82% for Q1 and Q2. These problems, resulting from the corrosive maritime atmosphere, have now been resolved and the site was operated at 100% efficiency for the rest of the year. The sites at Hazelrigg and High Muffles operated with an annual efficiency of 98% and 97% efficiency, experiencing only minor disruption.

The sites at Hazelrigg, Manchester and Weybourne were calibrated in Q2 and the sites at London, Auchencorth and High Muffles were calibrated in Q3. Quarterly field blanks have also been collected from each site. The Manchester sampler was replaced in June 2011 with a refurbished Hi. Vol. after an upgrade of the electricity supply at Manchester Law Courts. This completes the replacement of each TOMPs air sampler site with refurbished equipment.

In this report data for Q1, Q2, Q3 and Q4 2011 are presented for PCBs, co-planar PCBs and PCDDs and PCDFs and PBDEs.

**Table 1**. Summary of the bulked air volumes and number of samples for each quarter at all sites in 2011.

		Start	Time	Finish	Time	% Data capture
LON	Q1	07/01/2011	12:50:00	06/04/2011	11:45:00	100.0%
	Q2	06/04/2011	11:50:00	30/06/2011	10:40:00	100.0%
	Q3	30/06/2011	10:45:00	06/10/2011	11:35:00	100.0%
	Q4	06/10/2011	11:40:00	05/01/2012	12:00:00	100.0%
MAN	Q1	30/12/2010	14:10:00	07/04/2011	15:55:00	76.%
	Q2	07/04/2011	15:55:00	30/06/2011	11:00:00	83.33%
	Q3	30/06/2011	11:05:00	06/10/2011	10:45:00	100.0%
	Q4	06/10/2011	10:50:00	29/12/2011	10:35:00	100.0%
HR	Q1	30/12/2010	15:15:00	06/04/2011	14:55:00	96.9%
	Q2	06/04/2011	14:55:00	29/06/2011	15:50:00	100.0%
	Q3	29/06/2011	15:55:00	04/10/2011	16:25:00	100.0%
	Q4	04/10/2011	16:40:00	28/12/2011	15:30:00	97.7%
НМ	Q1	04/01/2011	11:15:00	07/04/2011	09:35:00	100.0%
	Q2	07/04/2011	09:40:00	30/06/2011	13:30:00	100.0%
	Q3	30/06/2011	13:35:00	06/10/2011	13:15:00	100.0%
	Q4	06/10/2011	13:20:00	29/12/2011	13:30:00	91.7%
AUCH	Q1	05/01/2011	12:00:00	30/03/2011	12:00:00	100.0%
	Q2	30/03/2011	12:00:00	06/07/2011	10:45:00	100.0%
	Q3	06/07/2011	10:45:00	28/09/2011	10:10:00	100.0%
	Q4	28/09/2011	10:10:00	04/01/2011	10:55:00	100.0%
WEY	Q1	04/01/2011	10:10:00	01/04/2011	10:20:00	10.5%
	Q2	01/04/2011	10:20:00	04/07/2011	09:20:00	81.9%
	Q3	04/07/2011	09:20:00	26/09/2011	09:20:00	100.0%
	Q4	26/09/2011	09:20:00	29/12/2011	09:35:00	100.0%

#### 7.2 PCDD/Fs: Results and discussion

Quarterly PCDD/F data are contained in the Appendix 1 and a summary presented in Figure 2. The annual mean PCDD/Fs concentrations measured in 2011 ranged from 0.05 to 12.5 fg TEQ/m³ and are similar to those we have reported for last 10 years. The highest concentrations were observed in the two urban sites (Manchester 12.5 fg TEQ/m³ and London 3.7 fg TEQ/m³) followed by Weybourne (2.5 fg TEQ/m³), Hazelrigg (2 fg TEQ/m³), High Muffles (1 fg TEQ/m³) and Auchencorth (0.05 fg TEQ/m³). As with previous years, the two urban sites exhibit higher concentrations of PCDD/Fs although the average concentration in London was closer to the sites at Weybourne and Hazelrigg. The seasonal pattern as shown by the quarterly data shows that the highest concentrations were measured in quarters Q1 and 4 with lower values reported for quarters 2 and 3.

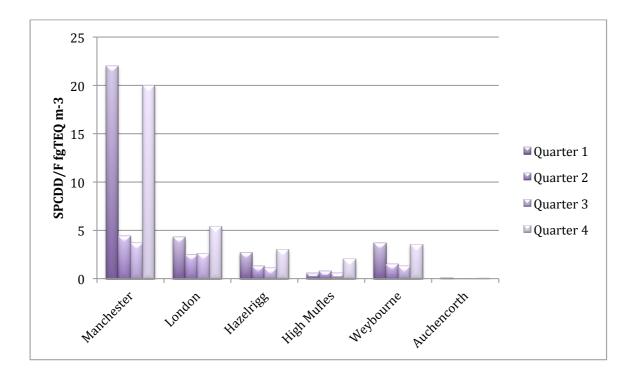
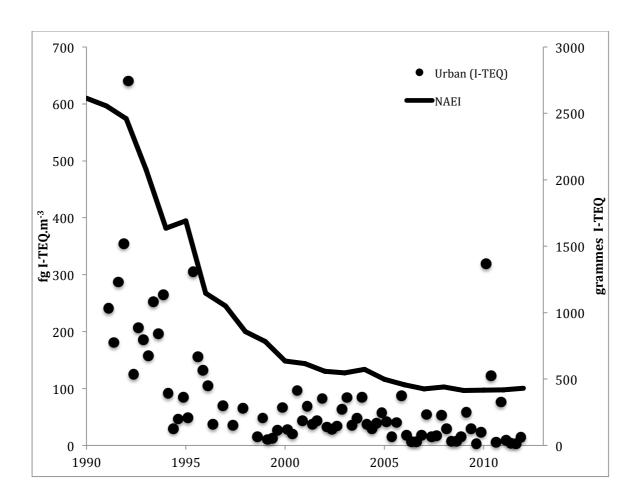


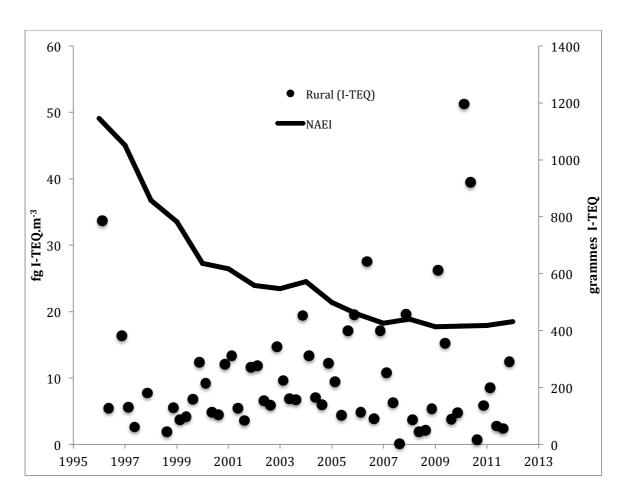
Figure 2. Quarterly PCDD/F data for each TOMPs site in 2011.

PCDD/Fs constitute two classes of chemicals that are formed unintentionally during combustion (e.g. waste incineration, burning of coal, wood etc.), the refining of petroleum, several metal treatment processes and during the synthesis of certain chlorinated chemicals (e.g. pentachlorophenol). The latter processes have reduced in importance over time. The observed seasonality of PCDD/Fs in air, where winter values exceed summer values, has

been seen in the TOMPs dataset and is widely reported in the literature, for example, Coutinho *et al.* (2007) reported average summer/winter ratios of 1:3 for sites in Portugal. Increased combustion (i.e., domestic space heating) and seasonal variations in the atmospheric boundary layer height are the likely causes. Generally seasonality is less obvious in urban areas than rural, as cities have more constant sources that emit throughout the year. The sources of PCDD/Fs to the UK atmosphere are presented in detail by Katsogiannis *et al.*, 2010 which discusses the TOMPs programme over the last 17 years.

The temporal trends for the urban and rural sites are presented in Figure 3, and the estimated atmospheric half lives (assuming first order kinetics) for London and Manchester are 4.8 and 5.2 years, respectively. The concentration data from Hazelrigg and High Muffles are much lower than the urban sites, but do not show a significant decrease over time. The Auchencorth Moss and Weybourne data sets are currently too short to determine any temporal trends. Data for UK PCDD/F emissions are provided by the National Atmospheric Emission Inventory over the period 1990 - 2011 (http://www.naei.defra.org). During this period PCDD/F emissions reduced from around 1100 g TEQ per year in 1990 to ~225 g TEQ per year in 2000 and further to ~200 g TEQ per year in 2011. The TOMPS program started in the same year (1990) and shows air concentrations generally falling over that period, although there is good evidence that this is actually part of a longer and more systematic decline. Media which are broadly reflective of trends in air – such as herbage and milk – show higher levels in the 1980s than 1990s (Hassanin et al., 2006; Alcock et al., 1996, 1998, Kjeller et al., 1996), and higher levels still in the 1960s and 1970s. Human dietary intakes of PCDD/Fs are estimated to have declined 4-5 fold between 1982 and 1992 (Van der Gon et al., 2005) and to have declined still further since (UKFSA et al., 2006).

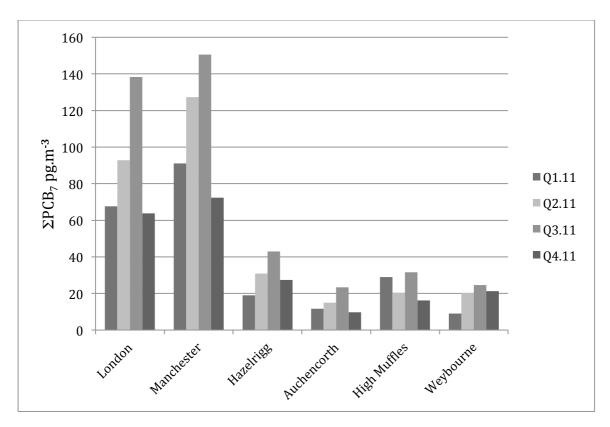




**Figure 3.** Long-term PCDD/F trend data from the TOMPs network for the urban (London and Manchester) and the rural (Hazelrigg, High Muffles, Auchencorth and Weybourne) sites. Data are presented as averages in fg I-TEQ.m<sup>-3</sup> and compared to the current NAEI estimates in grammes I-TEQ per annum

#### 7.3 PCBs: Results and discussions

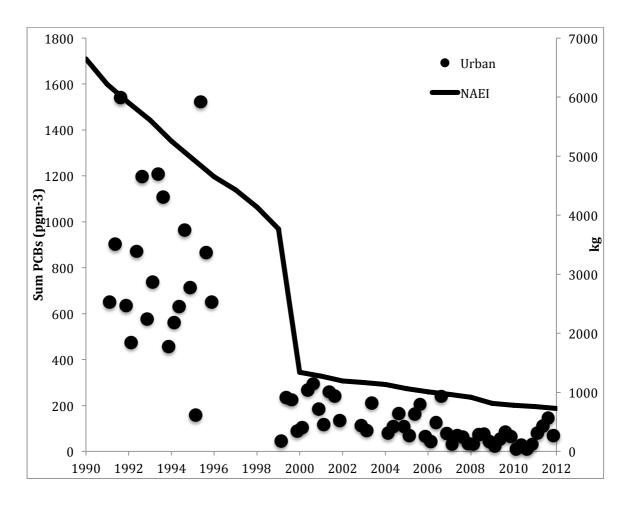
Quarterly congener PCB data are contained in the Appendix 2 and a summary presented in Figure 4. The quarterly PCB concentrations measured at each of the TOMPs sites ranged from 15 pg/m³ (Auchencorth) to 110 pg/m³ (Manchester) for the sum of seven indicator PCB congeners (PCBs 28,52,101,118,138,153,180). The data shows, as with previous years, that concentrations are proportional the population density i.e. higher for urban sites by a factor of 5. The urban sites at London and Manchester showed an increase in the PCB concentrations compared with previous average values over the last five years (2006-2010) of 60%, whilst the annual average concentrations for all the other sites (Hazelrigg, Weybourne, High Muffles and Auchencorth) were also slightly higher with a 30% increase. At each site the quarterly data showed a distinct seasonal pattern with higher levels in Q2 and Q3 which are characteristic of temperature driven diffusive sources.



**Figure 4.** Quarterly  $\Sigma PCB_7$  data at the TOMPs site for 2011.

Ambient PCBs concentrations are controlled by a range of factors but primarily by proximity of sampling sites to on-going sources. These sources are generally differentiated into primary and secondary sources. Primary sources of PCBs, which are mostly diffuse, include articles or preparations to which they were added, for example, as plasticizers in plastics, sealants, paints and oils. As a result of the application pattern for PCBs in indoor environments, primary sources are mostly found in areas with high population density and hence generally remain higher in urban environments. Emissions from secondary sources describes the process of re-emission or volatilization of PCBs from environmental compartments like soil and sediments which serve as reservoirs for persistent organic chemicals. Generally, the urban sites such as London and Manchester are still influenced by on-going diffuse primary releases of PCBs, whilst more rural sites are influenced by secondary sources and atmospheric transport. A detailed discussion of PCB sources to the UK atmosphere are discussed in detail by Schuster et al., 2010 which discusses the TOMPs programme over the period 1991 to 2008. With the addition of the 2011 data, the clearance rates (time taken for a 50% decline in concentration) provided by the TOMPs network averages 5 years with a range of 3.9 for London to 7.5 yrs for High Muffles across the sites. Time trend data for the urban

sites (London, Manchester) and rural sites (Hazelrigg, High Muffles, Auchencorth and Webourne) are presented in Figure 5. These data have been plotted against the estimated UK emission data from the NAEI (<a href="http://www.naei.defra.org">http://www.naei.defra.org</a>).



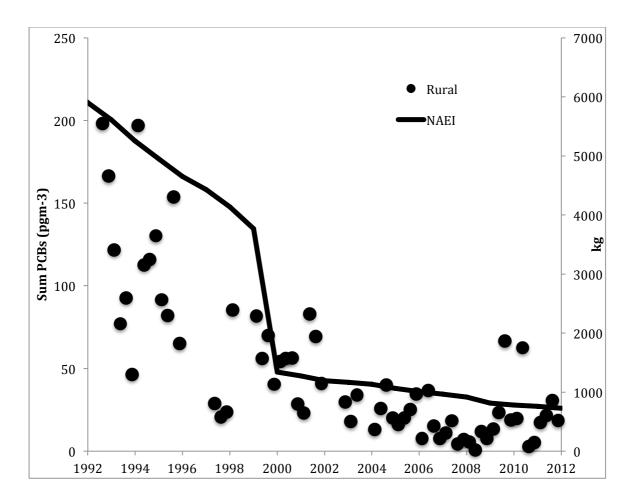
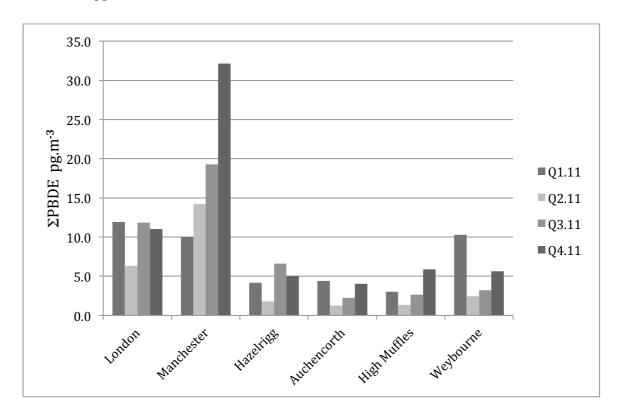


Figure 5. Long-term  $\Sigma_7PCB$  trend data from the TOMPs network for the urban (London and Manchester) and the rural (Hazelrigg, High Muffles, Auchencorth and Weybourne) sites. Data are presented as averages in pg m<sup>-3</sup> and compared to the current NAEI estimates in kg per annum

#### 7.4 PBDEs: Results and discussions

This year is the first complete year for the inclusion of PBDEs in the TOMPs network. Twenty two individual congeners have been measured and the data reported in Appendix 3. The most prominent congeners, accounting for between 23% and 91% of the  $\Sigma_{22}$ PBDEs were BDE47 and BDE99 (mean value 50%). These congeners were prominent in the commercial pentaBDE mix, accounting for 72% of the total. London and Manchester showed the highest annually averaged concentrations of  $\Sigma_{22}$ PBDEs at 10.3 pg/m³ and 18.9 pg/m³, respectively. The other sites were lower at 4.4, 3.0, 3.2, and 5.4 pg/m³ for Hazelrigg, Auchencorth, High Muffles and Weybourne, respectively. In 2004 Lee *et al.* carried out an ambient air sampling campaign at Hazelrigg, Chilton (Oxfordshire) and Mace Head (Galway) with 40 samples collected at each site. They reported BDE47 plus BDE99 concentrations ranging from 1.9 pg/m³ for Mace Head (a European background site) to 6.1 pg/m³ for Chilton and 7.9 pg/m³ for Hazelrigg. These data are similar to those measured at the TOMPs sites.

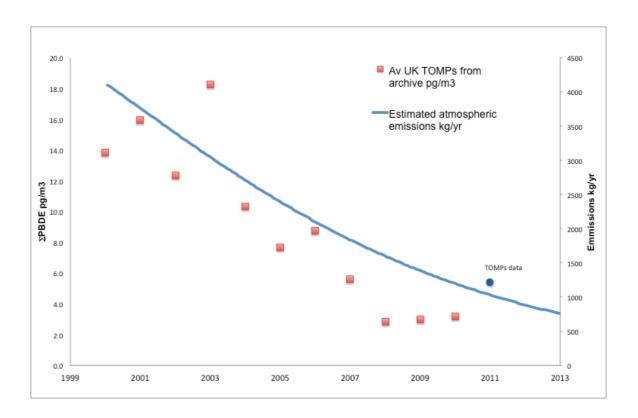


The TOMPs air sample archive has recently been used to provide information the time-trend of pentabrominated diphenylethers in the UK atmosphere. The re-analysis of PBDEs in the sample archive has focused on four of the six sites over a period ranging from 1999 to 2010. These time-trend data demonstrate a consistent decrease in concentration over recent the last years with the observed decline starting during the period 2001-2003. This is particularly

evident in the urban sites of Manchester and London and at the semi-rural site of Hazelrigg. The average ΣPBDE half-lives for these three sites were 3.4, 2.0 and 3.5 years, respectively. PBDEs are present in background levels, among the lowest reported in literature. Comparison of concentrations to estimated emissions and employment of PBDE profiles suggest that PBDEs in the UK atmosphere originate from primary emissions from products that contain mainly the penta-BDE technical mixture. The direct connection with source reduction and UK ambient air concentrations has been established using the EvnBETR model and hence these data provide useful information on the effectives of international agreements such as the Stockholm Convention and UN/ECE POPs protocol.

A manuscript covering the temporal trends of PBDEs in the UK environment has been recently published in Environmental Pollution.

Askin Birgul, Athanasios Katsoyiannis, Rosalinda Gioia, John Crosse, Mark Earnshaw, Nuno Ratola, Kevin C. Jones, Andrew J. Sweetman (2012) Atmospheric polybrominated diphenyl ethers (PBDEs) in the United Kingdom Original Research Article. Environmental Pollution, 169, Pages 105-111



#### 8 REFERENCES

- 1. Alcock, R. E.; Johnston, A. E.; McGrath, S. P.; Berrow, M. L.; Jones, K. C., Longterm changes in the polychlorinated biphenyl content of United-Kingdom soils. Environmental Science & Technology 1993, 27, (9), 1918-1923.
- 2. Breivik, K., Sweetman, A., Pacyna, J. M., Jones, K. C. 2002. Towards a global historical emission inventory for selected PCB congeners a mass balance approach: 2. Emissions. Science of Total Environment 290, 199-224.
- 3. Coutinho, M.; Pereira, M.; Borrego, C. Monitoring of ambient air PCDD/F levels in Portugal. Chemosphere 2007, 67, 1715–1721.
- 4. Danse, I. R., Jaeger, R. J., Kava, R., Kroger, M., London, W. M., Lu, F. C., Maickel, R. P., McKetta, J. J., Newwell, G. W., Shindell, S., Stare, F. J., Whelan, E. M. 1997. Position paper of the American Council on Science and Health: Public health concerns about environmental polychlorinated biphenyls (PCBs), Ecotoxicology and Environmental Safety, 38, 71-84.
- 5. Eduljee, G.H. 1988. PCBs in the environment. Chemistry in Britain, 24, 241-244.
- 6. Gouin T.; Thomas G. O.; Cousins I.; Barber J.; Mackay D. and Jones K. C. Airsurface exchange of polybrominated diphenyl ethers and polychlorinated biphenyls. Environ. Sci. Technol., 2002, 38, 1426-1434.
- 7. Hites, R.A. (2004) Polybrominated Diphenyl Ethers in the Environment and in People: A Meta-Analysis of Concentrations. Environ. Sci. Technol., 38 (4), 945–956
- 8. Hassanin, A.; Lee, R. G. M.; Johnston, A. E. and Jones, K. C. Reductions and changing patterns of ambient PCDD/Fs in the UK: Evidence and implications. *Chemosphere* **2006**, *65*, 530-539.
- 9. Harrad, S. J., Sewart, A. P., Alcock, R., Boumphrey, R., Burnett, V., Duarte-Davidson, R., Halsall, C., Sanders, G., Waterhouse, K., Wild, S. R., Jones, K. C. 1994. Polychlorinated biphenyls (PCBs) in the British environment: Sinks, sources and temporal trends. Environmental Pollution, 85, 131-146.
- 10. Kjeller, L. O.; Jones, K. C.; Johnston, A. E. and Rapper, C. Evidence for a decline in Atmospheric Emissions of PCDD/Fs in the U.K. *Environ. Sci. Technol.* **1996,** *30*, 1398 1403.
- 11. Katsoyiannis, A., Gioia, R., Sweetman, A.J. and Jones, K.C. (2010) Continuous Monitoring of PCDD/Fs in the UK Atmosphere: 1991-2008 Environ. Sci. Technol., 44, 5735–5740
- 12. Lee, R.G.M., Thomas, G.O. and Jones, K.C. (2004) PBDEs in the Atmosphere of Three Locations in Western Europe. Environ. Sci. Technol., 38 (3), 699–706
- 13. Publication Date (Web): December 30, 2003 (Article)
- 14. DOI: 10.1021/es035042c
- 15. Schuster, J., Gioia, R., Sweetman, A.J. and Jones, K.C. (2010) Temporal Trends and Controlling Factors for Polychlorinated Biphenyls (PCBs) in the UK Atmosphere (1991-2008). Environmental Science and Technology, 44, 8068-8074
- 16. Thomas, G. O.; Sweetman, A. J.; Parker, C. A.; Kreibich, H.; Jones, K. C. Development and validation of methods for the trace determination of PCBs in biological matrices. Chemosphere, 1998, 36, 2447-2459.
- 17. Van der Gon, D. H. A. C.; van het Bolscher, M.; Visschedijk, A, J, H.; Zandveld, P. Y. J. TNO report: Study to the effectiveness of the Persistent Organic Pollutants

- UNECE Protocol and cost of additional measures. Phase I. Estimation of the emission reduction resulting from the implementation of the POPs protocol. 2005.
- 18. de Voogt, P., Brinkman, U. A. T. 1989. Production, properties and usage of polychlorinated biphenyls. In: Kimbrough R. D., Jensen A A (Eds). Halogenated biphenyls, terphenyls, naphthalenes, debenzodioxins and related products. Elsevier-North, Amsterdam, Holland. Safety, 38, 71-84.
- 19. UKFSA. Dioxins and Dioxin-Like PCBs in farmed and wild fish and shellfish. Report 0306, 2006, at http://www.food.gov.uk/science/surveillance/.

Appendix 1. PCDD/Fs data. (All data in fgTEQ.m<sup>-3</sup>)

MANCHESTER 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	4.5	3.1	3.2	11	5.5
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	<2.2	< 2.2	< 2.2	<2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	16	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	1.8	0.45
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	3.3	0.84
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	3.5	0.87
1,2,3,4,6,7,8-HpCDD	1.3	1.3	0.47	2.4	1.4
OCDD	0.036	0.016	0.012	0.059	0.031
ΣTEQ dioxins and furans	22	4.4	3.7	20	13
3,4,4'5-TetraCB (PCB_81)	0.042	0.064	0.068	0.042	0.054
3,3',4,4'-TetraCB (PCB_77)	0.0053	0.0051	0.011	0.0076	0.0073
3,3',4,4',5-PentaCB (PCB_126)	2.9	3.5	3.3	3.1	3.2
3,3',4,4',5,5'-HexaCB					
(PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

LONDON 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	2.9	2	2.1	2.9	2.5
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	< 2.2	< 2.2	< 2.2	<2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	1.4	< 0.44
1,2,3,4,6,7,8-HpCDD	1.3	0.56	0.47	1.1	0.86
OCDD	0.026	0.012	0.014	0.029	0.02
ΣTEQ dioxins and furans	4.3	2.5	2.6	5.4	3.7
3,4,4',5-TetraCB (PCB_81)	0.029	0.046	0.054	0.032	0.04
3,3',4,4'-TetraCB (PCB_77)	< 0.004	< 0.004	0.0067	0.0054	0.0042
3,3',4,4',5-PentaCB (PCB_126)	< 0.44	< 0.44	2.2	< 0.44	0.54
3,3',4,4',5,5'-HexaCB (PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

HAZELRIGG 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	2	1	1.1	2.4	1.6
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	< 2.2	< 2.2	< 2.2	<2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDD	0.64	0.23	< 0.044	0.63	0.38
OCDD	0.016	0.0068	0.0062	0.018	0.012
ΣTEQ dioxins and furans	2.7	1.3	1.1	3	2
3,4,4',5-TetraCB (PCB_81)	0.011	0.014	0.018	0.01	0.013
3,3',4,4'-TetraCB (PCB_77)	< 0.004	< 0.004	< 0.004	< 0.004	< 0.004
3,3',4,4',5-PentaCB (PCB_126)	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
3,3',4,4',5,5'-HexaCB (PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

AUCHENCORTH 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	1.5	< 0.44	0.6	0.79	0.71
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	< 2.2	< 2.2	< 2.2	<2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDD	0.37	< 0.044	< 0.044	0.21	0.15
OCDD	0.011	0.0074	< 0.004	0.0066	0.0063
ΣTEQ dioxins and furans	0.11	< 0.01	0.035	0.059	0.052
3,4,4',5-TetraCB (PCB_81)	0.0067	0.0076	0.0073	< 0.004	0.0063
3,3',4,4'-TetraCB (PCB_77)	< 0.004	< 0.004	< 0.004	< 0.004	< 0.004
3,3',4,4',5-PentaCB (PCB_126)	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
3,3',4,4',5,5'-HexaCB (PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

HIGH MUFFLES 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	< 0.44	0.8	0.54	2	0.83
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	< 2.2	< 2.2	< 2.2	<2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDD	0.54	< 0.044	< 0.044	0.56	0.27
OCDD	0.023	< 0.004	< 0.004	0.015	0.0094
ΣTEQ dioxins and furans	0.56	0.8	0.54	2	0.97
3,4,4',5-TetraCB (PCB_81)	0.0086	0.013	0.0084	0.0054	0.0089
3,3',4,4'-TetraCB (PCB_77)	< 0.004	< 0.004	< 0.004	< 0.004	< 0.004
3,3',4,4',5-PentaCB (PCB_126)	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
3,3',4,4',5,5'-HexaCB (PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

WEYBOURNE 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
2,3,7,8-TCDF	3.7	1.3	1.3	2.6	2.2
1,2,3,7,8-PeCDF	< 0.22	< 0.22	< 0.22	< 0.22	< 0.22
2,3,4,7,8-PeCDF	< 2.2	< 2.2	< 2.2	< 2.2	< 2.2
1,2,3,4,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
2,3,4,6,7,8-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,7,8,9-HxCDF	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
1,2,3,4,7,8,9-HpCDF	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044
OCDF	< 0.0004	< 0.0004	< 0.0004	< 0.0004	< 0.0004
2,3,7,8-TCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,7,8-PeCDD	<4.4	<4.4	<4.4	<4.4	<4.4
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,6,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,7,8-HxCDD	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
1,2,3,4,6,7,8-HpCDD	< 0.044	0.19	< 0.044	0.81	0.25
OCDD	< 0.004	0.0066	0.006	0.022	0.0086
ΣTEQ dioxins and furans	3.7	1.5	1.3	3.5	2.5
3,4,4',5-TetraCB (PCB_81)	< 0.004	0.01	0.01	0.01	0.0078
3,3',4,4'-TetraCB (PCB_77)	< 0.004	< 0.004	< 0.004	< 0.004	< 0.004
3,3',4,4',5-PentaCB (PCB_126)	< 0.44	< 0.44	< 0.44	< 0.44	< 0.44
3,3',4,4',5,5'-HexaCB (PCB_169)	< 0.044	< 0.044	< 0.044	< 0.044	< 0.044

Appendix 2. TOMPs 2010 PCB data (All data in pg.m<sup>-3</sup>)

<b>LONDON 2011</b>	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011	
PCB_18	11.96	13.32	57.05	25.89	27.05	
PCB_22	15.67	21.18	20.80	8.27	16.48	
PCB_44	9.06	9.41	15.96	7.48	10.48	
PCB_49	34.91	17.38	10.19	4.63	16.78	
PCB_52	11.61	17.47	28.25	16.13	18.37	
PCB_70	5.07	7.31	12.67	5.88	7.73	
PCB_74	3.88	4.76	6.09	3.82	4.64	
PCB_87	2.92	4.70	4.64	2.37	3.66	
PCB_95	7.74	10.80	15.65	6.75	10.23	
PCB_99	2.29	3.48	3.88	2.30	2.99	
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_105	< 0.02	2.24	1.45	0.54	1.41	
PCB_110	5.51	8.37	8.67	5.19	6.94	
PCB_114	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_118	2.03	3.99	4.41	2.02	3.11	
PCB_123	< 0.02	0.63	0.49	< 0.02	0.56	
PCB_138	2.76	3.72	3.23	1.89	2.90	
PCB_141	0.69	1.07	0.97	0.59	0.83	
PCB_149	4.16	6.13	7.64	2.84	5.19	
PCB_151	1.80	2.31	3.16	1.67	2.23	
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_156	< 0.02	< 0.02	0.26	0.31	0.29	
PCB_157	< 0.02	0.09	0.47	0.19	0.25	
PCB_158	0.17	0.24	0.23	0.14	0.19	
PCB_167	0.28	0.20	< 0.02	< 0.02	0.24	
PCB_170	0.30	0.31	0.37	0.25	0.31	
PCB_174	0.70	0.88	0.86	0.40	0.71	
PCB_180	0.89	1.22	1.19	0.69	1.00	
PCB_183	0.50	0.65	0.53	0.27	0.49	
PCB_187	1.18	1.58	1.49	0.77	1.25	
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_194	< 0.02	0.17	< 0.02	< 0.02	0.17	
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02	
PCB_203	0.10	0.25	0.23	0.12	0.17	
PCB_153+132	4.38	6.63	7.08	3.23	5.33	
PCB_31+28	45.34	59.92	84.58	35.15	56.25	
PCB_41/64	1.48	3.45	19.35	21.56	11.46	
PCB_60/56	17.10	11.75	5.70	4.16	9.68	
PCB_90/101	2.65	3.86	14.07	6.77	6.84	
ΡCΒ_Σ7ΡCΒ	67.64	92.82	138.40	63.86	90.68	

MANCHESTER 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
PCB_18	10.50	11.56	36.11	16.97	18.79
PCB_22	12.09	16.77	7.82	3.53	10.05
PCB_44	10.98	16.85	15.87	7.74	12.86
PCB_49	43.21	47.95	8.69	4.48	26.08
PCB_52	20.16	31.60	38.08	23.99	28.46
PCB_70	9.35	12.51	20.18	8.61	12.66
PCB_74	6.85	8.13	5.24	3.20	5.86
PCB_87	8.25	11.78	16.18	7.23	10.86
PCB_95	20.13	24.49	44.26	17.73	26.65
PCB_99	5.82	8.00	10.98	5.82	7.65
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_105	2.93	5.41	3.32	2.00	3.42
PCB_110	14.90	20.25	27.19	13.98	19.08
PCB_114	0.00	0.00	0.30	< 0.02	0.10
PCB_118	6.78	10.62	12.92	6.41	9.18
PCB_123	0.85	1.35	1.28	0.70	1.04
PCB_138	4.86	6.40	5.46	3.52	5.06
PCB_141	1.09	1.34	1.61	0.71	1.19
PCB_149	7.98	10.91	13.50	6.67	9.76
PCB_151	2.92	2.86	4.22	2.38	3.09
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_156	< 0.02	< 0.02	0.26	0.40	0.33
PCB_157	< 0.02	< 0.02	0.10	< 0.02	0.10
PCB_158	0.53	0.75	0.67	0.43	0.60
PCB_167	< 0.02	0.33	< 0.02	< 0.02	0.33
PCB_170	0.39	0.46	0.32	0.31	0.37
PCB_174	0.50	0.85	0.67	0.39	0.60
PCB_180	0.71	1.27	0.94	0.77	0.92
PCB_183	0.44	0.63	0.49	0.33	0.47
PCB_187	1.06	1.37	1.08	0.71	1.06
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_194	0.26	0.16	< 0.02	< 0.02	0.21
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_203	0.12	0.18	0.12	0.12	0.14
PCB_153+132	7.96	11.00	11.50	6.76	9.31
PCB_31+28	38.06	51.29	52.70	18.40	40.11
PCB_41/64	18.02	25.99	12.01	16.09	18.03
PCB_60/56	2.98	3.80	4.54	2.46	3.44
PCB_90/101	19.40	25.83	41.84	18.89	26.49
ΡCΒ_Σ7ΡCΒ	91.17	127.39	150.51	72.34	110.35

HAZELRIGG 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
PCB_18	2.35	7.85	23.04	12.30	11.39
PCB_22	6.79	25.33	8.31	4.16	11.15
PCB_44	< 0.02	< 0.02	4.98	3.12	4.05
PCB_49	2.49	9.89	3.72	2.05	4.54
PCB_52	0.73	0.65	7.20	4.48	3.26
PCB_70	2.08	3.62	1.41	1.02	2.03
PCB_74	2.14	2.61	0.87	0.76	1.59
PCB_87	0.56	0.81	0.66	0.58	0.65
PCB_95	2.69	1.89	5.62	4.02	3.55
PCB_99	0.82	1.25	1.14	0.76	0.99
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_105	< 0.02	< 0.02	0.18	0.13	0.15
PCB_110	1.35	2.39	1.93	1.81	1.87
PCB_114	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_118	0.77	0.95	1.59	1.15	1.11
PCB_123	0.43	0.69	0.25	< 0.02	0.45
PCB_138	0.74	0.72	0.60	0.49	0.64
PCB_141	0.32	0.49	0.19	< 0.02	0.33
PCB_149	1.46	1.20	1.60	0.96	1.31
PCB_151	1.06	1.06	1.27	0.45	0.96
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_156	< 0.02	< 0.02	0.04	< 0.02	0.04
PCB_157	< 0.02	< 0.02	0.06	< 0.02	0.06
PCB_158	0.06	0.23	0.10	< 0.02	0.13
PCB_167	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_170	0.06	0.08	0.12	0.08	0.08
PCB_174	0.15	0.11	0.15	0.10	0.13
PCB_180	0.20	0.15	0.28	0.23	0.21
PCB_183	0.10	0.12	0.12	0.07	0.10
PCB_187	0.28	0.35	0.36	0.25	0.31
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_194	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_203	< 0.02	0.06	0.05	0.07	0.06
PCB_153+132	1.30	1.19	1.39	0.94	1.21
PCB_31+28	12.76	23.36	30.40	18.73	21.31
PCB_41/64	3.72	4.75	6.65	5.25	5.09
PCB_60/56	1.64	0.95	6.50	3.31	3.10
PCB_90/101	3.24	4.73	3.03	2.52	3.38
PCB_Σ7PCB	18.97	30.81	42.89	27.39	30.01

AUCHENCORTH 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
PCB_18	2.12	3.79	16.02	4.77	6.68
PCB_22	3.92	5.98	3.79	1.35	3.76
PCB_44	2.84	0.30	0.90	0.78	1.20
PCB_49	3.81	2.62	1.83	0.79	2.26
PCB_52	1.54	1.26	2.62	1.58	1.75
PCB_70	0.75	0.89	0.97	0.51	0.78
PCB_74	1.14	1.27	0.60	0.30	0.83
PCB_87	0.32	0.18	0.26	0.15	0.23
PCB_95	0.94	0.89	1.22	0.86	0.98
PCB_99	0.28	0.25	0.25	0.20	0.24
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_105	< 0.02	< 0.02	0.11	0.06	0.08
PCB_110	0.57	0.52	0.13	0.37	0.40
PCB_114	0.07	< 0.02	< 0.02	< 0.02	0.07
PCB_118	0.21	0.39	0.25	0.20	0.26
PCB_123	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_138	0.39	0.33	0.19	0.13	0.26
PCB_141	0.11	0.10	0.07	< 0.02	0.09
PCB_149	0.59	0.53	0.67	0.48	0.57
PCB_151	0.30	0.23	0.25	0.16	0.24
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_156	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_157	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_158	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_167	0.12	0.05	< 0.02	< 0.02	0.08
PCB_170	0.11	0.06	0.07	0.05	0.07
PCB_174	0.12	0.06	0.11	0.07	0.09
PCB_180	0.20	0.14	0.24	0.07	0.16
PCB_183	0.10	0.07	0.07	0.06	0.07
PCB_187	0.14	0.14	0.20	0.17	0.16
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_194	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_203	< 0.02	0.04	0.07	0.06	0.05
PCB_153+132	0.61	0.54	0.56	0.41	0.53
PCB_31+28	8.16	12.03	18.73	6.82	11.43
PCB_41/64	4.08	4.98	3.94	1.60	3.65
PCB_60/56	0.57	0.59	0.71	0.28	0.54
PCB_90/101	0.82	0.75	0.95	0.65	0.79
PCB_Σ7PCB	11.72	15.05	23.28	9.66	14.93

HIGH MUFFLES 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
PCB_18	5.56	5.44	14.88	9.04	8.73
PCB_22	5.49	8.81	4.43	2.13	5.21
PCB_44	< 0.02	< 0.02	2.11	1.56	1.84
PCB_49	9.18	20.06	1.74	1.07	8.01
PCB_52	3.29	2.14	3.27	2.79	2.87
PCB_70	1.25	1.07	1.23	0.79	1.08
PCB_74	1.03	1.65	0.98	0.62	1.07
PCB_87	0.57	< 0.02	0.27	0.26	0.37
PCB_95	5.91	1.39	1.47	1.14	2.47
PCB_99	0.44	< 0.02	0.38	0.31	0.38
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_105	0.38	< 0.02	0.10	< 0.02	0.24
PCB_110	1.41	0.82	0.72	0.67	0.90
PCB_114	0.10	< 0.02	< 0.02	< 0.02	0.10
PCB_118	0.39	0.26	0.32	0.28	0.31
PCB_123	0.31	< 0.02	< 0.02	< 0.02	0.31
PCB_138	1.02	0.52	0.36	0.24	0.54
PCB_141	0.38	0.14	0.08	< 0.02	0.20
PCB_149	3.08	0.90	0.80	0.74	1.38
PCB_151	1.99	0.45	0.43	0.34	0.80
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_156	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_157	< 0.02	< 0.02	0.05	< 0.02	0.05
PCB_158	0.07	< 0.02	< 0.02	< 0.02	0.07
PCB_167	< 0.02	0.13	< 0.02	< 0.02	0.13
PCB_170	0.06	0.07	0.07	0.09	0.07
PCB_174	0.19	0.12	0.14	0.08	0.13
PCB_180	0.25	0.19	0.26	0.23	0.23
PCB_183	0.15	0.10	0.10	0.08	0.11
PCB_187	0.29	0.26	0.26	0.20	0.25
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_194	< 0.02	< 0.02	0.04	0.00	0.02
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_203	< 0.02	< 0.02	0.08	0.10	0.09
PCB_153+132	2.24	0.96	0.86	0.61	1.17
PCB_31+28	17.66	15.67	25.56	11.16	17.51
PCB_41/64	3.80	6.78	5.30	5.23	5.28
PCB_60/56	0.78	0.74	0.92	0.56	0.75
PCB_90/101	4.55	0.94	1.23	1.08	1.95
PCB_Σ7PCB	29.00	20.43	31.56	16.11	24.28

WEYBOURNE 2011	Q1.11	Q2.11	Q3.11	Q4. 11	Average 2011
PCB_18	1.29	4.40	15.10	11.46	8.06
PCB_22	2.98	4.21	2.85	2.17	3.05
PCB_44	1.42	1.81	2.18	2.16	1.89
PCB_49	2.52	2.39	2.09	1.86	2.21
PCB_52	1.88	3.10	3.87	4.24	3.27
PCB_70	2.98	1.05	1.11	1.36	1.63
PCB_74	0.83	0.70	0.77	0.74	0.76
PCB_87	< 0.02	< 0.02	0.48	0.63	0.55
PCB_95	0.61	1.44	1.73	2.14	1.48
PCB_99	0.00	0.47	0.47	0.67	0.40
PCB_104	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_105	< 0.02	0.30	0.16	0.26	0.24
PCB_110	0.44	0.98	1.04	1.34	0.95
PCB_114	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_118	0.15	0.46	0.43	0.50	0.38
PCB_123	< 0.02	0.10	0.11	0.11	0.10
PCB_138	< 0.02	0.55	0.33	0.50	0.46
PCB_141	< 0.02	0.10	0.05	0.10	0.08
PCB_149	0.29	0.84	0.93	1.12	0.79
PCB_151	0.16	0.37	0.41	0.47	0.35
PCB_155	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_156	< 0.02	< 0.02	0.09	0.07	0.08
PCB_157	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_158	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_167	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_170	< 0.02	0.08	0.08	0.13	0.10
PCB_174	< 0.02	0.13	0.08	0.14	0.12
PCB_180	< 0.02	0.20	0.22	0.27	0.23
PCB_183	< 0.02	0.08	0.08	0.11	0.09
PCB_187	< 0.02	0.19	0.20	0.24	0.21
PCB_188	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_189	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_194	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_199	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
PCB_203	< 0.02	< 0.02	< 0.02	0.09	0.09
PCB_153+132	< 0.02	0.91	0.83	1.12	0.95
PCB_31+28	6.26	14.14	17.79	13.18	12.84
PCB_41/64	2.77	2.12	4.59	5.79	3.82
PCB_60/56	< 0.02	0.61	0.58	0.77	0.65
PCB_90/101	0.85	1.36	1.58	2.04	1.46
ΡCΒ_Σ7ΡCΒ	8.99	20.25	24.62	21.34	18.80

Appendix 3. PBDE data for 2011. (All data in pg.m<sup>-3</sup>)

PBDE LONDON 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	0.40	< 0.02	0.40
BDE_28	0.20	0.35	0.53	0.28	0.34
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	< 0.02	0.32	0.32
BDE_37	< 0.02	< 0.02	1.40	0.62	1.01
BDE_47	2.26	2.04	2.25	1.48	2.01
BDE_49	0.21	0.27	0.23	< 0.02	0.24
BDE_66	< 0.02	< 0.02	0.22	< 0.02	0.22
BDE_71	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	0.27	< 0.02	0.27
BDE_99	4.17	1.83	2.29	1.64	2.48
BDE_100	0.67	0.27	0.30	0.28	0.38
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	< 0.02	< 0.02	0.59	0.69	0.64
BDE_153	0.39	0.27	0.44	0.60	0.43
BDE_154	0.63	< 0.02	< 0.02	< 0.02	0.63
BDE_166	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_183	0.54	< 0.02	1.15	2.68	1.46
BDE_196	1.54	0.78	0.92	1.15	1.10
BDE_197	1.31	0.50	0.86	1.28	0.99
<b>BDE 47 + 99</b>	6.42	3.87	4.54	3.12	4.49

PBDE MANCHESTER 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	0.30	0.35	0.32
BDE_28	< 0.02	0.56	0.45	0.34	0.45
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_37	< 0.02	< 0.02	0.53	0.42	0.47
BDE_47	2.97	7.26	7.79	4.40	5.61
BDE_49	0.36	0.43	0.42	0.79	0.50
BDE_66	< 0.02	0.28	0.41	0.61	0.43
BDE_71	< 0.02	< 0.02	< 0.02	0.23	0.23
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	< 0.02	0.28	0.28
BDE_99	1.52	3.26	3.39	2.84	2.75
BDE_100	0.45	0.65	0.82	0.96	0.72
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	< 0.02	< 0.02	0.41	1.18	0.79
BDE_153	0.34	0.36	0.70	2.09	0.87
BDE_154	0.67	0.34	0.50	1.28	0.70
BDE_166	< 0.02	< 0.02	0.40	0.79	0.60
BDE_183	0.70	< 0.02	1.38	9.09	3.72
BDE_196	1.48	1.07	0.93	2.99	1.62
BDE_197	1.52	< 0.02	0.87	3.51	1.97
BDE 47 + 99	4.49	10.52	11.17	7.25	8.36

PBDE AUCHENCORTH 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_28	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_37	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_47	0.94	0.83	0.96	0.70	0.86
BDE_49	0.17	< 0.02	< 0.02	0.10	0.14
BDE_66	0.14	< 0.02	< 0.02	< 0.02	0.14
BDE_71	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_99	0.37	0.19	0.38	0.67	0.40
BDE_100	0.18	< 0.02	< 0.02	0.12	0.15
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	0.19	< 0.02	< 0.02	< 0.02	0.19
BDE_153	0.26	< 0.02	< 0.02	0.19	0.23
BDE_154	0.20	< 0.02	< 0.02	0.24	0.22
BDE_166	0.15	< 0.02	< 0.02	< 0.02	0.15
BDE_183	0.65	0.22	0.57	0.83	0.57
BDE_196	0.59	< 0.02	0.33	0.60	0.51
BDE_197	0.54	< 0.02	< 0.02	0.58	0.56
BDE 47 + 99	1.31	1.02	1.35	1.37	1.26

PBDE HIGH MUFFLES 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_28	< 0.02	< 0.02	0.25	< 0.02	0.25
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_37	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_47	0.76	0.91	1.04	0.81	0.88
BDE_49	0.12	0.17	0.19	0.22	0.18
BDE_66	< 0.02	< 0.02	0.11	0.14	0.12
BDE_71	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_99	0.40	0.25	0.38	0.69	0.43
BDE_100	< 0.02	< 0.02	0.10	0.15	0.12
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_153	0.22	< 0.02	0.14	0.35	0.24
BDE_154	0.30	< 0.02	< 0.02	0.20	0.25
BDE_166	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_183	< 0.02	< 0.02	< 0.02	1.53	1.53
BDE_196	0.70	< 0.02	0.45	0.86	0.67
BDE_197	0.53	< 0.02	< 0.02	0.93	0.73
BDE 47 + 99	1.16	1.16	1.43	1.50	1.31

PBDE HAZELRIGG 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_28	0.13	0.12	0.18	< 0.02	0.14
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	0.28	0.15	0.22
BDE_37	< 0.02	< 0.02	0.11	< 0.02	0.11
BDE_47	0.90	0.86	1.56	0.84	1.04
BDE_49	0.12	0.11	0.24	< 0.02	0.16
BDE_66	< 0.02	< 0.02	0.16	0.14	0.15
BDE_71	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_99	0.75	0.46	1.07	0.73	0.75
BDE_100	0.16	0.13	0.30	0.15	0.19
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	< 0.02	< 0.02	0.18	0.26	0.22
BDE_153	0.22	0.12	0.26	0.26	0.21
BDE_154	0.37	< 0.02	0.31	0.25	0.31
BDE_166	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_183	0.34	< 0.02	1.12	1.18	0.88
BDE_196	0.61	< 0.02	0.43	0.60	0.55
BDE_197	0.56	< 0.02	0.41	0.47	0.48
BDE 47 + 99	1.65	1.32	2.63	1.56	1.79

PBDE WEYBOURNE 2011	Q1.11	Q2.11	Q3.11	Q4.11	Average 2011
BDE_17	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_28	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_32	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_35	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_37	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_47	< 0.02	0.52	0.73	0.74	0.66
BDE_49	0.42	< 0.02	0.18	< 0.02	0.30
BDE_66	< 0.02	< 0.02	0.13	0.12	0.13
BDE_71	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_75	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_77	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_85	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_99	< 0.02	0.61	0.57	0.62	0.60
BDE_100	< 0.02	< 0.02	< 0.02	0.13	0.13
BDE_119	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_138	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_153	0.93	0.15	0.22	0.45	0.44
BDE_154	1.53	< 0.02	< 0.02	0.25	0.89
BDE_166	< 0.02	< 0.02	< 0.02	< 0.02	< 0.02
BDE_183	1.24	0.37	0.76	1.80	1.04
BDE_196	3.46	0.41	0.35	0.73	1.24
BDE_197	2.73	0.37	0.29	0.80	1.05
BDE 47 + 99	< 0.02	1.13	1.30	1.37	1.27