

2013 Annual Report for the UK Black Carbon Network

**D Butterfield
S Beccaceci
P Quincey
A Lilley
C Bradshaw
G Fuller
D Green
A Font Font**

OCTOBER 2014

2013 Annual Report for the UK Black Carbon Network

D Butterfield, S Beccaceci, P Quincey, A Lilley and C Bradshaw
Analytical Science Division, NPL

G Fuller, D Green and A Font Font
Environmental Research Group, King's College London

© Queen's Printer and Controller of HMSO, year 2014

ISSN: 1754-2928

National Physical Laboratory
Hampton Road, Teddington, Middlesex, TW11 0LW

Extracts from this report may be reproduced provided the source is acknowledged
and the extract is not taken out of context.

Approved on behalf of NPLML by Martyn Sene, Operations Director

EXECUTIVE SUMMARY

This report covers the operation of the UK Black Carbon Network and the data collected by the Network in 2013. The Aethalometer instrument used on the Network makes measurements of Black Carbon (BC) and UV component.

The National Physical Laboratory (NPL) in partnership with the Environmental Research Group at King's College London was awarded the contract to restructure and run the UK Black Smoke Network by the Department for Environment, Food and Rural Affairs (Defra) in September 2006. By 2007 the network was making measurements at 21 sites. In 2008 the black smoke samplers were replaced by model AE22 Aethalometers. The Network was reduced to 20 sites in October 2009, when the Bradford site was closed after Defra reviewed its monitoring requirements. In early 2012 the Network underwent major restructuring to focus the network on targeted monitoring of specific emission sources. In late 2012 Aethalometers were installed at Goonhilly and Ballymena to investigate emissions from shipping and domestic fuel use, respectively.

The 2013 data capture for Aethalometer measurements was 97%. This is a remarkably high figure for an automatic monitoring network and shows the reliability of the Aethalometers and the effectiveness of daily on-line surveillance of the equipment.

2013 urban annual mean Black Carbon concentrations on the Network (with the corresponding 2012 concentrations in brackets) ranged from 1.0 (0.7) $\mu\text{g.m}^{-3}$ at Dunmurry Kilmakee to 6.9 (8.9) $\mu\text{g.m}^{-3}$ at Marylebone Road. Harwell (rural background) reported an average concentration of 0.5 (0.5) $\mu\text{g.m}^{-3}$, with similar values at the newly installed rural sites at Auchencorth Moss and Detling. The network mean for Black Carbon concentration was 1.5 (1.9) $\mu\text{g.m}^{-3}$.

The annual mean UV component concentrations ranged from 0.1 (0.0) $\mu\text{g.m}^{-3}$ at Auchencorth Moss to 1.2 (1.1) $\mu\text{g.m}^{-3}$ at Strabane. Roadside sites show many negative spikes in the UV component concentration, thought to be measurement artefacts caused by volatile components in fresh vehicle exhaust plumes. This effect is most prevalent at Marylebone Road. The network mean for UV component concentration was 0.3 (0.3) $\mu\text{g.m}^{-3}$. The figures in brackets are again the corresponding concentrations for 2012.

The new network design implemented in early 2012 allows urban increments in Black Carbon and UV component concentrations to be determined for London, Birmingham and Glasgow, and roadside increments for London and Birmingham. The urban increment for Black Carbon was similar for all locations while the roadside increment was roughly proportional to road traffic volumes, especially buses and taxis. There was no significant urban or roadside increment in UV component concentration.

Daily averages of the measurements show that the highest concentrations of Black Carbon are found on weekdays, with the weekends generally having lower values. The hourly averages of Black Carbon broadly show a commuter traffic-based signature, with the exception of Strabane, Ballymena and Dunmurry, in Northern Ireland, which show elevated levels from 15:00hrs to 23:00hrs, which is indicative of local residential heating.

From the diurnal plots it can be seen that the main driver behind the UV component concentrations is domestic fuel use, with elevated concentrations in the evenings at the urban background sites. Sites in Northern Ireland show the largest evening effect due to the higher use of oil and solid fuel (coal and wood) for domestic heating. From the daily averages it can be seen that the UV component concentration is fairly consistent on weekdays with greater concentrations at the weekend.

Comparisons between Black Carbon concentrations and Elemental Carbon concentrations showed good linear relationships between the measurements at North Kensington and Harwell, with R^2 values

of 0.86 and 0.97. The scatter at Marylebone Road was larger than in previous years ($R^2 = 0.54$). For North Kensington and Harwell the slopes were 1.48 and 1.56, with intercepts were 0.41 and 0.17 $\mu\text{g}\cdot\text{m}^{-3}$. The slope and intercept for Marylebone road were 1.34 and 1.58 $\mu\text{g}\cdot\text{m}^{-3}$, albeit with the low R^2 value.

Comparisons between particulate mass concentrations and Black Carbon concentrations showed that Black Carbon makes up a significant proportion of the particulate mass concentration at roadside sites. At Marylebone Road the Black Carbon concentration comprises 24% of the PM_{10} concentration and 34% of the $\text{PM}_{2.5}$ concentration, while at Birmingham Tyburn roadside Black Carbon forms 16% and 19% of PM_{10} and $\text{PM}_{2.5}$ respectively.

Monthly means of Black Carbon concentrations were examined over the period 2009 to 2013 to evaluate trends. The only site with a significant trend in Black Carbon concentrations was Marylebone Road. Here, black carbon concentrations have dropped over the last 2 years. Over the same period there were no sites with a significant slope in UV component concentration.

Using a published method for converting Black Smoke Index into Black Carbon concentration it can be shown that there is no significant discontinuity in results from the two methods and, apart from Strabane, there is no obvious long-term trend in the Black Carbon concentrations over the past 14 years.

CEN are currently formulating a standard for the measurement of elemental carbon and organic carbon deposited on filters. As part of this work there are field validation trials where automatic instruments, including an Aethalometer, are being evaluated in parallel with filter samplers.

CONTENTS

1.0 INTRODUCTION.....	1
1.1 GENERAL.....	1
1.2 BLACK CARBON	1
1.3 MEASUREMENT METHOD	2
1.3.1 Aethalometer instrument and data processing	2
1.3.2 Sampling	2
2.0 NETWORK INFRASTRUCTURE	3
2.1 NETWORK SITES AND DESIGN.....	3
2.2 CHANGES TO THE NETWORK.....	5
2.3 NETWORK OPERATION.....	5
3.0 QUALITY ASSURANCE AND QUALITY CONTROL (QA/QC).....	8
3.1 SITE AUDITS.....	8
3.1.1 Sampler Leak Rate and Calibration of Sample Flow.....	8
3.1.2 Instrument Performance.....	9
4.0 MEASUREMENT UNCERTAINTY	12
4.1 SAMPLE VOLUME.....	12
4.2 MEASUREMENT OF ABSORPTION	12
4.3 CORRECTION FOR SPOT DARKENING.....	12
4.4 PRELIMINARY OVERALL MEASUREMENT UNCERTAINTY	13
5.0 RESULTS	14
5.1 TIME SERIES	14
5.1.1 Black Carbon	14
5.1.2 UV component.....	17
5.2 AVERAGES AND DATA CAPTURE	21
5.2.1 Black Carbon	21
5.2.2 UV component.....	24
5.2.3 Data Capture	26
5.3 TEMPORAL VARIATIONS.....	27
5.4 COMPARISONS WITH OTHER POLLUTANTS.....	54
5.4.1 Elemental Carbon.....	55
5.4.2 Polycyclic Aromatic Hydrocarbons (PAH)	57
5.4.3 Particulate Mass	60
5.5 TRENDS.....	61
5.5.1 Short-Term Trends by Site.....	61
5.5.2 Network Short-Term Trends.....	72
6.0 EUROPEAN STANDARDISATION	75
7.0 CONCLUSIONS	76
APPENDIX 1 MEASUREMENTS BY THE CARBOTRAF PROJECT IN GLASGOW	77

1.0 INTRODUCTION

1.1 GENERAL

The National Physical Laboratory (NPL) in partnership with the Environmental Research Group at King's College London was awarded the contract to restructure and run the UK Black Smoke Network by the Department for Environment, Food and Rural Affairs (Defra) in September 2006. During 2007 the number of sites in the network expanded from 14 sites to 21 sites, as samplers were installed into mainly Automatic Urban and Rural Network (AURN) sites. By March 2007 all of the 21 sites were operational.

In 2008 the black smoke samplers were replaced by model AE22 Aethalometers. The Network was reduced to 20 sites in October 2009 when the Bradford site was closed after Defra reviewed its monitoring requirements.

In January 2012 a new contract was issued by Defra to run the Network, and this involved a major reorganisation of the monitoring sites. The focus of the Network changed to provide targeted monitoring of the major urban conurbations of London, Birmingham and Glasgow, plus additional sites to monitor specific emission sources. This involved the closing of 11 existing sites and the opening of 4 new sites, leaving a Network of 13 sites, plus Harwell which is run as part of Defra's Particle Numbers and Speciation Network. By targeting these urban areas a more detailed apportionment of emissions of Black Carbon emission sources should be provided.

Later in 2012 Defra took up two options to increase monitoring of specific emission sources which resulting in a site being installed at Goonhilly in Cornwall to assess Black Carbon from shipping, and at Ballymena in Northern Ireland to extend the work on monitoring emissions from solid fuel and biomass burning. The Goonhilly monitoring continued until November 2013 and the Ballymena monitoring until March 2014. Aethalometer measurements stopped at the Norwich Lakenfields site in May 2013 as the site was no longer provided additional information on black carbon emissions above that already collected from the other sites on the Network.

1.2 BLACK CARBON

Black Carbon (BC) is a measure of airborne soot-like carbon (in $\mu\text{g}\cdot\text{m}^{-3}$) based on the optical absorption of specific wavelengths by particulates collected on a filter. Ideally it is a similar metric to Elemental Carbon (EC), a measure of soot-like carbon determined by thermo-optical (chemical) techniques, though in practice the EC fraction of total carbon depends strongly on the method chosen. BC has a close relationship to the Black Smoke measure monitored by the network and its predecessors for many decades before the installation of the Aethalometers¹, though again this can be affected by the instruments and circumstances.

BC is typically formed through the incomplete combustion of fossil fuels, biofuel, and biomass, and is emitted in both anthropogenic and naturally occurring soot. It consists of pure carbon in several forms. Black carbon warms the planet by absorbing heat in the atmosphere and by reducing albedo (the ability to reflect sunlight) when deposited on snow and ice. Black Carbon stays in the atmosphere for periods of days to weeks, whereas CO₂ has an atmospheric lifetime of more than 100 years.

The terminology to be used for 'Black Carbon' data is currently receiving attention within, for example, the Global Atmosphere Watch special aerosol advisory group. This is mainly concerned with highlighting the assumptions used to convert optical data to mass concentration data. The procedures used for the Black Carbon Network are described within this report.

¹ P Quincey, A relationship between Black Smoke Index and Black Carbon concentration, Atmospheric Environment 41 (2007) 7964–7968

1.3 MEASUREMENT METHOD

1.3.1 Aethalometer instrument and data processing

Aethalometers quantify Black Carbon on filter samples based on the transmission of light through a sample. The sample is collected onto a quartz tape, and the change in absorption coefficient of the sample is measured by a single pass transmission of light through the sample, measured relative to a clean piece of filter. The system evaluates changes in two optical sensors (sample and reference), with the light source both on and off, such that independent measurements of the change in attenuation of the sample are produced for averaging periods of typically five minutes. The absorption coefficient for material added during the period, σ [m^{-1}], is calculated from the attenuation change, the filter area, and volume of the sample. This is converted to a Black Carbon concentration for the period, as a first approximation, using a mass extinction coefficient [$16.6 \text{ m}^2 \text{ g}^{-1}$] chosen by the manufacturer to give a good match to Elemental Carbon. In practice this mass extinction coefficient will vary with factors such as particle size, sample composition and quantity of material already on the filter, as discussed below.

The Aethalometers run on the Network operate at 2 wavelengths, 880nm and 370 nm. The 880nm wavelength is used to measure the Black Carbon (BC) concentration of the aerosol, while the 370nm wavelength gives a measure of the “UV component” of the aerosol. At wavelengths shorter than about 400 nm, certain classes of organic compounds (such as polycyclic aromatic hydrocarbons, and also certain compounds present in tobacco smoke and smoke from wood burning) start to show strong UV absorbance. The UV component can therefore in principle be used as an indicator of wood and solid fuel emissions.

The UV component concentration presented in this report is obtained by subtracting the measured BC concentration from the concentration measured by the 370nm source. The UV component is not a real physical or chemical material, but a parameter based on UV absorption due to the mix of organic compounds measured at this wavelength. This metric termed ‘UVPM’ is expressed in units of ‘BC Equivalent’.

It is well known that the assumption of constant mass extinction coefficient does not hold as the filter spot darkens, leading to nonlinearity in the Aethalometer response. The effect of this nonlinearity means that the Aethalometer has reduced sensitivity to black carbon at high filter tape loadings. To correct for this nonlinearity, the model developed by A Virkkula² has been used to correct for increased attenuation due to spot darkening during sampling. This uses the simple equation:

$$BC_{corrected} = (1 + k \cdot ATN) \cdot BC_{uncorrected}$$

where ATN is the light attenuation by the filter spot, and k is a parameter determined for each filter spot such that continuity between adjacent filter spots is greatly improved. All of the Black Carbon and UV component results in this report have been corrected by this method.

1.3.2 Sampling

At all sites, ambient air is drawn into the sampling system through a standard stainless steel rain cap mounted on the end of a vertical stainless steel tube. Size selection of the sampled aerosol is made by a PM_{2.5} cyclone placed close to the inlet of the Aethalometer. All of the tubing before the cyclone is constructed from stainless steel. Sampling has been standardised across the network by using this size selective inlet before the Aethalometer, which was not possible with the Black Smoke method.

² A Virkkula et al, A Simple Procedure for Correcting Loading Effects of Aethalometer Data, Journal of Air and Waste Management Association, 57:1214-1222, 2007

2.0 NETWORK INFRASTRUCTURE

The following sections present the design of the Network, describe its operation and the changes to the Network in 2013.

2.1 NETWORK SITES AND DESIGN

Figure 1 shows the locations of the Aethalometers during 2013. The sites are located to target the measurement of traffic emissions of Black Carbon in urban areas, solid fuel and biomass emissions in Northern Ireland, Cardiff and Norwich and shipping emissions from the English Channel.

Urban and traffic increments were targeted by having a rural background, an urban background and a roadside / kerbside siting combination across each conurbation. These site combinations are shown in Table 1.

Conurbation	Site Name	Site Classification
Glasgow	Auchencorth Moss	Rural Background
	Glasgow Centre	Urban Background
	Glasgow Kerbside	Traffic
Birmingham	Harwell	Rural Background
	Birmingham Tyburn Background	Urban Background
	Birmingham Tyburn Roadside	Traffic
London	Harwell	Rural Background
	North Kensington	Urban Background
	Marylebone Road	Traffic
	Detling	Rural Background

Table 1 Sites to measure emissions of Black Carbon from traffic sources

The installation of a site at Glasgow Kerbside has been seriously delayed due to planning restrictions. A suitable site is still to be found.

Seven other sites make up the Network. These sites measure specific non-traffic emission sources and are listed in Table 2.

Site Name	Site classification	Emission Source
Belfast Centre	Urban Background	Solid fuel use
Lisburn Dunmurry	Urban Background	Solid fuel use
Strabane	Urban Background	Solid fuel use
Ballymena	Urban Background	Solid fuel use
Norwich Lakenfields	Urban Background	Possible solid fuel use
Cardiff 12	Urban Background	Domestic
Goonhilly	Rural Background	Shipping

Table 2 Sites to measure non-traffic related emission sources

The sites making up the Network are shown on a map in Figure 1.

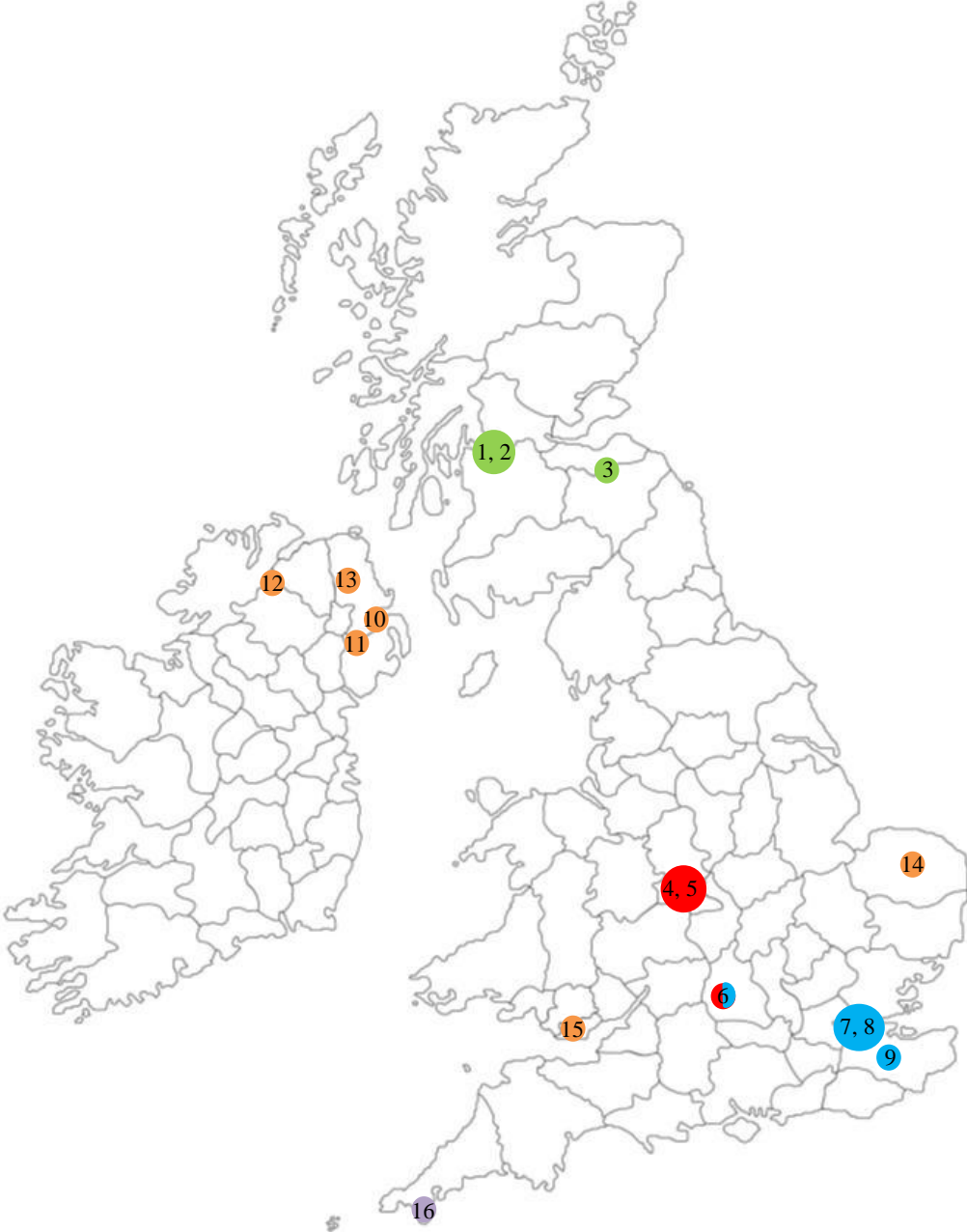


Figure 1 Sites on the BC Network after 2012 reorganisation
Key on next page

Key:

Emission source	Key	Site Name
Glasgow Urban Area	1	Glasgow Kerbside
	2	Glasgow Centre
	3	Auchencorth Moss
Birmingham Urban Area	4	Birmingham Tyburn Roadside
	5	Birmingham Tyburn Background
Birmingham Urban Area + London Urban Area	6	Harwell
London Urban Area	7	North Kensington
	8	Marylebone Road
	9	Detling
Solid Fuel Use	10	Belfast Centre
	11	Lisburn Dunmurry
	12	Strabane
	13	Ballymena
Possible Solid Fuel Use	14	Norwich Lakenfields
Domestic Emissions	15	Cardiff 12
Shipping Emissions	16	Goonhilly

2.2 CHANGES TO THE NETWORK

Since the installation of the Norwich site, unexpectedly high UV component concentrations have been measured for an urban area with non-solid fuel heating. These concentrations have been slowly decreasing over time, so it was decided that this site was no longer giving any additional information on domestic emissions. Monitoring at the site ceased on 14th May 2013.

In 2012 the Glasgow Centre site was closed due to redevelopment of the surrounding area. An urban background replacement site was found in 2013 and the Glasgow Townhead site commissioned in the last quarter of 2013. The Aethalometer did not survive its storage period very well and had numerous performance issues during the commissioning, resulting in valid measurements only starting in December 2013.

During 2012/13 there was a 1 year project to try to measure the impact of shipping emissions on black carbon concentrations at a coastal rural background site downwind of shipping lanes. An Aethalometer was installed on Goonhilly Downs in Cornwall from 27th November 2012 to 21st November 2013.

To give more information on the emissions of Black Carbon and UV component sources from non-smokeless fuels in Northern Ireland, an Aethalometer was installed into the Ballymena AURN site in early November 2012. Ballymena is also a part of the UK Polycyclic Aromatic Hydrocarbons (PAH) Network. Measured PAH concentrations here are higher than expected from predicted fuel usage. Aethalometer measurements at Ballymena will continue into 2014.

2.3 NETWORK OPERATION

The operation of the Network was set up to mirror that of the AURN, to include a Central Management and Control Unit (CMCU) and a Quality Assurance and Quality Control Unit (QA/QC). The Environmental Research Group at King's College London (King's) carries out the CMCU activities. These activities include the routine collection of data from site, initial data validation and instrument fault finding, routine liaison with the Local Site Operators (LSO) and the Equipment Support Unit (ESU). The QA/QC activities are performed by NPL and include: site audits, inter-laboratory performance schemes, data ratification and reporting.

As the Aethalometer produces real-time continuous data, it was decided to perform remote data collection and diagnostics at each site via a modem to maximise data capture and minimise LSO costs. A summary of this activity is outlined below:

Measurements are collected from the 16 sites on the Network on a daily basis. Measurements of Black Carbon, UV carbon, flow, raw attenuation signals and tape life since the last data collection are requested from the Aethalometer and automatically loaded into King's database. The 5 minute mean measurements are averaged to 15 minute means in line with measurements made using gaseous and particulate monitors on the AURN. A valid 15 minute measurement is only calculated where two or more valid 5 minute measurements exist in that 15 minute period. A range of sensibility checks are undertaken at this point to ensure measurements are above zero and below a maximum limit ($100 \mu\text{g m}^{-3}$); the flow data is also checked to ensure it is 4 l/min ($\pm 10\%$).

The data from each site is assessed using a range of algorithms/criteria, which determine whether the site requires a manual check; this is 'risk-based' data checking and provides a method for improving the efficiency of the manual checking procedure. The list of algorithms/criteria examine whether:

- Data warning flags have been attached to the data, either from the instrument or from the sensibility checks during processing.
- Data checking resulted in any notes or actions on the previous day.
- There are any services, local site operator visits or audits being undertaken the previous day.
- The data is stable for more than 6 consecutive 15 minute periods.
- The data capture over the previous 24 hours is less than 90 %.
- The site was not manually checked the previous day.

If any of these tests produce a positive result, the site is included in a list of sites to be examined manually. Where necessary, this manual validation is undertaken using the MONNET software package every working day; a screen shot of the 5 day data checking graph is shown in Figure 2. This shows the Black Carbon and UV carbon measurements and the flow measured by the instrument. Where NO_x measurements are available from the site (such as North Kensington and Marylebone Road) these are included as a method of assessing the impact of local traffic emissions. Further manual checks are made comparing the measurements between sites across the network to identify any outliers.

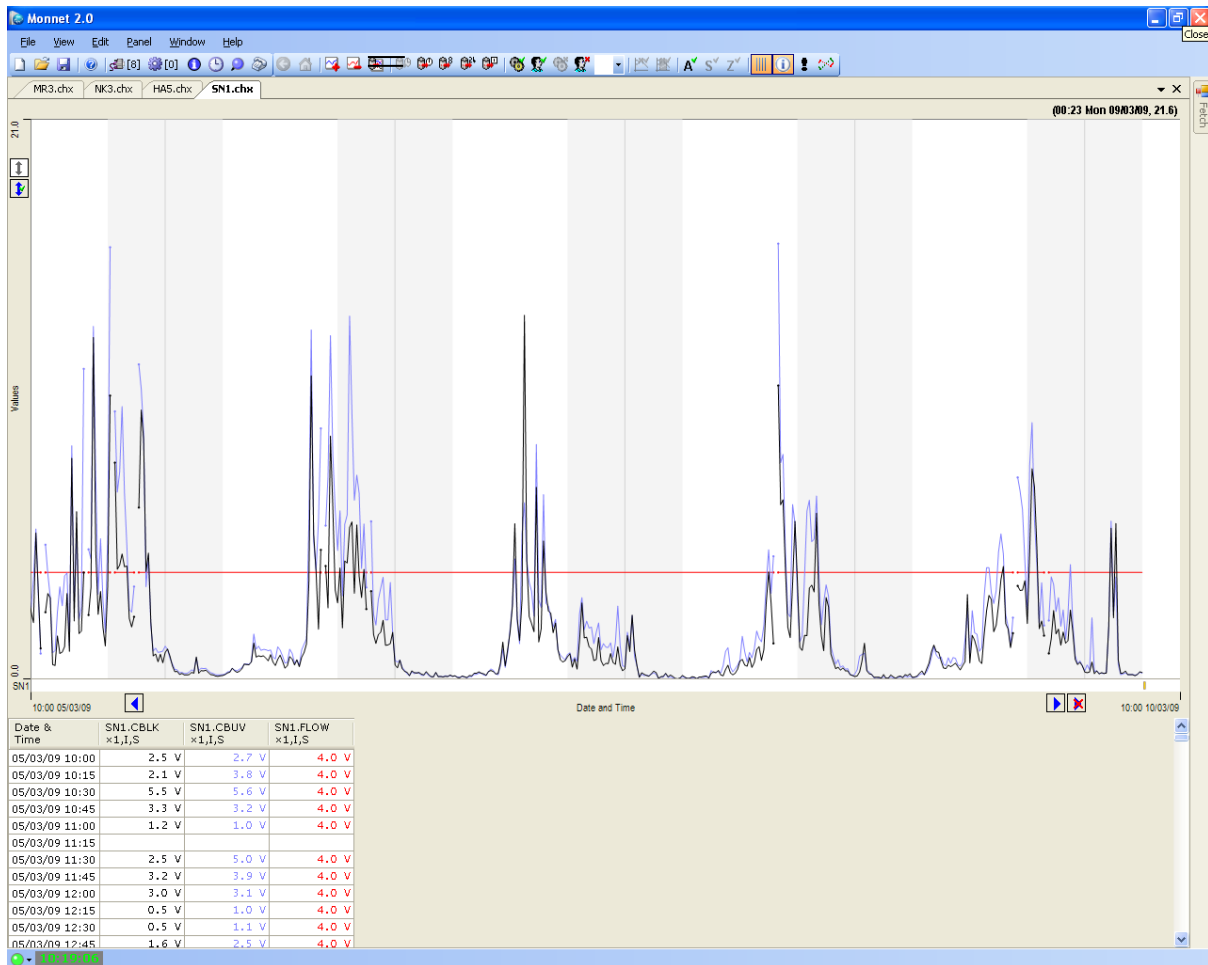


Figure 2 MONNET data checking graph

Issues raised during the manual data checking are noted in the database, this information is retained and passed to NPL to inform the ratification process. Occasionally, issues raised during data checking require an intervention from either the LSO or the ESU. If this is the case a visit request is sent to either the LSO or ESU. The reports generated from these visits are processed at King's and stored according to the site that they pertain to. The directory is mirrored to the web server and accessible via a password protected web portal for access during ratification.

3.0 QUALITY ASSURANCE AND QUALITY CONTROL (QA/QC)

Quality Assurance and Quality Control activities cover two main areas: site audits and instrument performance. The first addresses sampling issues and the second ensures the consistency and accuracy of the measurement of Black Carbon concentration.

3.1 SITE AUDITS

Table 3 gives the site audit dates and serial numbers of the Aethalometer audited.

Site	Date	Serial No.
Goonhilly	07/03/2013	852
Cardiff	08/07/2013	686
Birmingham Background	09/07/2013	859
Birmingham Roadside	09/07/2013	869
Marylebone Road	16/07/2013	866
Detling	17/07/2013	853
Belfast	20/08/2013	863
Dunmurry Kilmakee	20/08/2013	861
Ballymena	21/08/2013	849
Strabane	22/08/2013	848
Auchencorth Moss	25/09/2013	862
North Kensington	17/10/2013	850
Glasgow Townhead	31/10/2013	856
Harwell	12/11/2013	851

Table 3 Site Audit Visits

3.1.1 Sampler Leak Rate and Calibration of Sample Flow

The leak rate for Aethalometers is measured by simultaneously measuring the flow rate at the input and exhaust of the analyser and requires the use of two calibrated flow meters.

The absolute value of the inlet flow measured during the leak test is used to calibrate the sample flow of the instrument.

Both flow meters used were calibrated against National Standards. When taking into account the repeatability of the measurements in the field, the flow inlet and exhaust flows were measured with an uncertainty of $\pm 2.5\%$, expressed with a level of confidence of 95%.

According to the manufacturer, the maximum acceptable leak rate is 20%. Black Carbon concentrations are not corrected for leak rate, but the leak rate is included in the uncertainty budget.

Table 4 gives the measured leak rates and sample flows for each site:

Site	Leak Rate, %	Indicated Flow, lpm	Inlet Flow, lpm
Goonhilly	7.8	4.0	4.067
Cardiff	6.4	4.0	4.333
Birmingham Background	6.7	4.0	4.110
Birmingham Roadside	6.3	4.0	4.343
Marylebone Road	5.3	4.0	3.680
Detling	4.3	4.0	4.340
Belfast	14.3	4.0	4.775
Dunmurry Kilmakee	11.1	4.0	3.980
Ballymena	4.8	4.0	3.910
Strabane	6.7	4.0	4.130
Auchencorth Moss	8.7	4.0	4.200
North Kensington	6.5	4.0	4.240
Glasgow Townhead	N/A	N/A	N/A
Harwell	11.1	3.9	3.967

Table 4 Aethalometer leak rates and sample flows

The Glasgow Aethalometer had serious performance issues during the audit and was repaired by the equipment support after the audit visit.

3.1.2 Instrument Performance

The best simple indication of instrument performance can be gained by examining the zero noise of the Aethalometer, as this gives an indication of the optical and electrical stability over the measurement period. This is carried out by generating nominally particle free air using a High Efficiency Particle (HEPA) filter and examining the reported concentrations over an extended period of time. The Aethalometer concentration falls quickly to a stable value around zero, with variations due to noise in the optical system and electronics. Figure 3 shows a typical Aethalometer response to this test running on the normal time base of 5 minutes. The data are from the Strabane site.

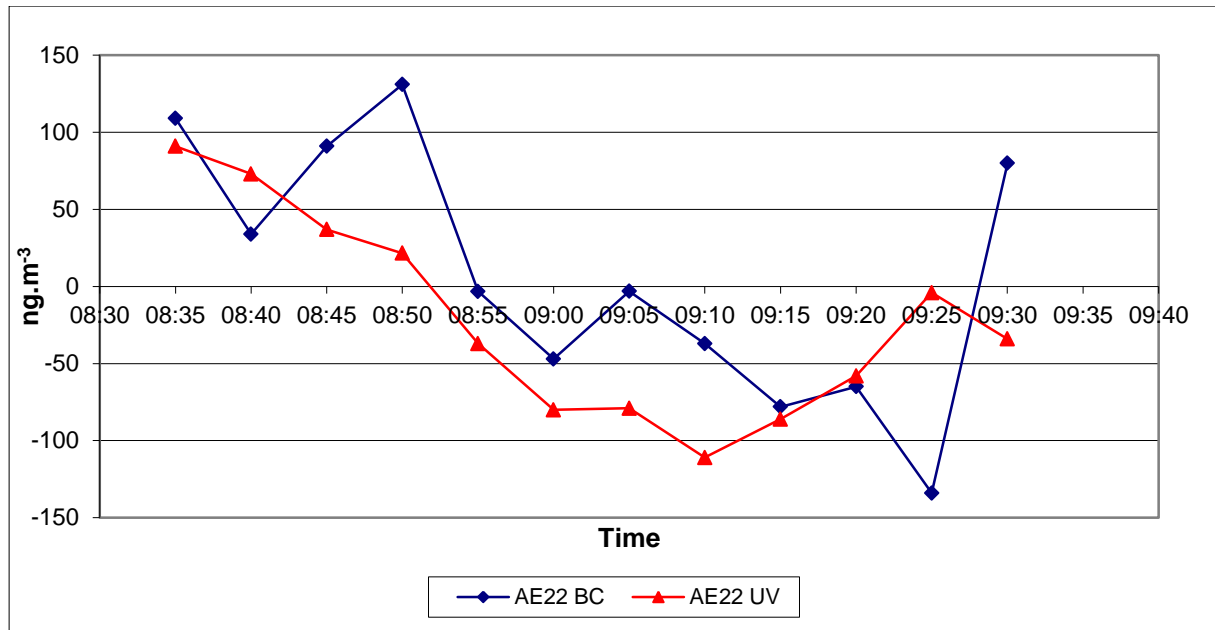


Figure 3 Strabane Aethalometer sampling HEPA filtered air

It should be noted that the concentrations normally reported by the Network are in $\mu\text{g.m}^{-3}$ and the above concentrations are in ng.m^{-3} . These concentrations during the zero noise tests are normally at or below the stated detection limit of the instrument, $0.1\mu\text{g.m}^{-3}$ (100ng.m^{-3}).

The zero noise is calculated as the standard deviation of the recorded concentrations multiplied by the student t-factor for the number of measurements. In this case the zero noise has been calculated for a 15 minute averaging period in line with the shortest averaging period used for recording data. The results for each site are given in Table 5.

Site	BC, ng.m^{-3}	UV Channel, ng.m^{-3}
Goonhilly	116	63
Cardiff	140	82
Birmingham Background	747	428
Birmingham Roadside	436	556
Marylebone Road	652	477
Detling	251	336
Strabane	183	145
Belfast	192	152
Ballymena	651	266
Dunmurry Kilmakee	176	100
Auchencorth Moss	145	106
North Kensington	202	130
Glasgow Townhead	144	77
Harwell	72	88

Table 5 Zero Noise of BC and UV component channels

It should be noted that the UV Channel in Table 5 is not the UV component concentration, but the result taken from the UV channel. Section 1.3.1 gives a description of how the UV component is calculated. The results in Table 5 are for measurements made with an Aethalometer time base of 5 minutes, these 5 minute readings are then averaged over 15 minutes before validation and ratification takes place. If the above 5 minute results are converted to a 15 minute noise result then the average Aethalometer noise recorded for the BC and UV channels is: 169 ng.m^{-3} and 88 ng.m^{-3} respectively. The shortest time period over which data is published is 1 hour, therefore the average noise on a 1 hour average concentration is expected to be 85 ng.m^{-3} and 62 ng.m^{-3} , for BC and UV respectively. These figures are considered acceptable, being far less than the network mean black carbon concentration of $1.52 \text{ } \mu\text{g.m}^{-3}$ and the network mean UV component of $0.3 \text{ } \mu\text{g.m}^{-3}$.

QA/QC methods are continually being reviewed and improved for the network.

4.0 MEASUREMENT UNCERTAINTY

4.1 SAMPLE VOLUME

From measurements at the site audit the sample volume can be determined with an uncertainty of $\pm 10.7\%$, expressed with a level of confidence of 95%. Included in this uncertainty are contributions from flow rate accuracy, repeatability, drift and leaks.

The leak rate is not used to correct the results, but is included as an uncertainty if the sampler passes the leak test at audit. The manufacturer's tolerance for leak rate is 20%. In the case of this uncertainty calculation the average value of leak rate determined in the 2013 audits was used. As leak rate is considered to be a rectangular distribution, its contribution to the standard uncertainty in sample volume is 4.3%.

4.2 MEASUREMENT OF ABSORPTION

The Aethalometer measurement does not depend on any absolute calibration of the detectors' response signals, but instead relies upon their ability to determine very small relative changes in optical transmission. Determining the zero noise of the system gives relevant information on the instrument's ability to measure small changes in optical transmission. Results from the HEPA filter zero noise tests show that the stability of the optical / electrical system is approximately $\pm 0.09 \mu\text{g.m}^{-3}$ for hourly means. Converting this into a standard uncertainty this represents a contribution of 5.6%, compared to the network mean of $1.52 \mu\text{g.m}^{-3}$.

4.3 CORRECTION FOR SPOT DARKENING

The Virkkula² model was used to correct the measured concentrations to account for the nonlinearity of Aethalometer black carbon and UV component measurements with attenuation. This effect and its correction introduce an uncertainty into the measurements. At most sites the correction can be seen to work well on the 15-minute data, in that there is minimal discontinuity when the spot location changes, and the associated uncertainty is considered to be small compared to other components. At sites where the concentration is changing quickly, such as Marylebone Road, this uncertainty in the 15-minute data becomes significant although this is decreased when hourly mean concentrations are calculated.

The differences between using individual spot corrections versus seasonal or monthly corrections have been examined, with the conclusion that where possible individual spot correction is the most suitable method.

The uncertainty due to the spot darkening cannot be directly determined and has not been included in the overall measurement uncertainty.

4.4 PRELIMINARY OVERALL MEASUREMENT UNCERTAINTY

When the contributions from sample volume and optical /electrical stability are combined, the overall measurement uncertainty for hourly Black Carbon concentrations is 15.4%, expressed with a level of confidence of 95%. The only source of uncertainty in the overall measurement uncertainty that reduces when producing longer term averages from the hourly data is the zero noise. The following is the overall measurement uncertainties for different averaging periods, expressed with a level of confidence of 95%:

Hourly	15.4%
Monthly	10.7%
Yearly	10.7%

Uncertainties year-on-year are consistent and compare well with the EU requirement of 25% for particulate mass concentration measurements.

This is an indicative measurement uncertainty for the Aethalometer method and is calculated from the results of the 2013 audit data. The site specific overall measurement uncertainty may differ from this value, and any effect from the spot darkening correction will be additional.

5.0 RESULTS

The concentration data for 2013 are presented in the following sections as time series graphs, summary graphs and tables of the annual mean concentration and data capture.

All of the Black Carbon and UV component data have been corrected for spot darkening using the Virkkula method².

The hourly data set for Black Carbon and UV component concentrations can be downloaded from Defra's UK-AIR: Air Information Resource Web Pages found at:

<http://uk-air.defra.gov.uk/>

5.1 TIME SERIES

The following sections present time series graphs of the Black Carbon and UV component concentrations.

5.1.1 Black Carbon

The following charts show the Black Carbon concentrations measured by the UK Black Carbon Network for 2013. The time resolution of the measurements is hourly. Data has been split into regions of the UK for presentation purposes. The maximum y-axis on these charts has been set to 45 $\mu\text{g.m}^{-3}$ to enable easy comparison between charts, except for Figure 8 which shows concentrations measured at rural locations.

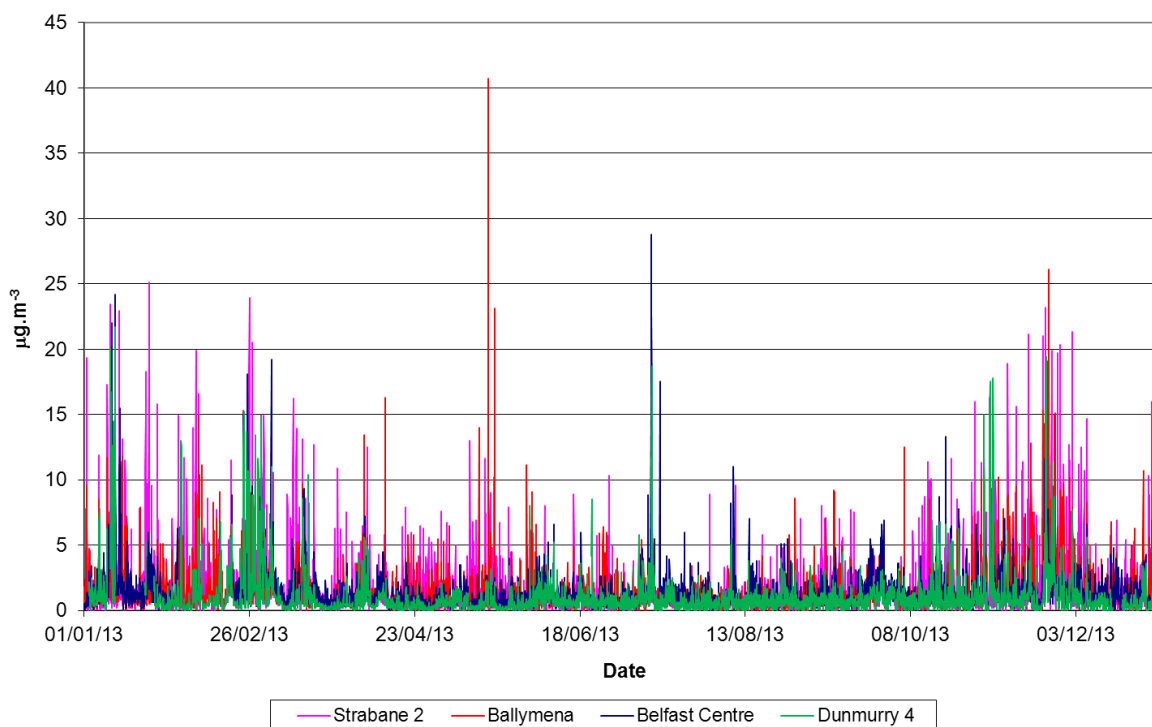


Figure 4 Black Carbon concentrations during 2013 in Northern Ireland

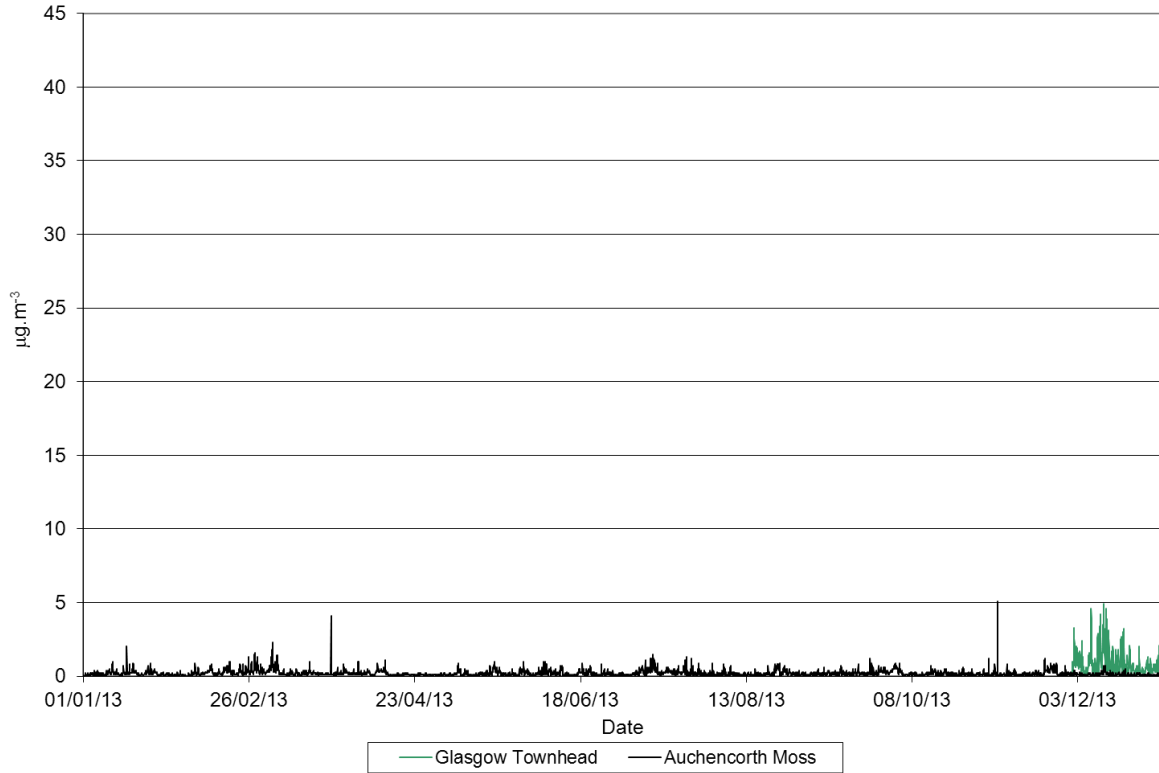


Figure 5 Black Carbon concentrations during 2013 in Scotland

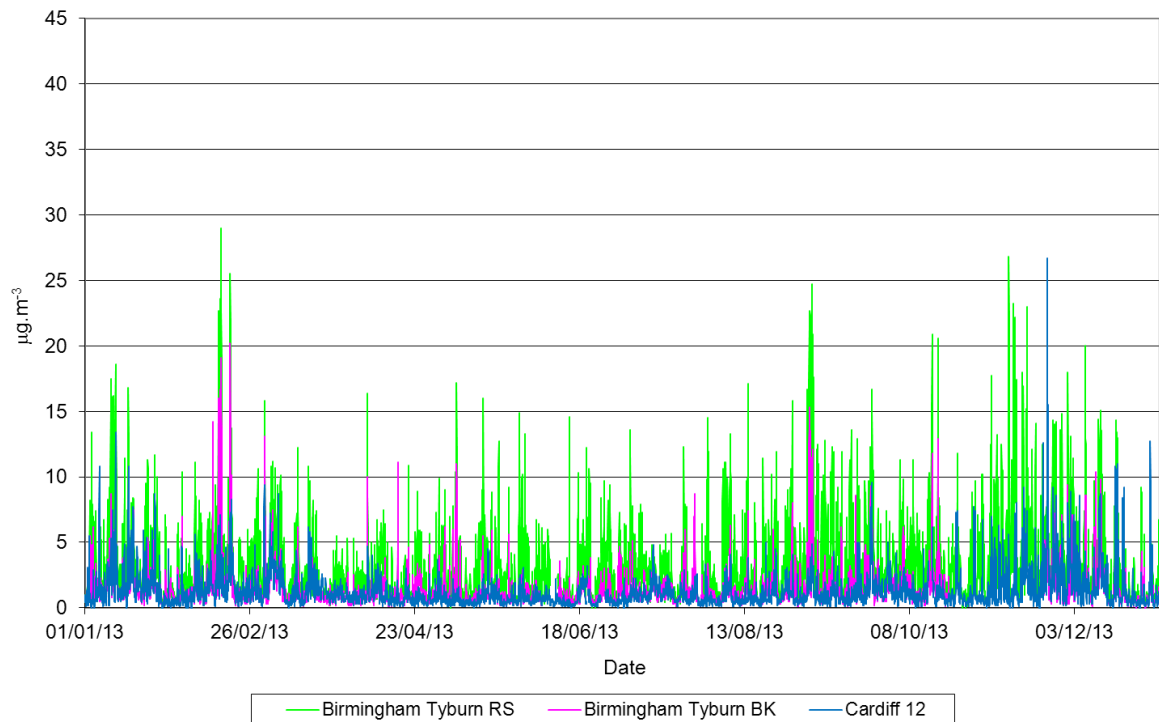


Figure 6 Black Carbon concentrations during 2013 in Wales and the Midlands

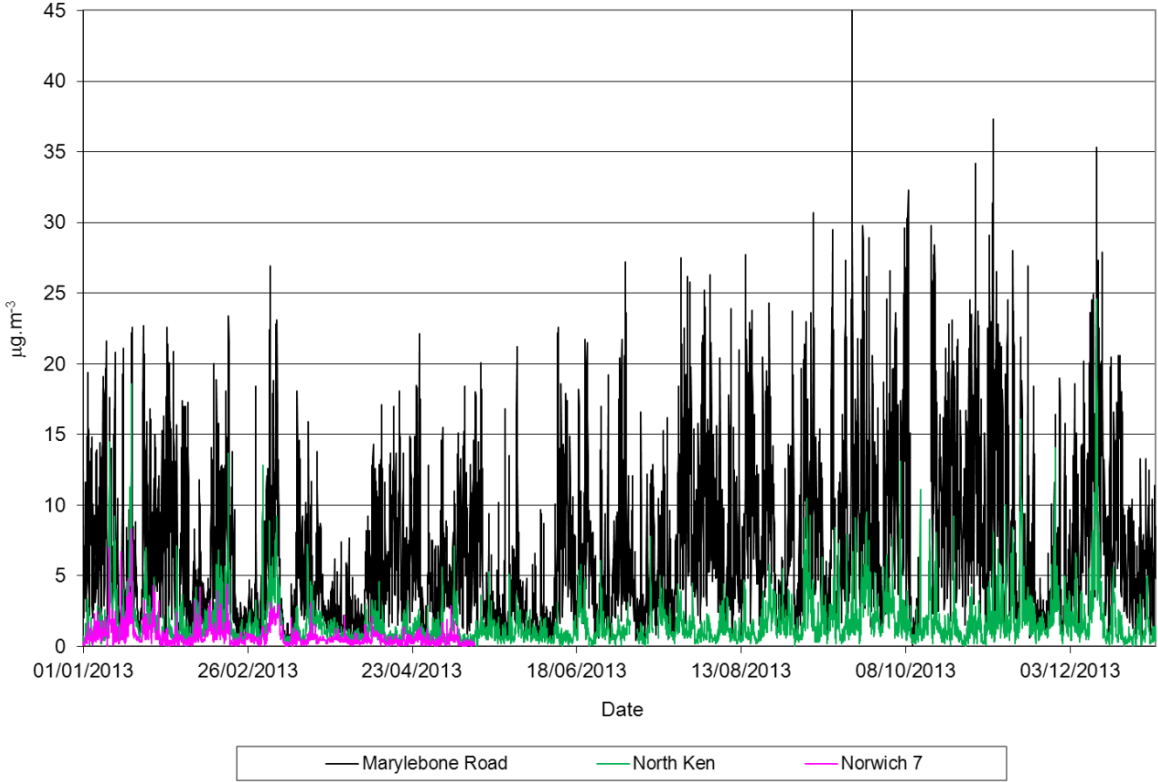


Figure 7 Black Carbon concentrations during 2013 in Southern and Eastern England

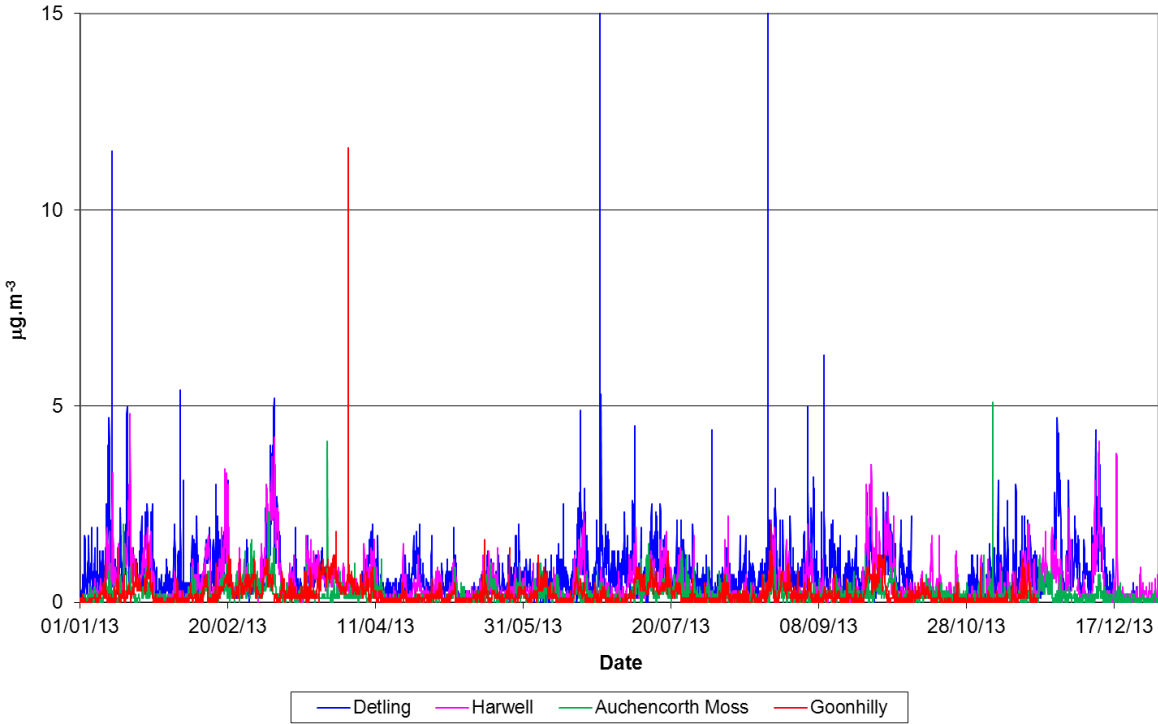


Figure 8 Black Carbon concentrations during 2013 at Rural Locations

Northern Irish sites measured increased concentrations during the cold months in January, February, November and December.

As before there is a noticeable drop in BC concentrations from late Christmas Eve at all sites until the end of December, probably due to reduced road traffic across the country.

The spike at Goonhilly in April is due to a local gorse fire upwind of the site, requiring the attendance of the fire brigade. The two spikes in the summer at Detling are probably due to localised burning of green waste.

Elevated Black Carbon concentrations were also measured on the evenings of 4th and 5th November at many sites across the UK, indicating emissions from celebrations associated with bonfire night. Especially high concentrations were recorded at Birmingham Roadside with concentrations reaching over $18 \mu\text{g.m}^{-3}$.

A detailed analysis of the Goonhilly data to determine the impact of shipping on black carbon concentrations will be written up and published.

5.1.2 UV component

The following charts show the UV component concentrations measured by the UK Black Carbon Network for 2013. The time resolution of the measurements is hourly. Data has been split into regions of the UK for presentation purposes. The y-axis on the UV component time series graphs have not been fixed to the same value for every chart, because the UV component is much more dependent on local site-specific conditions.

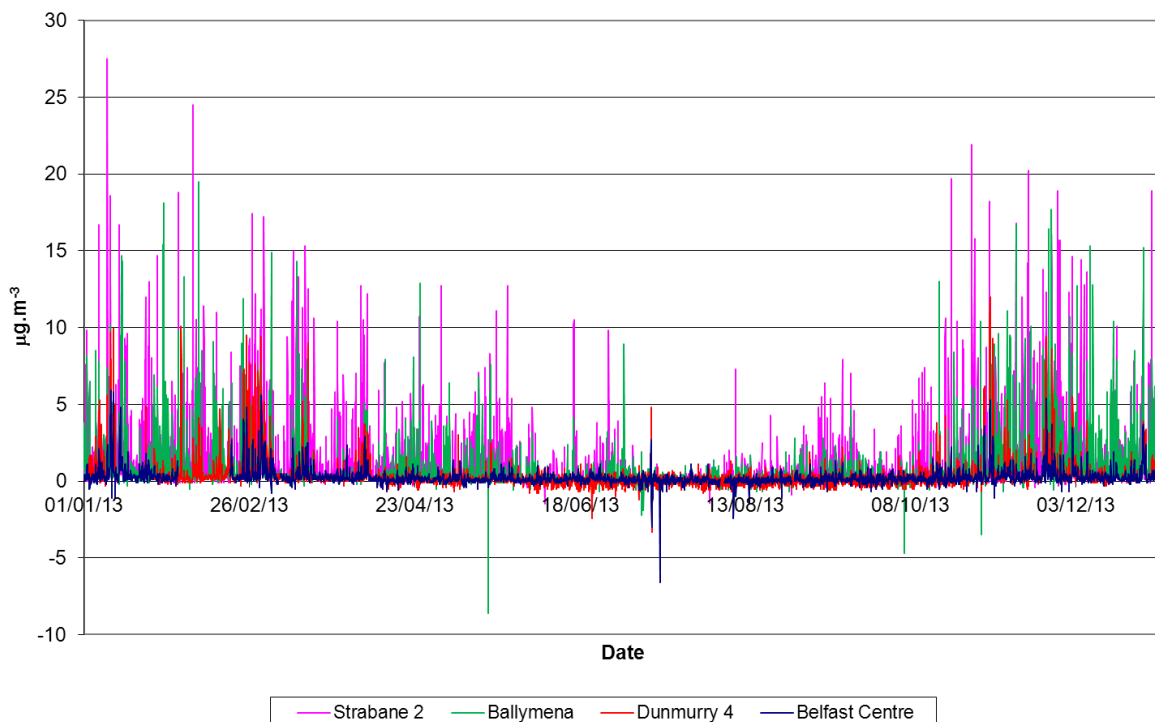


Figure 9 UV component concentrations during 2013 in Northern Ireland

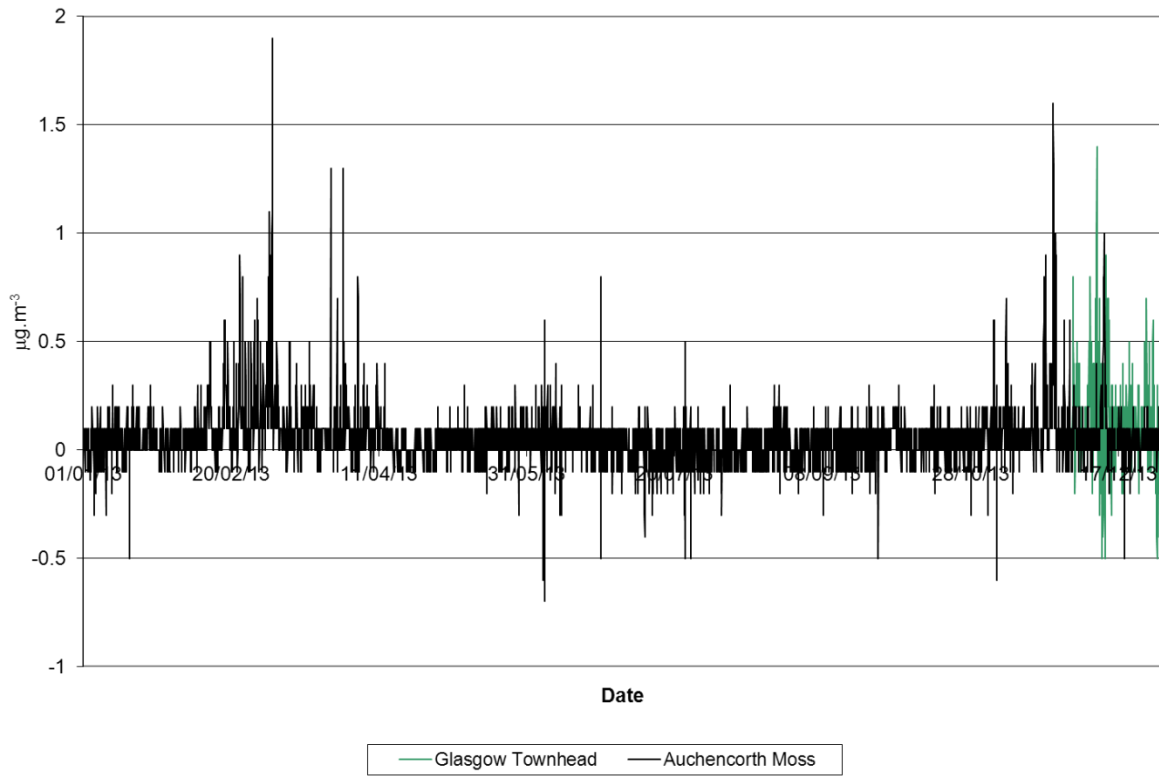


Figure 10 UV component concentrations during 2013 in Scotland

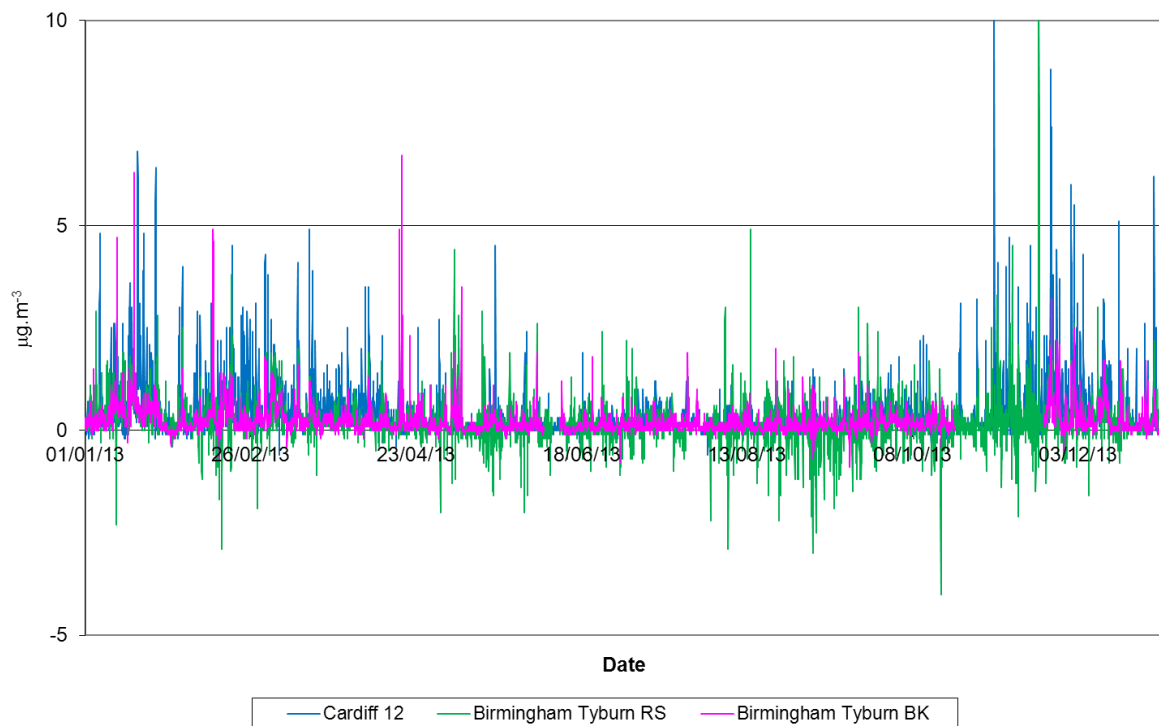


Figure 11 UV component concentrations during 2013 in Wales and the Midlands

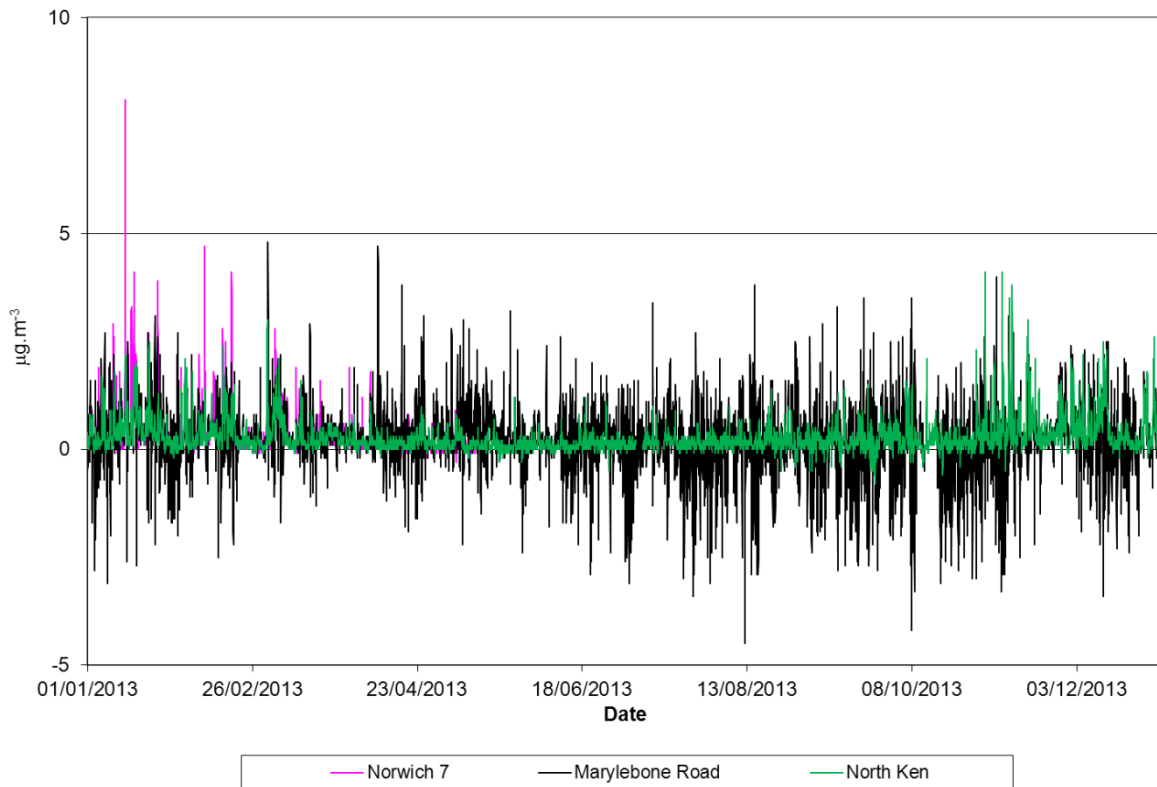


Figure 12 UV component concentrations during 2013 in Southern and Eastern England

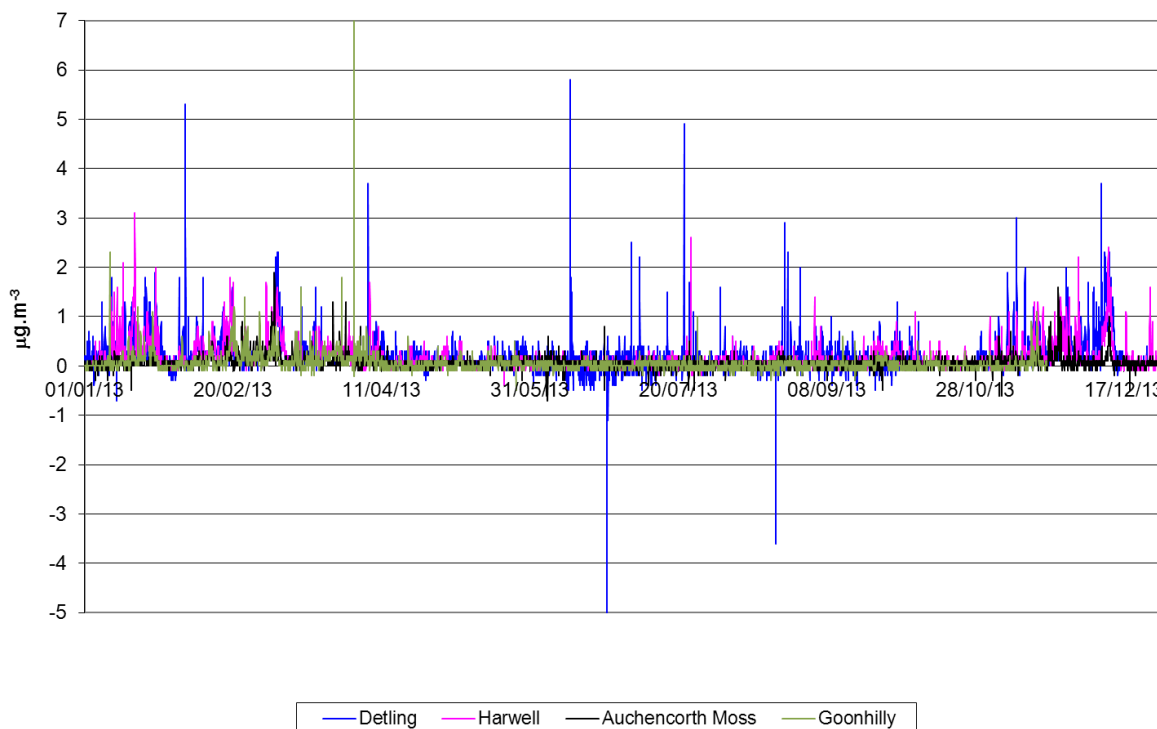


Figure 13 UV component concentrations during 2013 at Rural Locations

The Northern Irish sites measured increased UV component concentrations during the cold periods in January, February, November and December. Evidence from the UV component concentrations for 2013 suggests that the heating season runs from January to the end of May in Strabane and Ballymena

and from mid-September to the end of the year. Concentrations across 2013 were similar to those measured in 2012, but with a longer heating season.

Elevated UV component concentrations were also measured on the evenings of 4th and 5th November at many sites across the UK, indicating emissions from celebrations associated with bonfire night. Concentrations of approximately $12\mu\text{g.m}^{-3}$ were measured at the Dunmurry site.

A detailed analysis of the Goonhilly data will be written up and published in 2014.

The cause of the very short-term negative concentration spikes in the “UV component”, especially at roadside sites, is not clear. It may be due to the semi-volatile nature of the aromatic organic species that adsorb at the 370nm wavelength. Combustion exhaust streams may contain filterable particles at high concentrations together with semi-volatile UV-active material that will be temporarily retained on the filter tape leading to a distinct increase in UV absorption. Over time these organic species boil off the tape and reduce the enhanced UV adsorption. If equilibrium between organic species deposit and boil off is not reached due to cleaner air being sampled, the amount of organic enhancement will drop and will lead to negative UV component concentrations.

Another possible reason for positive and negative spikes in roadside data is the internal timing of the measurement process within the Aethalometer. The UV Channel reading is made around 20 seconds after the Black Carbon channel reading. If concentrations are changing rapidly, the subtraction of the Black Carbon concentration from the “UV” concentration could give misleading results.

These effects will be most prevalent at Marylebone Road due to the closeness of the inlet to the kerb, high traffic flow and predominantly diesel based exhaust signatures from HGVs, buses and taxis. A similar but smaller response is also found at Birmingham Tyburn Roadside, which still has high traffic flows but less diesel based exhaust emissions than Marylebone Road. This effect is not prevalent at Birmingham Tyburn Background which is approximately 60m away from the roadside site.

5.2 AVERAGES AND DATA CAPTURE

The following sections present the annual average Black Carbon and UV component concentrations along with the data capture statistics

5.2.1 Black Carbon

Table 6 gives the annual mean for each site for 2013.

Site	Mean concentration $\mu\text{g.m}^{-3}$
Auchencorth Moss	0.2
Ballymena	1.4
Belfast Centre	1.4
Birmingham Tyburn BK	1.4
Birmingham Tyburn RS	3.2
Cardiff	1.2
Detling	0.7
Dunmurry Kilmakee	1.0
Glasgow Townhead	0.9
Goonhilly	0.2
Harwell	0.5
Marylebone Road	6.9
North Kensington	1.7
Norwich Lakenfields	0.7
Strabane	1.5

Note: Glasgow Townhead, Norwich and Goonhilly are not full calendar year means.

Table 6 Annual Mean Black Carbon Concentrations for 2013

The annual mean concentrations are presented as a bar graph (Figure 14) to aid the comparison of sites:

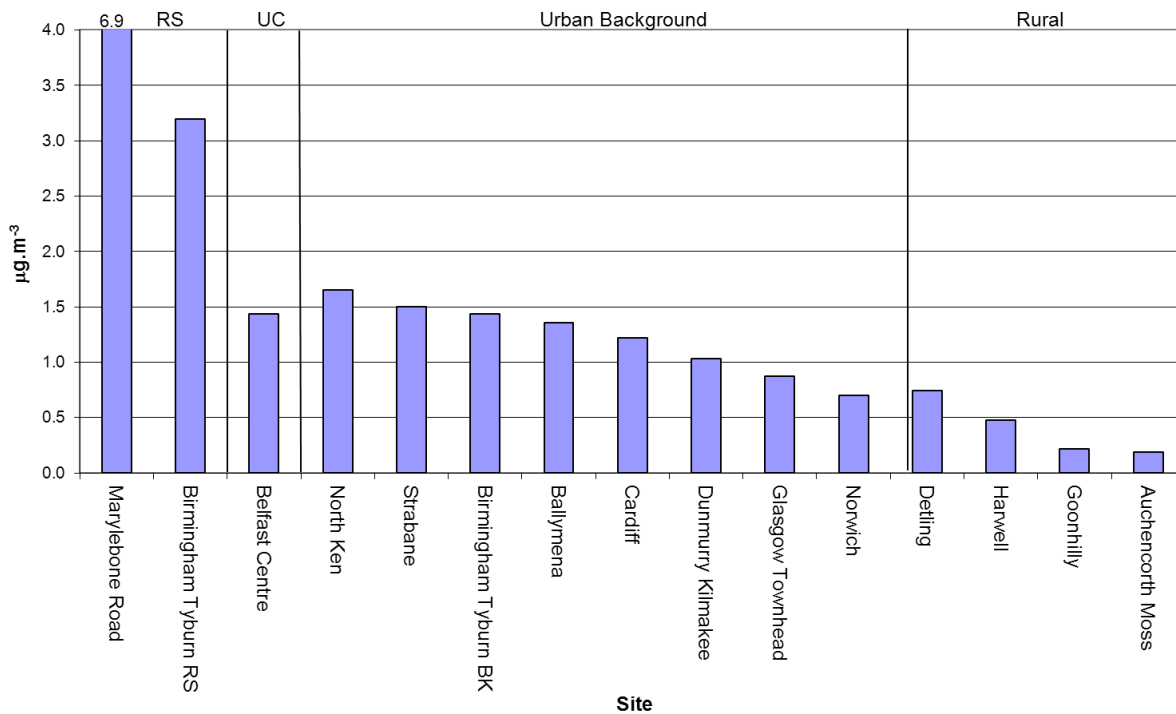


Figure 14 Annual Mean Black Carbon Concentrations for 2013

Black Carbon concentration increments between rural, urban background and roadside sites for London, Birmingham and Scotland have been analysed from periods when all the sites were producing valid data, and are shown in Table 7. The figures for 2012 are given in Table 8 for comparison.

Conurbation	Increment, µg.m ⁻³	
	Urban	Roadside
London	1.1	5.2
Birmingham	0.9	1.7
Scotland	0.8	N/A

Note The Scotland increment is only for the month of December due to the delays in the commissioning of the Glasgow Townhead site.

Table 7 Increments in Black Carbon concentrations between rural, background and roadside sites in 2013

Conurbation	Increment, µg.m ⁻³	
	Urban	Roadside
London	1.0	7.2
Birmingham	1.3	1.8
Scotland	1.2	N/A

Note: Birmingham Tyburn Roadside, Auchencorth Moss and Detling are not full calendar year means.

Table 8 Increments in Black Carbon concentrations between rural, background and roadside sites in 2012

The urban increment for London is very similar to 2012, while the roadside increment is 2.0 µg.m⁻³

(28%) less than in 2012. The urban increment at Birmingham is $0.4 \mu\text{g.m}^{-3}$ (50%) lower while the roadside increment is similar to 2012. The difference in the Scotland urban increment is not significant due to the low time coverage of the Glasgow Townhead site.

It can be seen that the urban increment for all three areas is similar, while the roadside increment for London is much larger than that for Birmingham. This is due to the much larger traffic flow and different vehicle profile of the Marylebone Road site compared to the Birmingham Tyburn roadside site. Highways Agency traffic count data for 2013 for the two roads passing the monitoring stations are given in Table 9.

Road (Count Point ID)	Motor cycles	Cars Taxis	Buses Coaches	All HGVs	All Motor Vehicles
Marylebone Road (27236)	3491	50954	2718	3376	71754
Tyburn Road (56399)	136	25032	236	1559	31415
Ratio	25.7	2.0	11.5	2.2	2.3

Table 9 2013 Average daily traffic count data for Marylebone and Tyburn Roads

The Marylebone Road roadside increment in Black Carbon concentration in 2013 was a factor of 3 higher than the Tyburn Road increment, somewhat higher than the ratio of numbers of cars / taxis and HGVs, between the sites at 2.0 and 2.2 respectively. However there are 11.5 times more buses and coaches and 25.7 times more motor cycles passing the Marylebone Road site compared to the Birmingham site, which would indicate that these are probably the predominant source of black carbon emissions at Marylebone Road.

5.2.2 UV component

Table 10 gives the annual average for each site for 2013.

Site	Mean concentration $\mu\text{g.m}^{-3}$
Auchencorth Moss	0.1
Ballymena	0.8
Belfast Centre	0.2
Birmingham Tyburn BK	0.2
Birmingham Tyburn RS	0.2
Cardiff	0.4
Detling	0.2
Dunmurry Kilmakee	0.4
Glasgow Townhead	0.1
Goonhilly	0.0
Harwell	0.2
Marylebone Road	0.2
North Kensington	0.3
Norwich Lakenfields	0.3
Strabane	1.2

Note: Glasgow Townhead, Norwich and Goonhilly are not full calendar year means.

Table 10 Annual Mean UV component Concentrations for 2013

The annual mean concentrations are presented as a bar graph (Figure 15) to aid the comparison of sites:

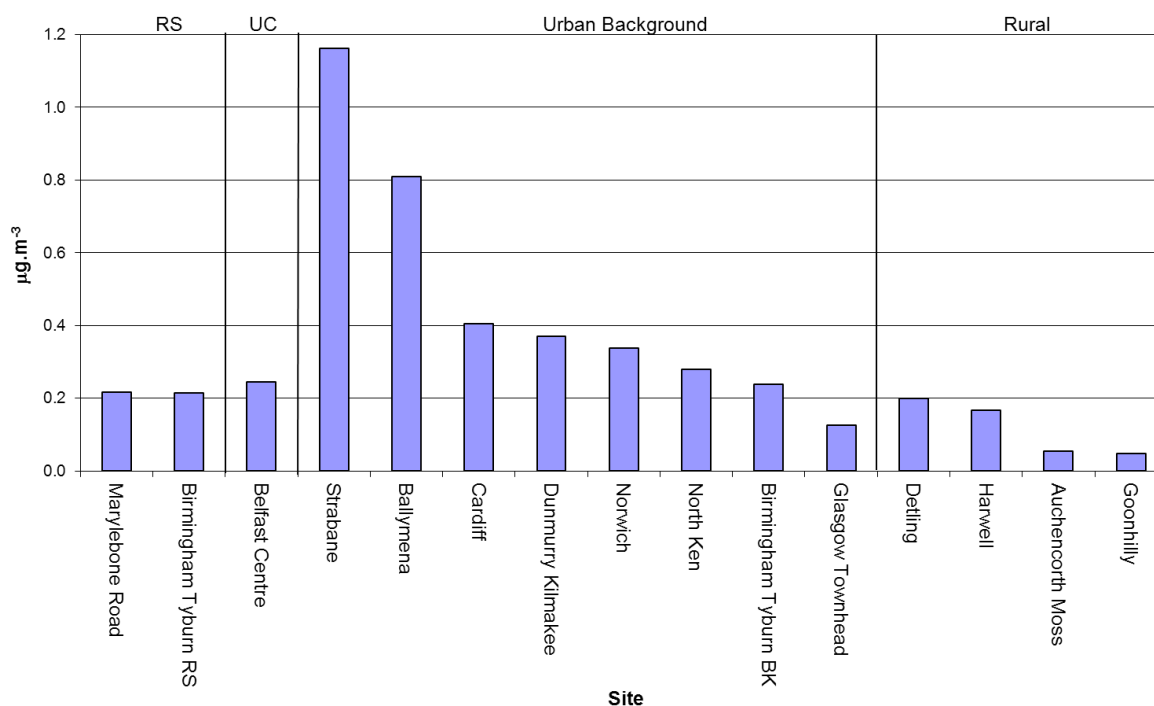


Figure 15 Annual Mean UV component Concentrations for 2013

UV component concentration increments between rural, urban background / centre and roadside sites for London, Birmingham and Scotland have been analysed from periods when all the sites were producing valid data, and are shown in Table 11.

Conurbation	Increment, $\mu\text{g.m}^{-3}$	
	Urban	Roadside
London	0.1	-0.1
Birmingham	0.1	0.0
Scotland	0.1	N/A

Table 11 Increments in UV component concentrations between rural, background and roadside sites in 2013

There is no significant difference in increments between 2012 and 2013.

It can be seen that the urban and roadside increments at all sites are similar, as domestic emissions in the three areas are of a similar magnitude due to similar fuel types and that road traffic is not a significant source for the UV component.

Using the same method, the urban increment in UV component concentration in Northern Ireland has been calculated relative to Belfast where gas heating has largely displaced oil and coal. The results shown in Table 12.

Site	Increment compared to Belfast, $\mu\text{g.m}^{-3}$
Dunmurry	0.1
Ballymena	0.6
Strabane	0.9

Table 12 Increment in UV component concentration in Northern Ireland

The increments at Dunmurry, Ballymena and Strabane are not surprising as domestic heating in Belfast is predominantly gas fired, while there is a history of solid fuel usage for secondary heating in Dunmurry, and a significant usage of non-smokeless fuel usage in Strabane. Ballymena is supplied by natural gas, but the benzo[a]pyrene (BaP) concentrations measured by the PAH Network are higher than might be expected. A similar increase in the UV component is also found here. Correlations between BaP and the UV component can be seen in Section 5.4.2. It can therefore be concluded that there is a significant emission source of both these components in the Ballymena area, probably from the use of non-smokeless fuel. The monitoring site is not within Ballymena's smoke control zone but is within its Air Quality Management Area for PM_{10} .

5.2.3 Data Capture

Table 13 gives the data capture for each site for 2013. Due to the Network changes during 2013 the time coverage for the complete calendar year for each site has also been given.

Site	Data Capture %	Time Coverage %
Auchencorth Moss	99	99
Ballymena	99	99
Belfast Centre	94	94
Birmingham Tyburn BK	90	90
Birmingham Tyburn RS	98	98
Cardiff	99	99
Detling	88	88
Dunmurry Kilmakee	93	93
Glasgow Townhead	100	8
Goonhilly	99	88
Harwell	99	99
Marylebone Road	99	99
North Kensington	99	99
Norwich Lakenfields	100	36
Strabane	99	99

Table 13 Data capture rates of the Aethalometers for 2013

The average data capture for the Network is 97% and there is only one site, Detling (88%), with a data capture of below 90%. The Aethalometer installed at Detling had two major breakdowns during the year, both of which required the analyser to be swapped with a hot spare.

The sites with time coverage less than 90% are those sites that either opened or closed part way through 2013.

5.3 TEMPORAL VARIATIONS

The following section presents analysis of the 2013 Black Carbon and UV component concentrations with respect to the hour of the day. Charts of variations over the day of the week and the month in the year are made using the data from 2009 – 2013, to avoid bias introduced by single year measurements. Y-axes vary by site.

All results have been grouped by site classification. The site order for the Roadside and Urban Centre sites is by decreasing Black Carbon concentration, while the site order for the Urban Background and Rural Background sites is by decreasing UV component concentration. The units on the y-axis are $\mu\text{g.m}^{-3}$ for Black Carbon and equivalent $\mu\text{g.m}^{-3}$ for the UV component.

The data has been plotted in local time (GMT/BST), as the pollution sources are primarily attributed to human activity sources and not solar driven atmospheric chemistry sources.

The 2013 data are presented in Figures 16 to 19.

The 2009-2013 data is presented in Figures 20 to 23.

Chart Key

For all of the charts, the continuous central line is the mean value and the shaded area about this line represents the uncertainty in the mean y-value due to the spread of the results over that averaging period, expressed with a level of confidence of 95%. It is not the overall measurement uncertainty. The shaded area on the x-axis in the by month chart (bottom centre) is for display purposes only, to allow the uncertainty in the mean value to be seen more clearly.

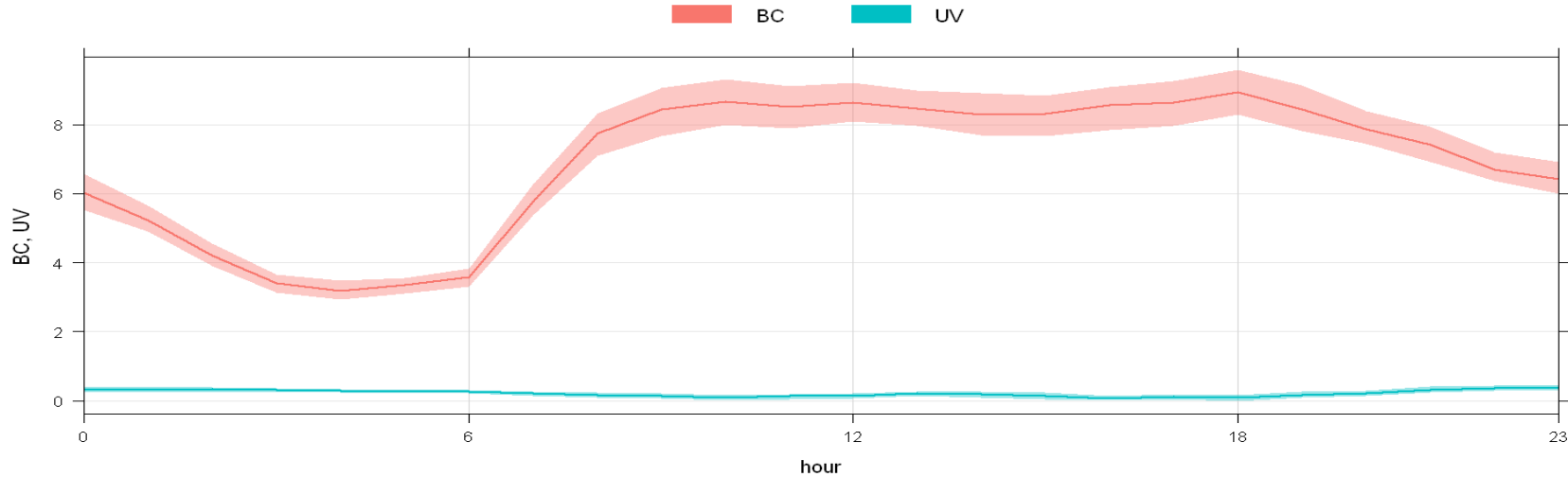
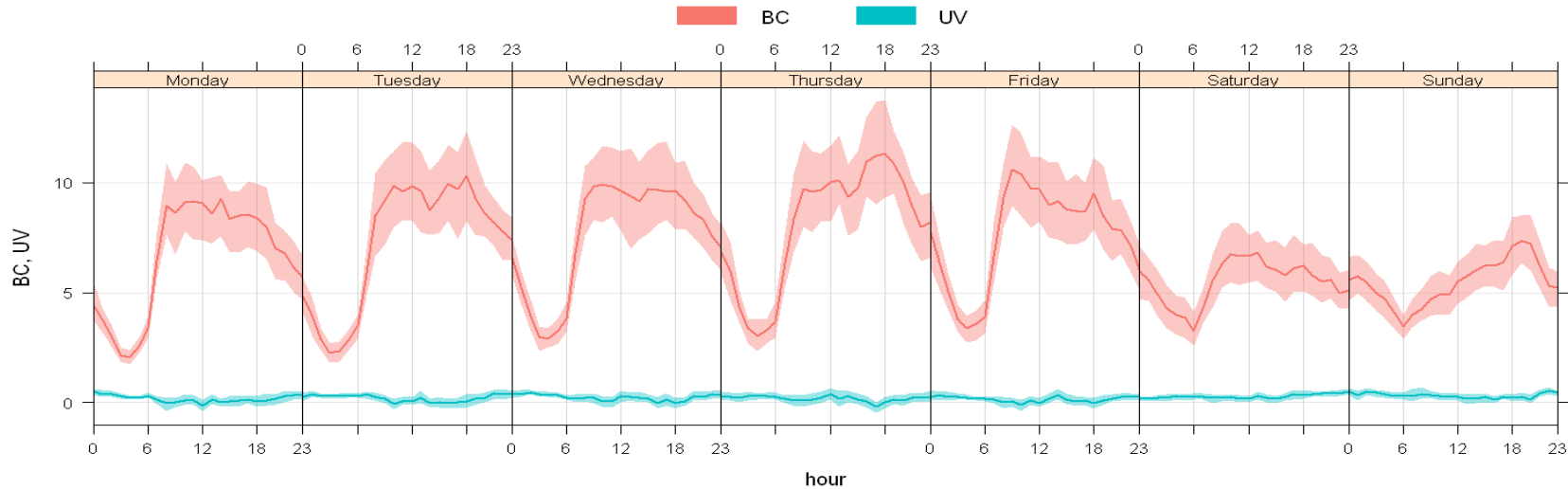
Acknowledgement

Figures 16 to 23 are generated using the Open-Air Tools run on the R software platform^{3,4}.

3 DC Carslaw and K Ropkins, (2012) OpenAir --- an R package for air quality data analysis, Environmental Modelling & Software. Volume 27-28, 52-61.

4 DC Carslaw and K Ropkins (2012). OpenAir: Open-source tools for the analysis of air pollution data, R package version 0.5-23

2013 Data



Marylebone Road

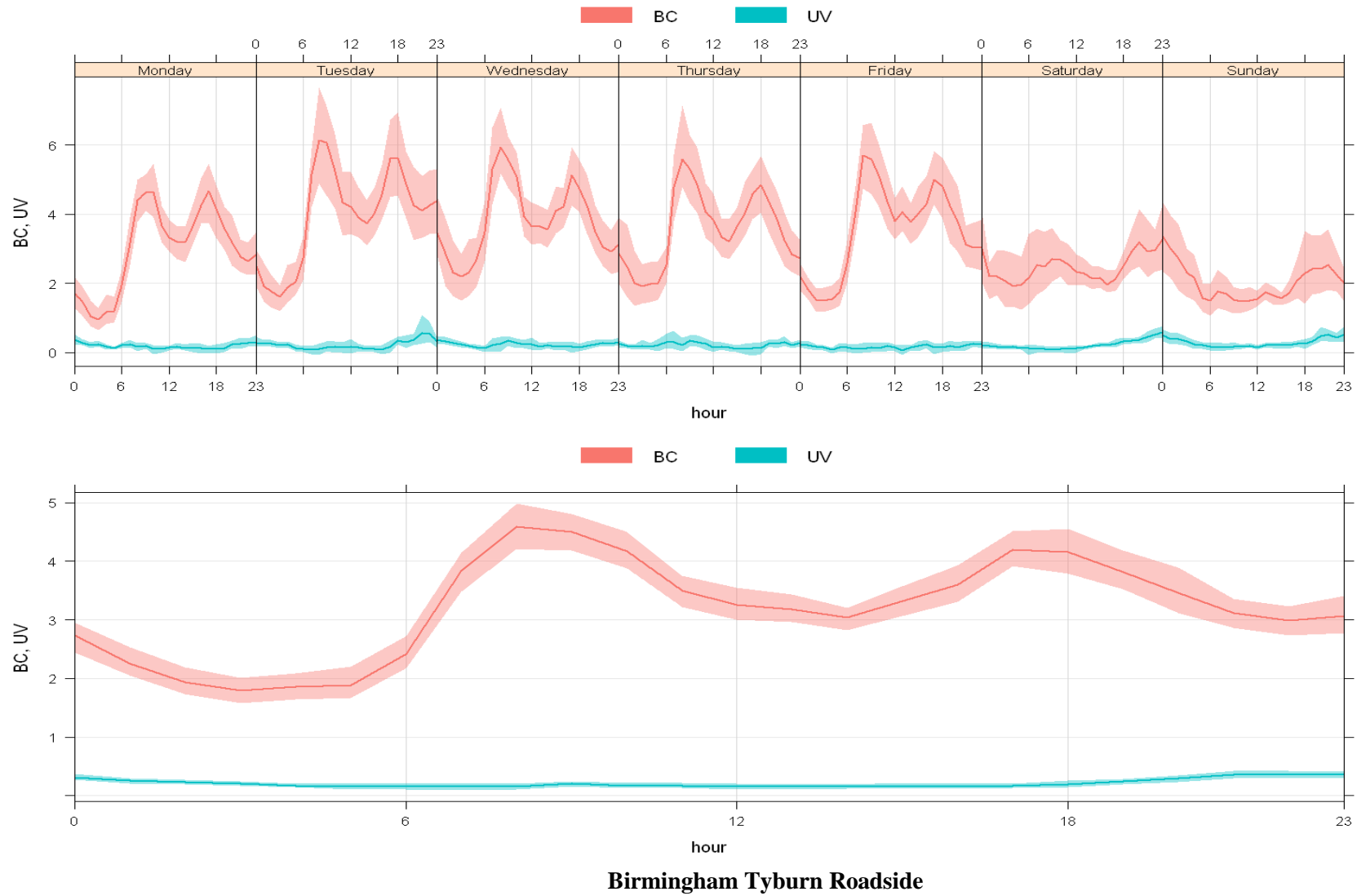
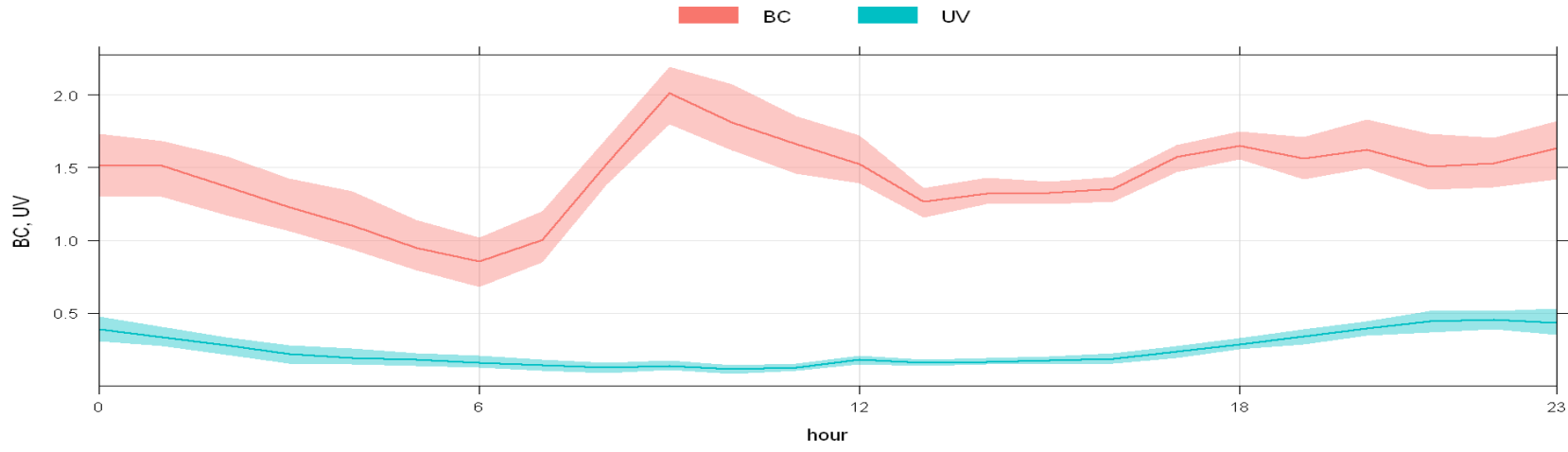
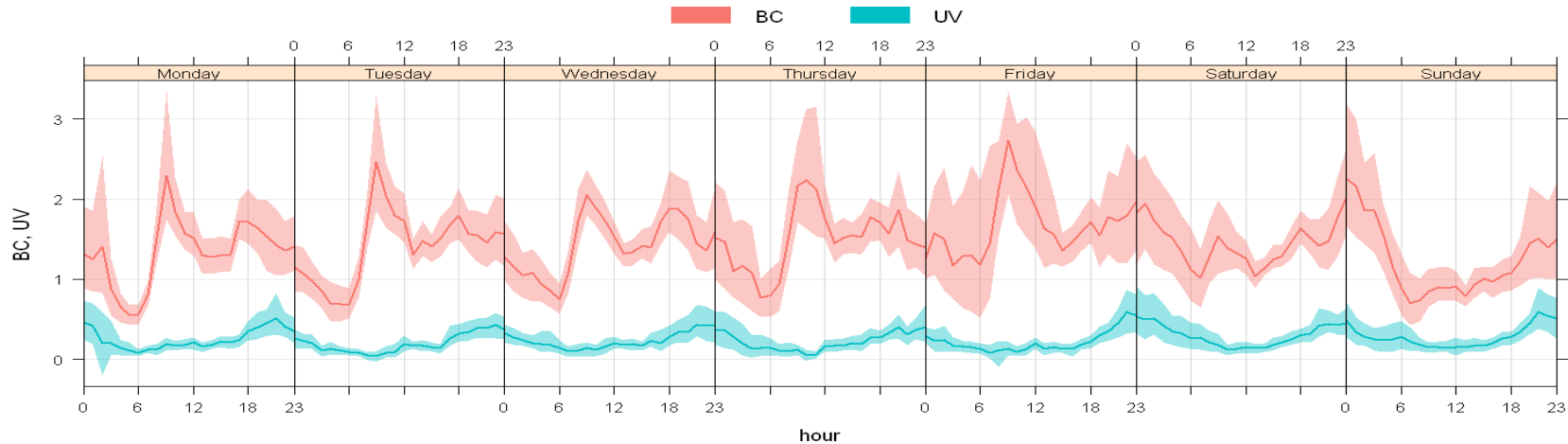
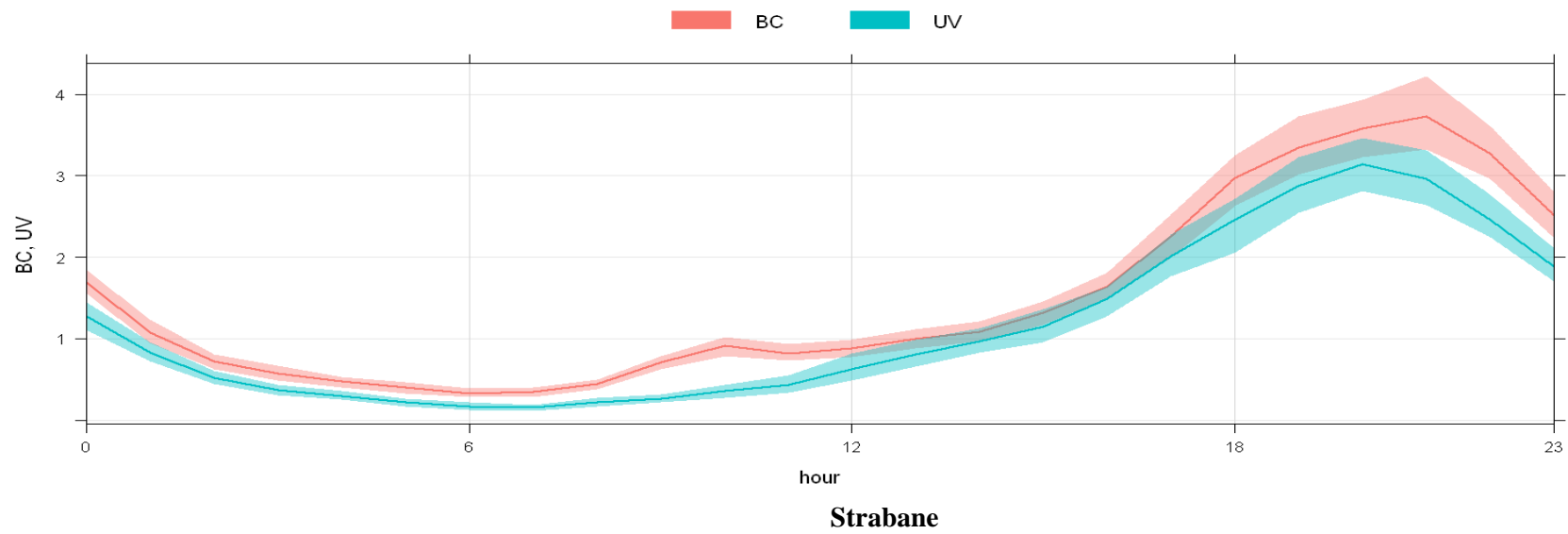
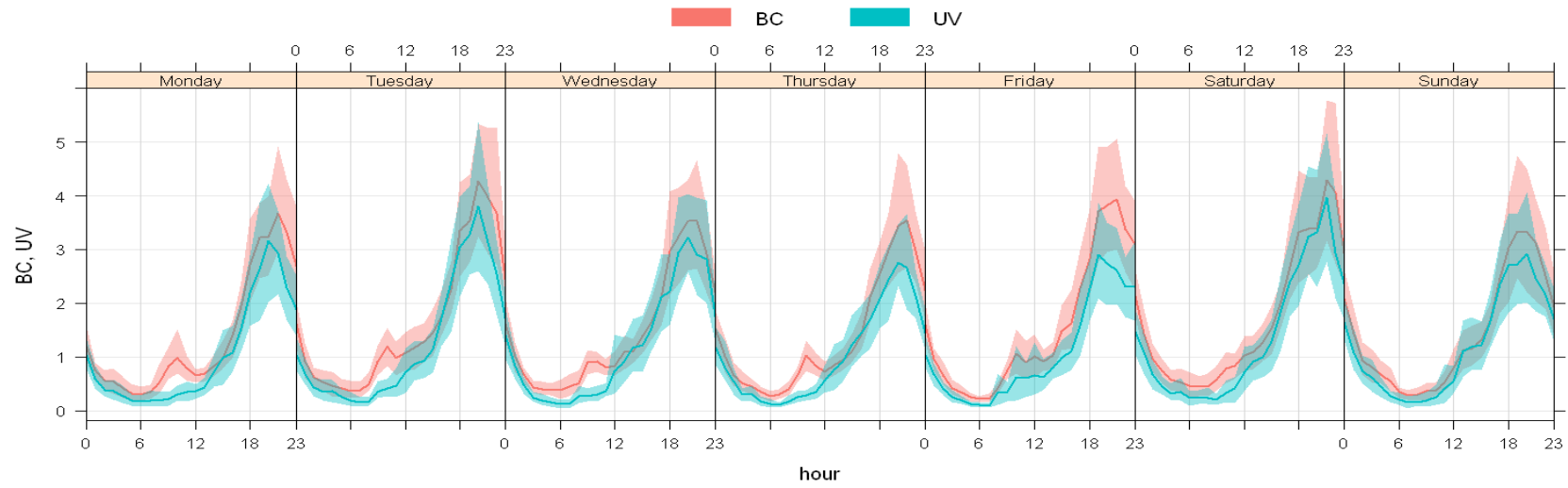


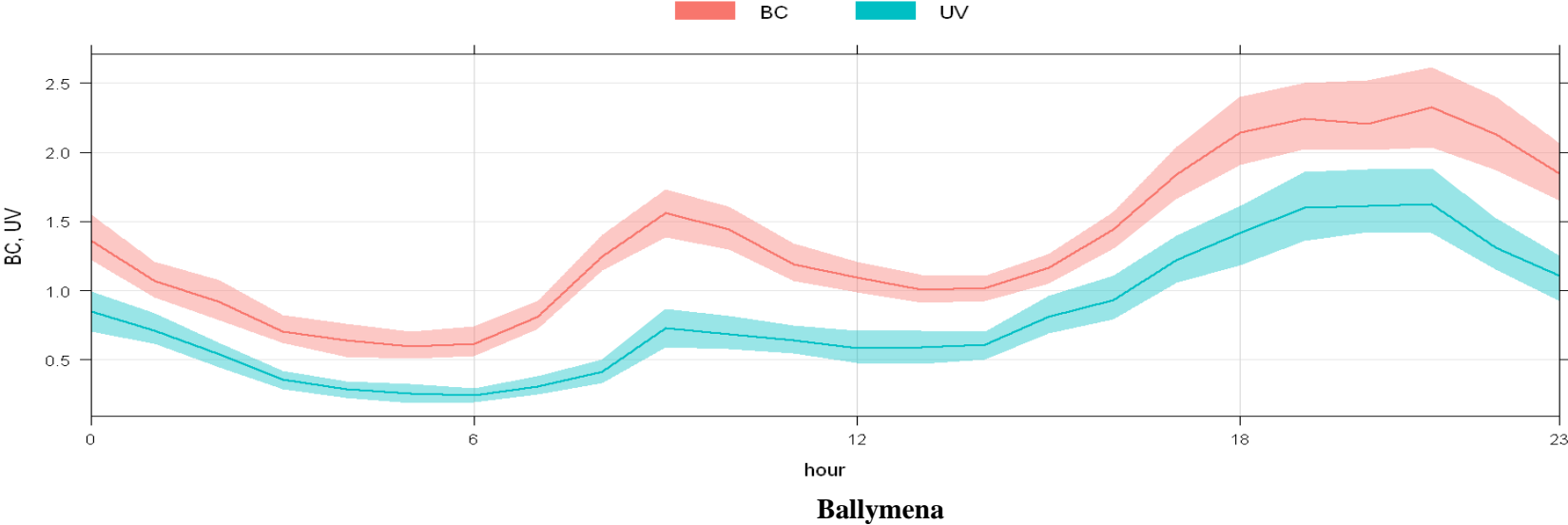
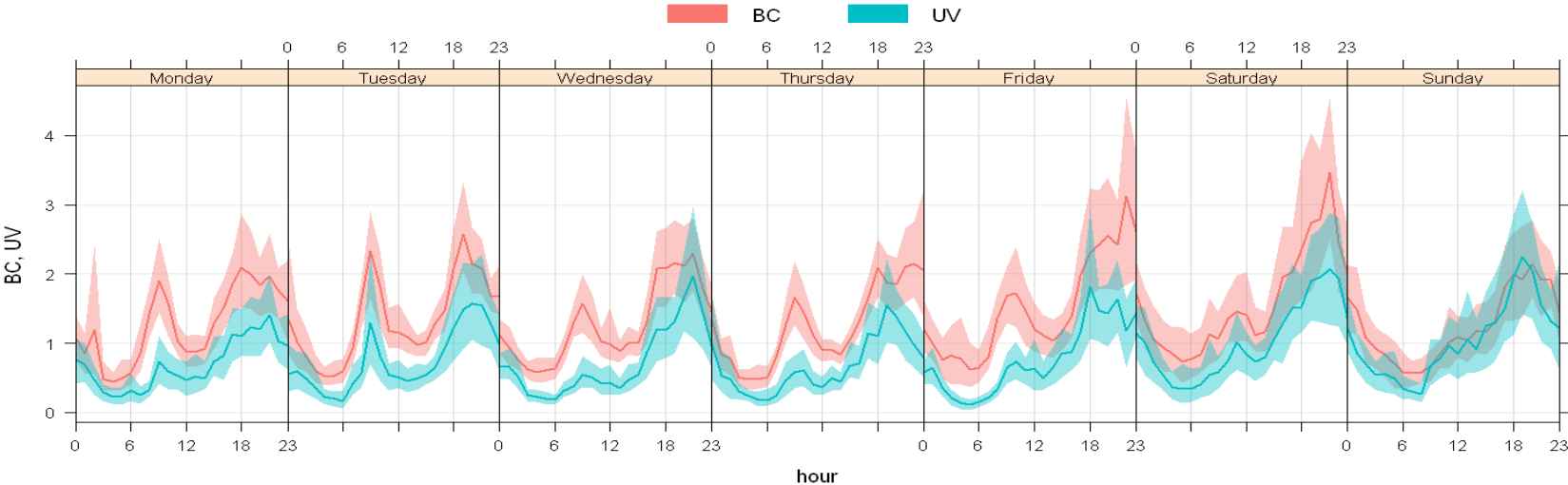
Figure 16 2013 Roadside sites

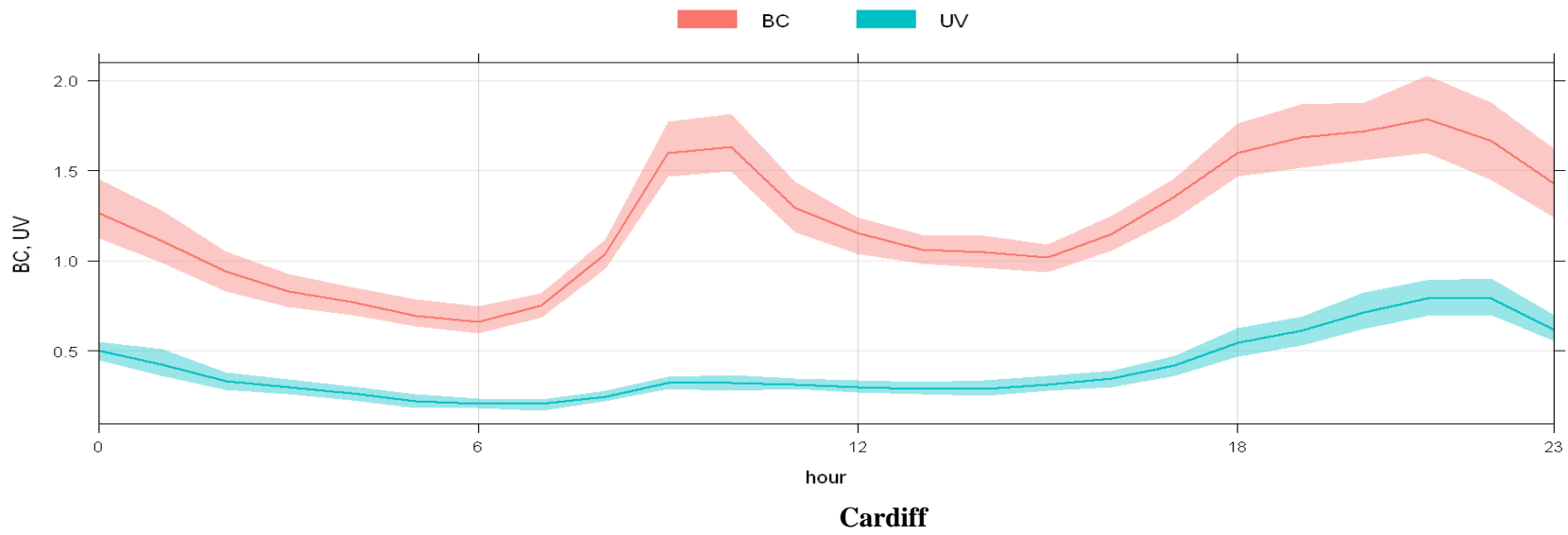
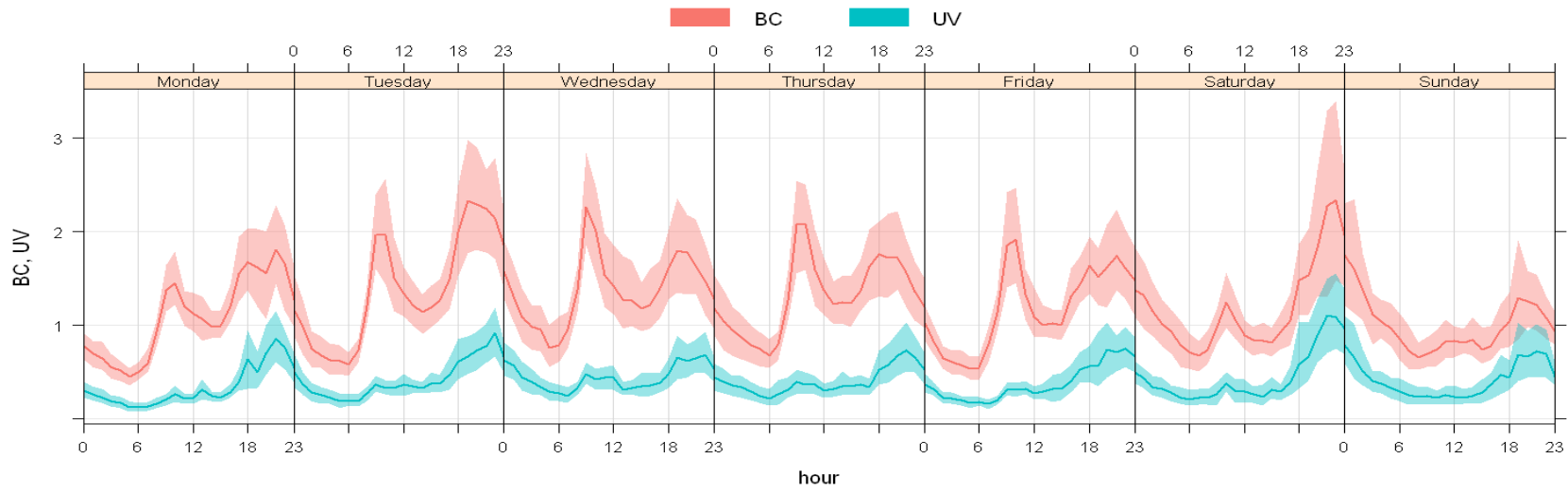


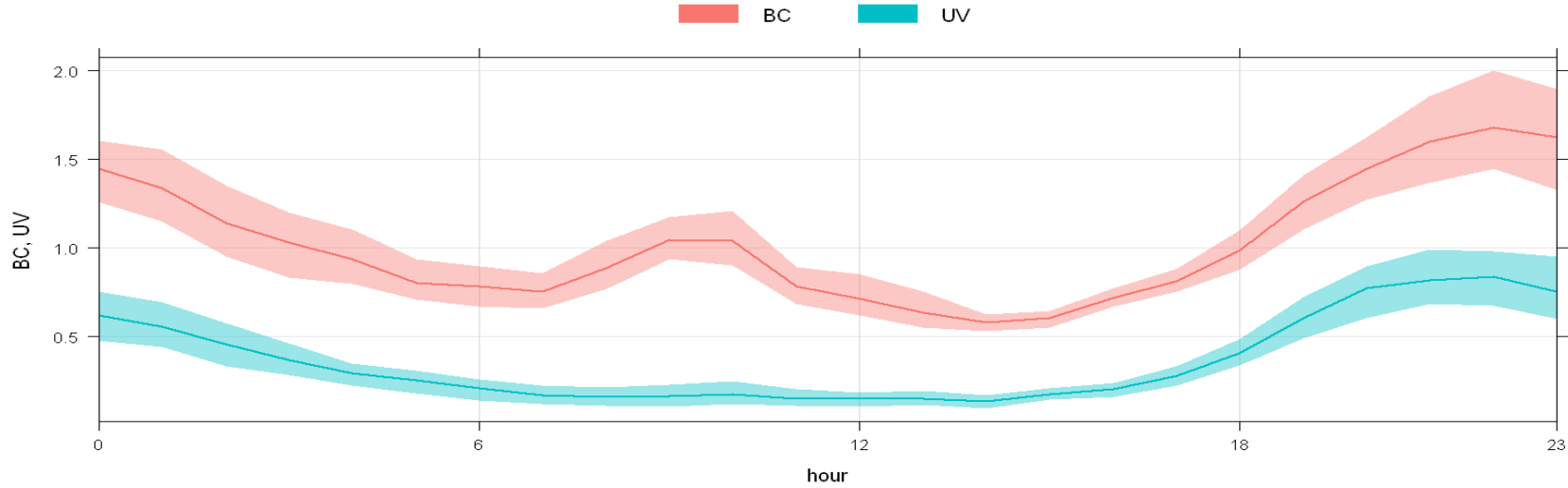
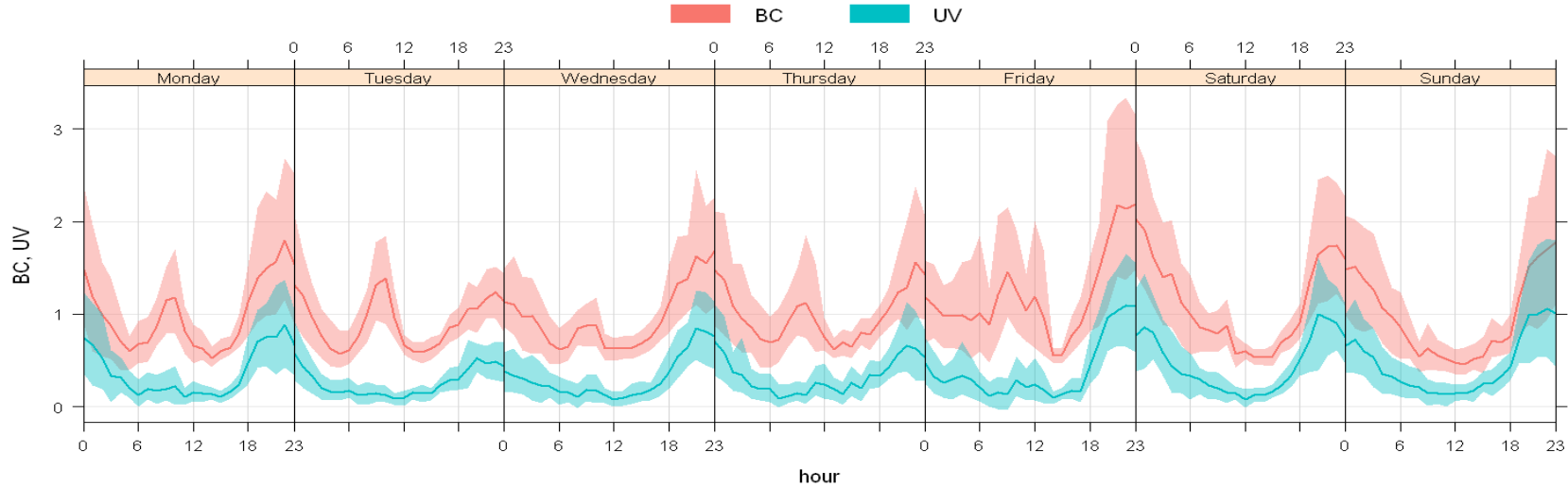
Belfast Centre

Figure 17 2013 Urban Centre sites

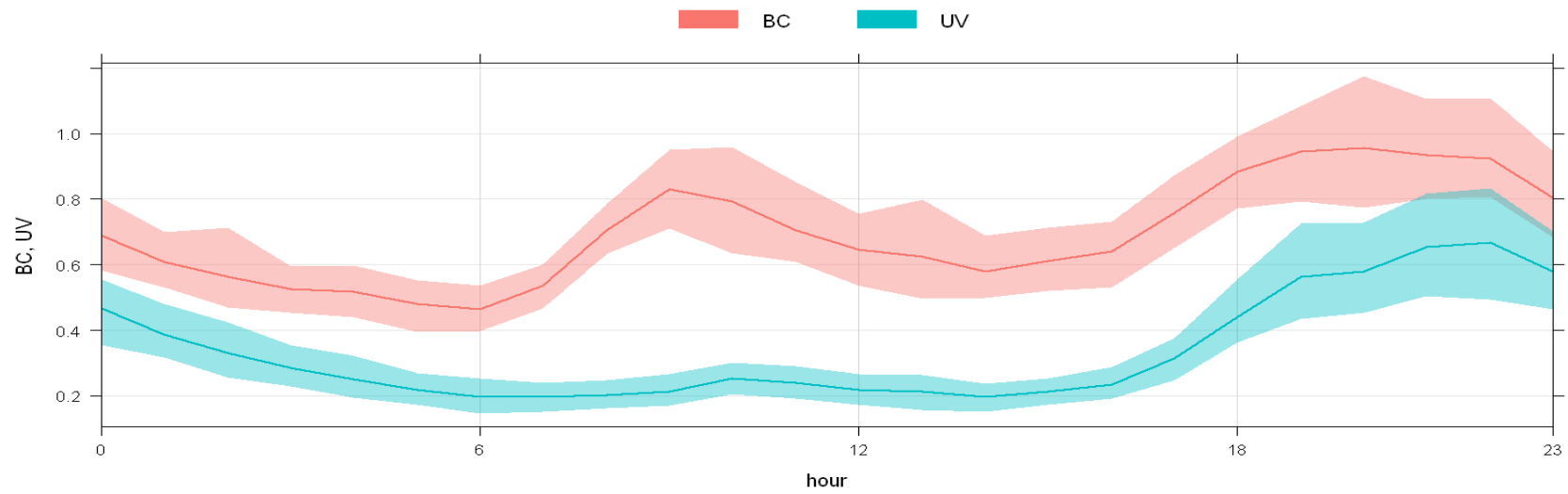
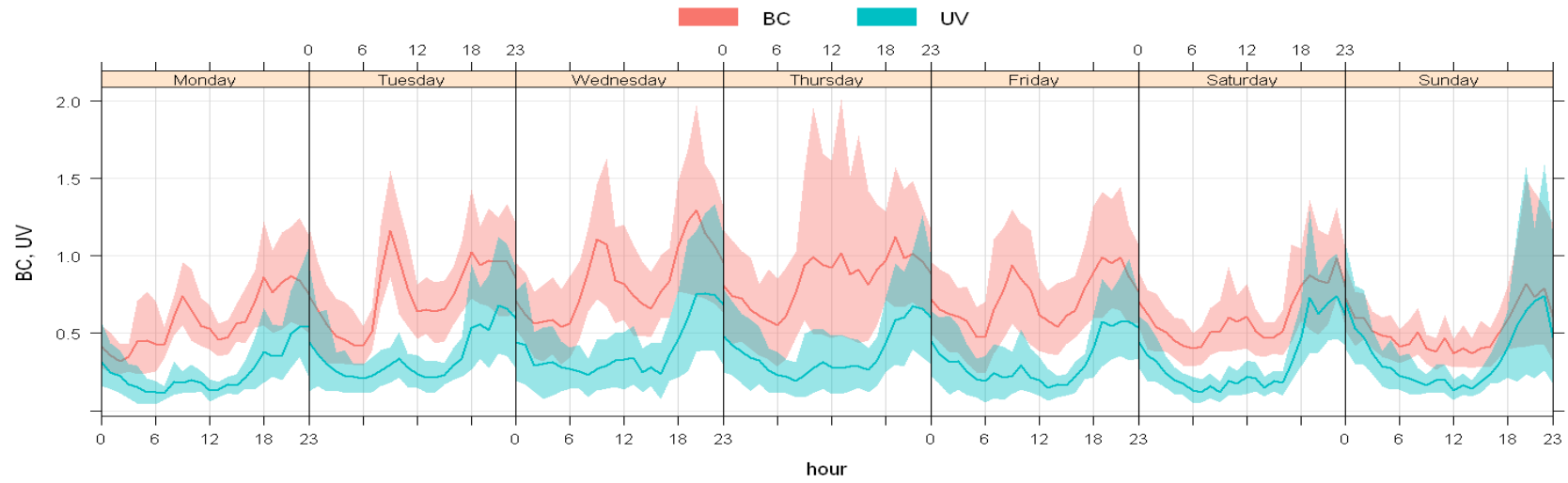




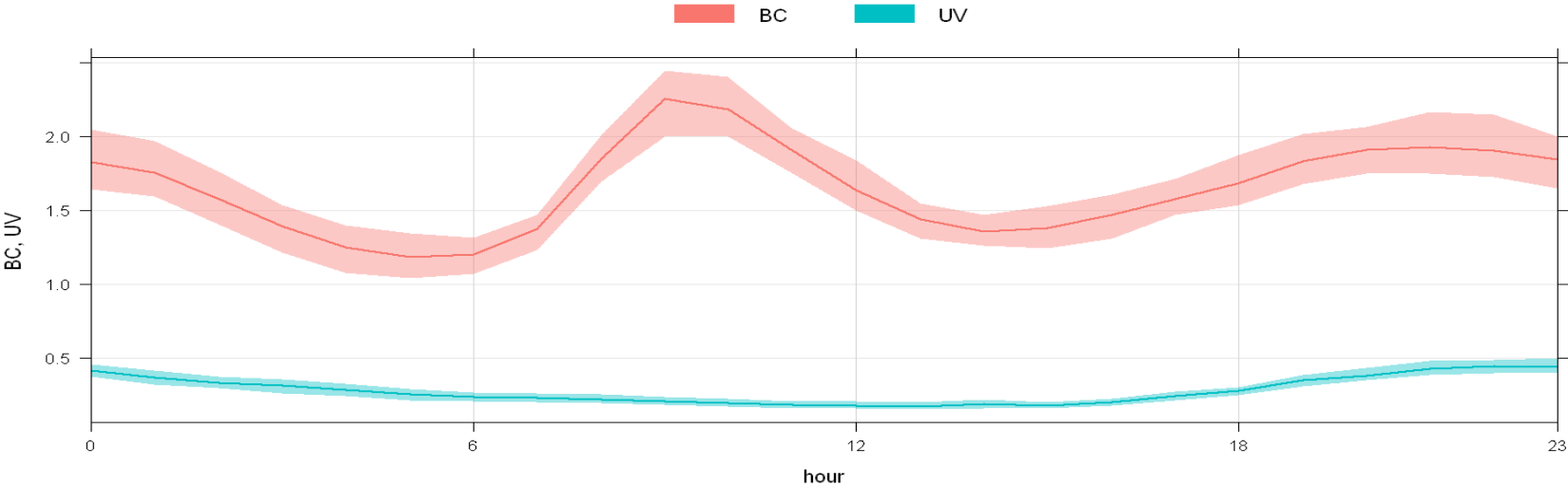
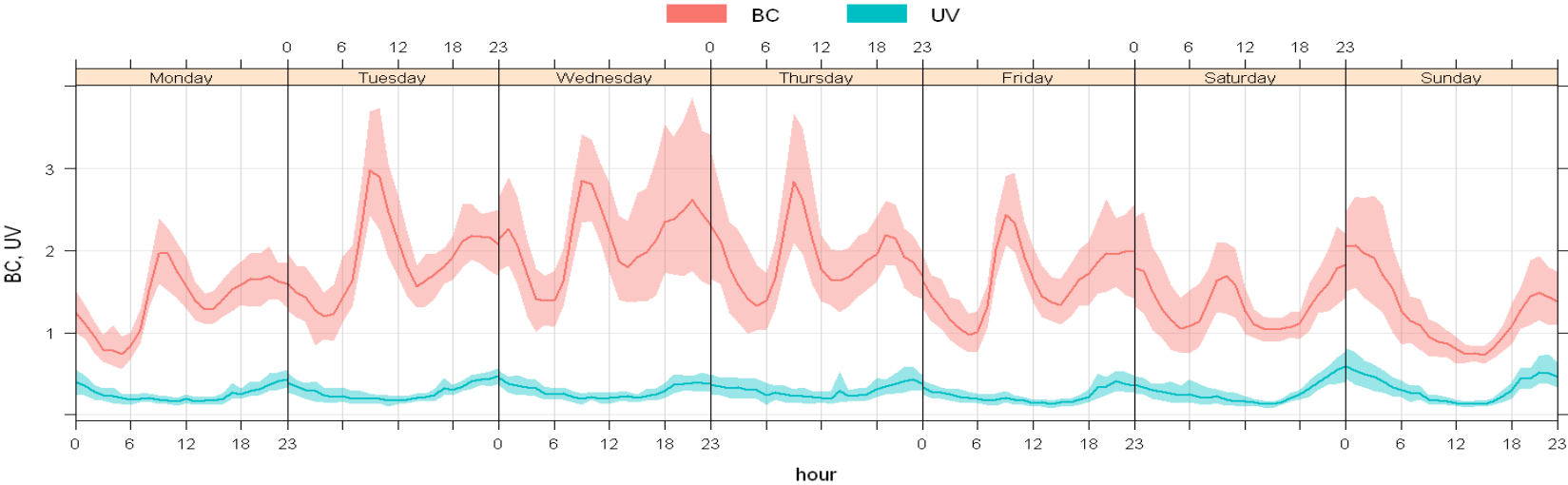




Dunmurry Kilmakee



Norwich Lakenfields



North Kensington

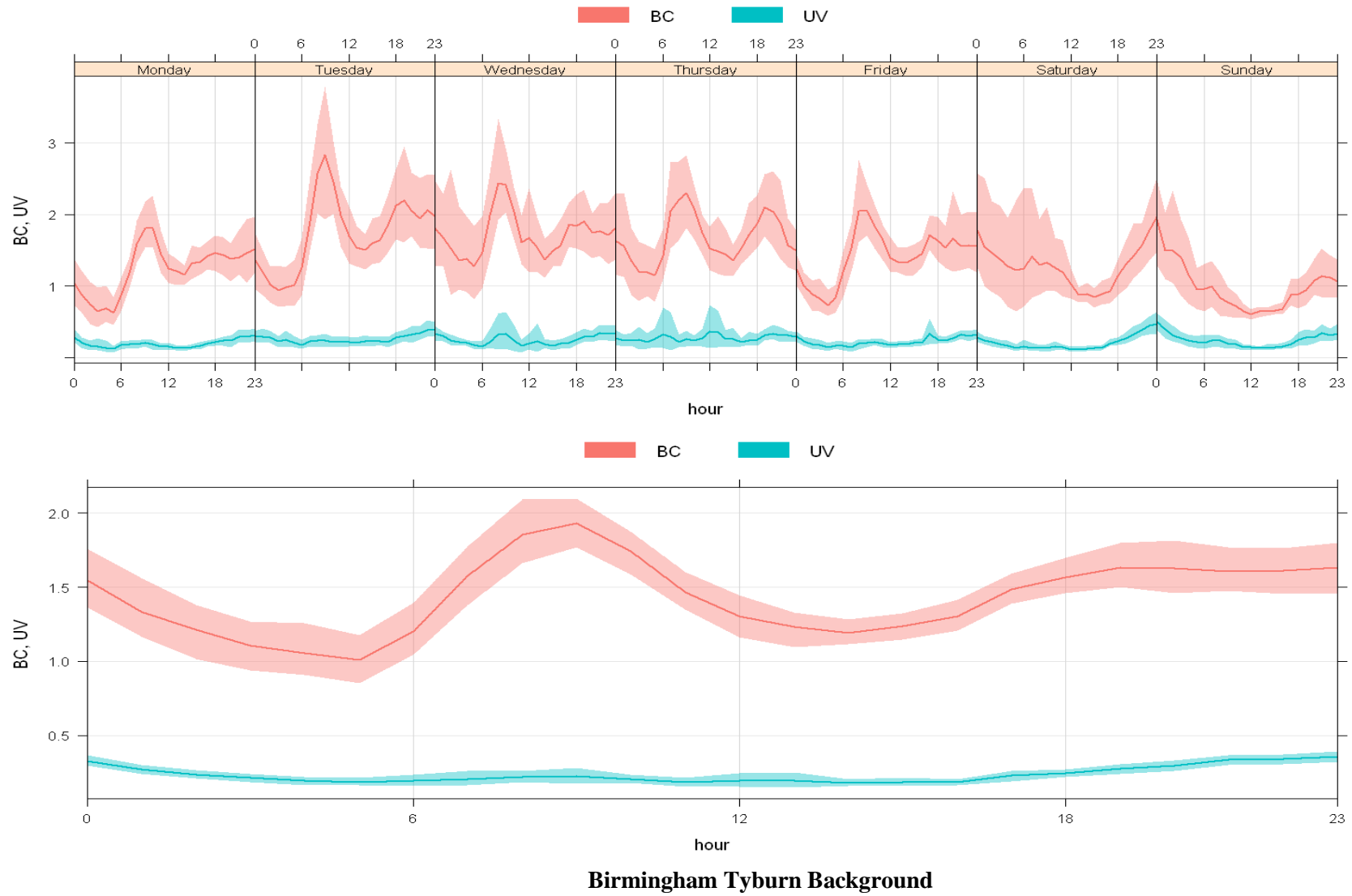
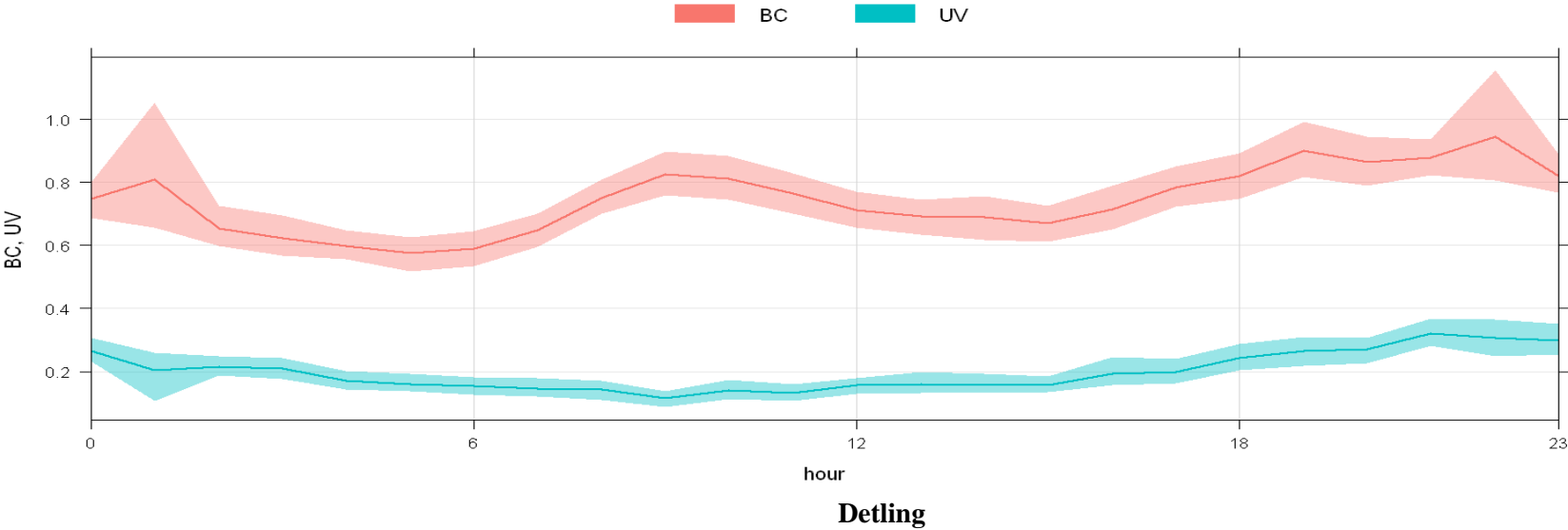
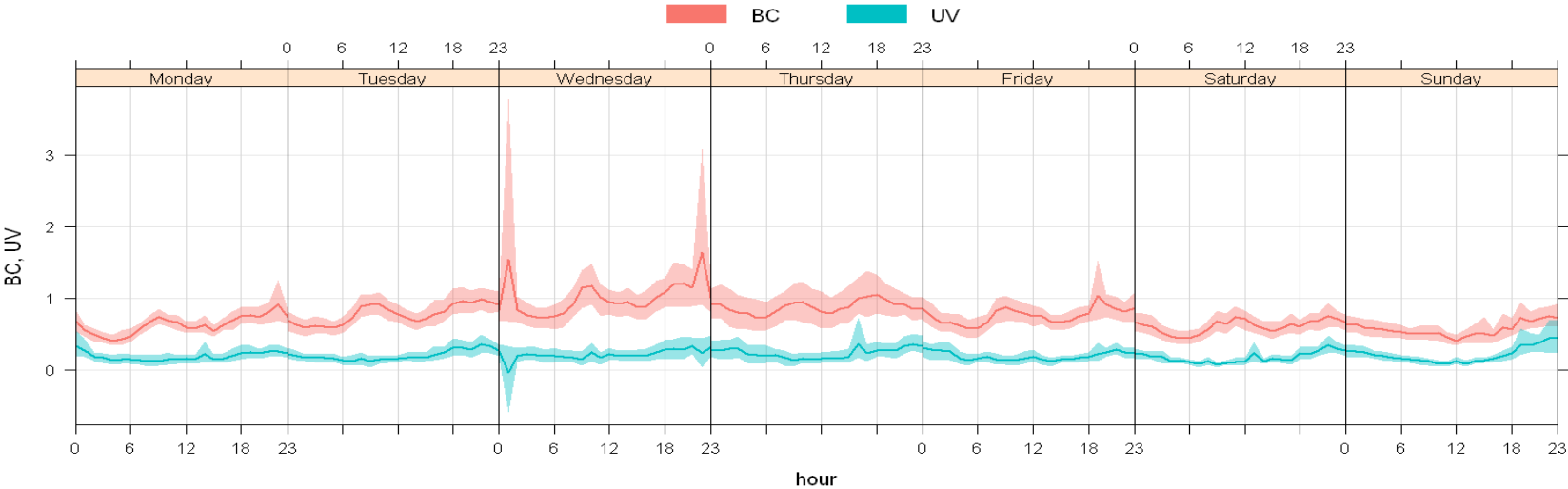
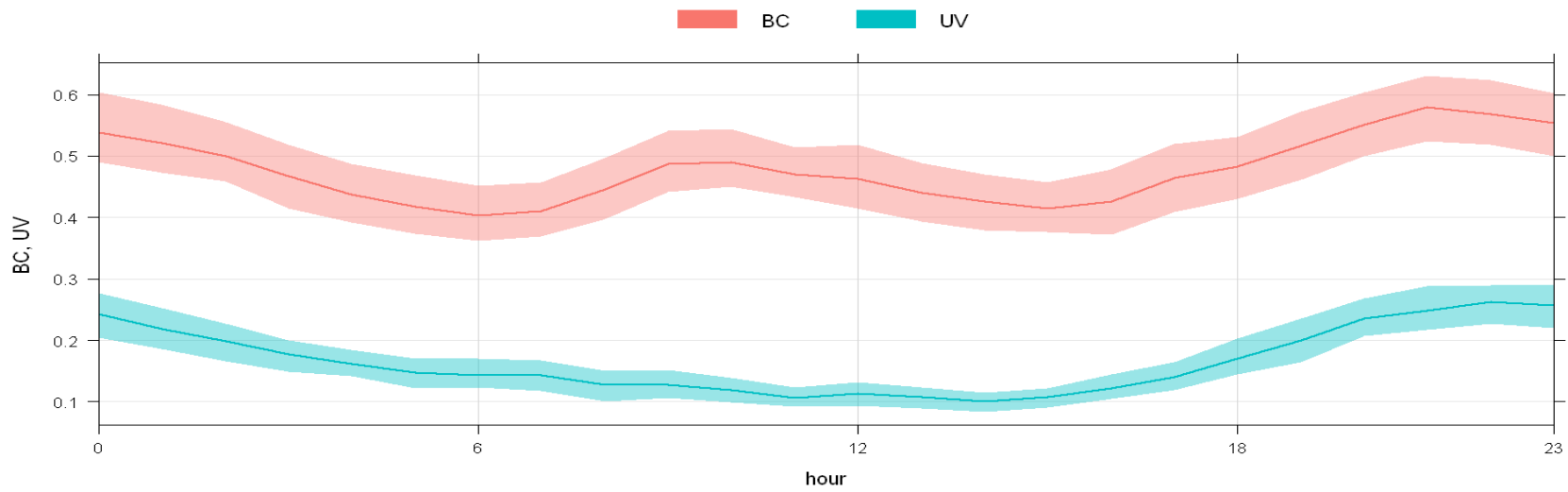
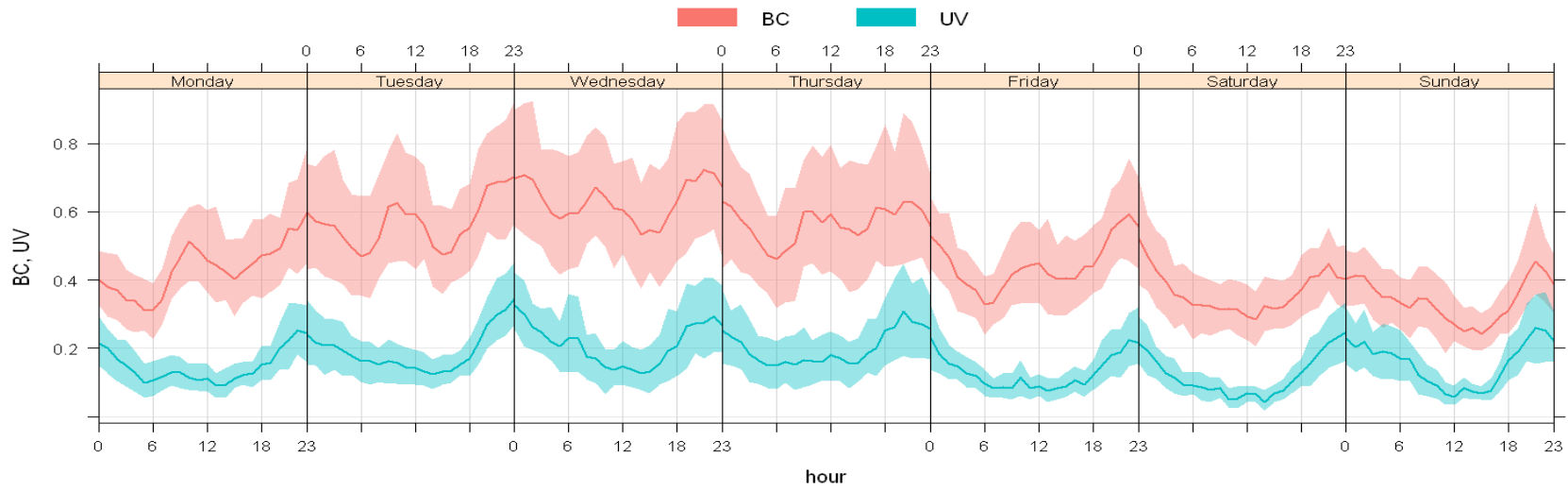
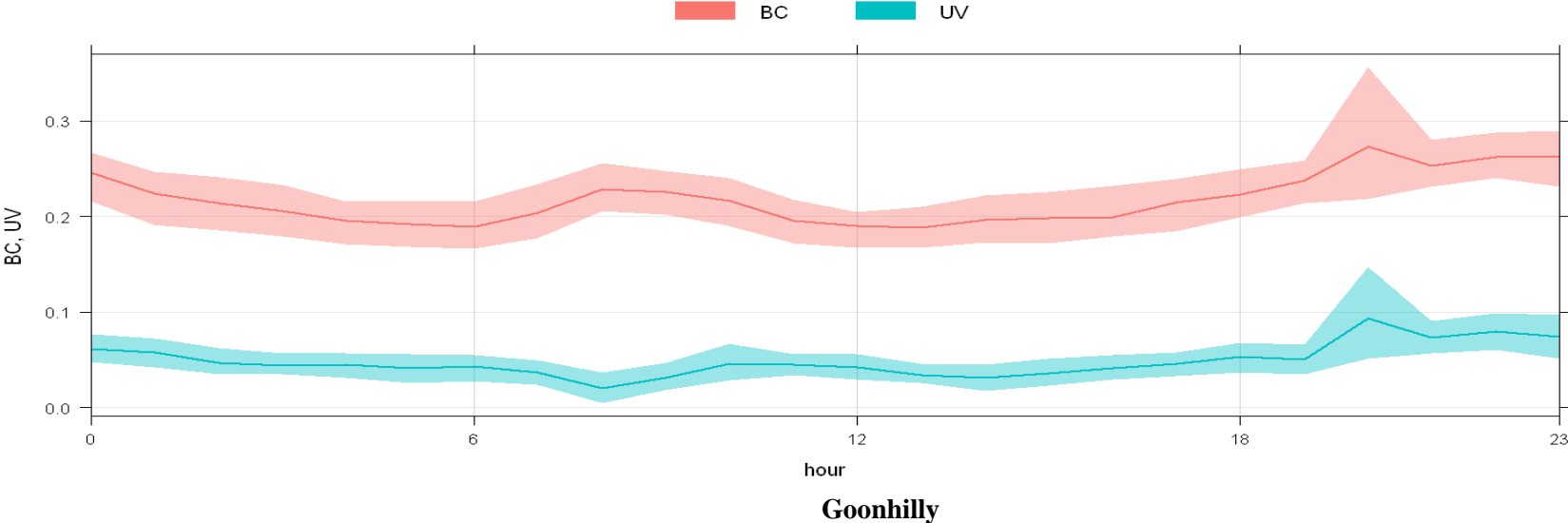
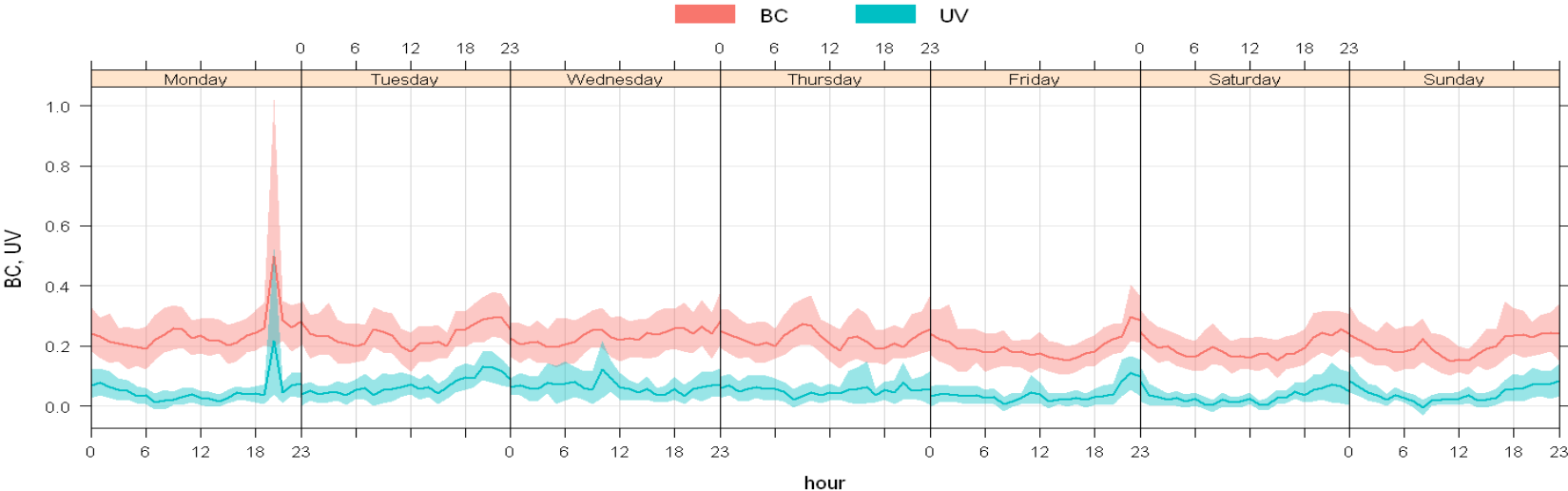


Figure18 2013 Urban Background sites





Harwell



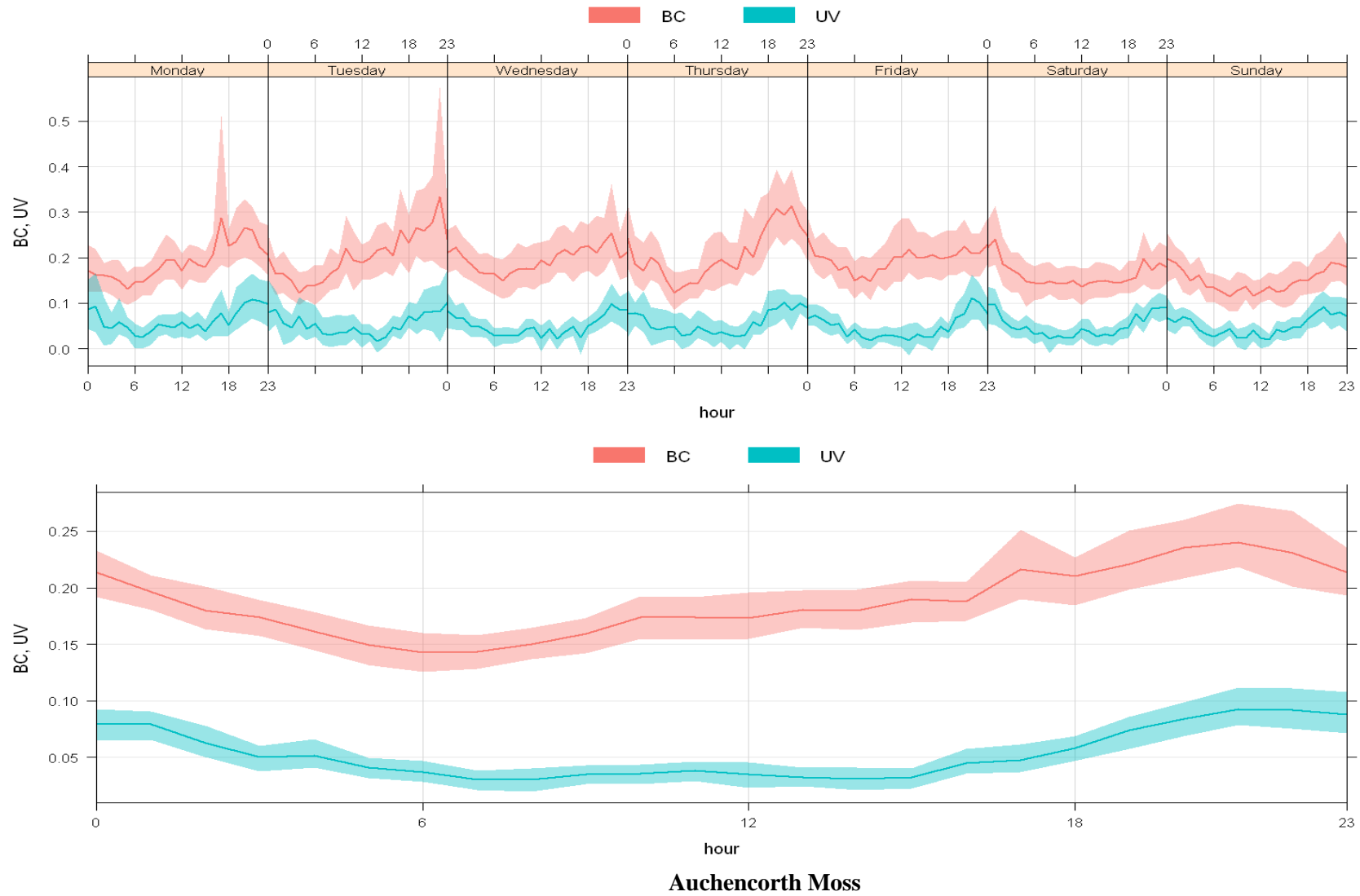


Figure 19 2013 Rural Background sites

2009 – 2013 Data

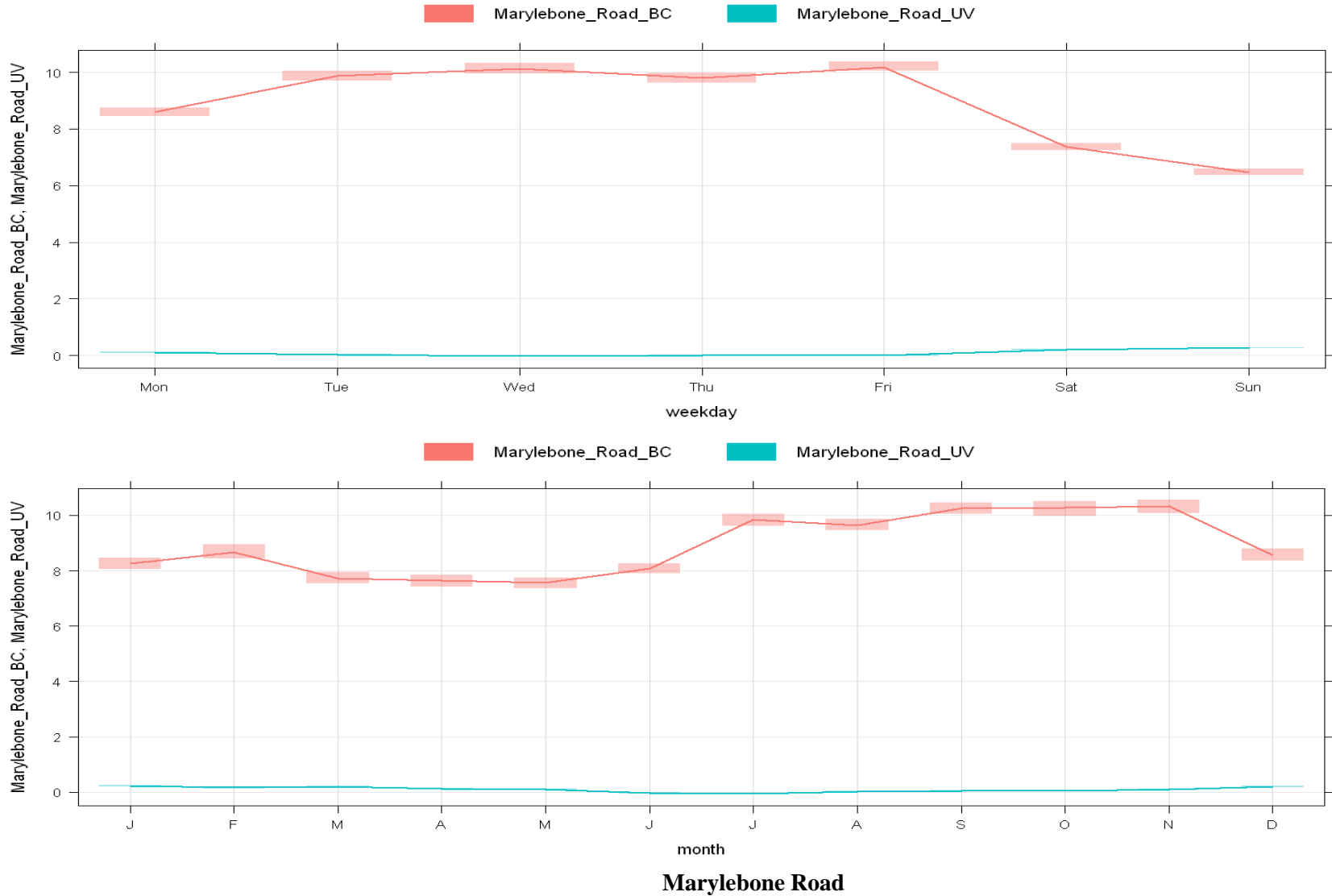


Figure 20 2009 – 2013 Roadside sites

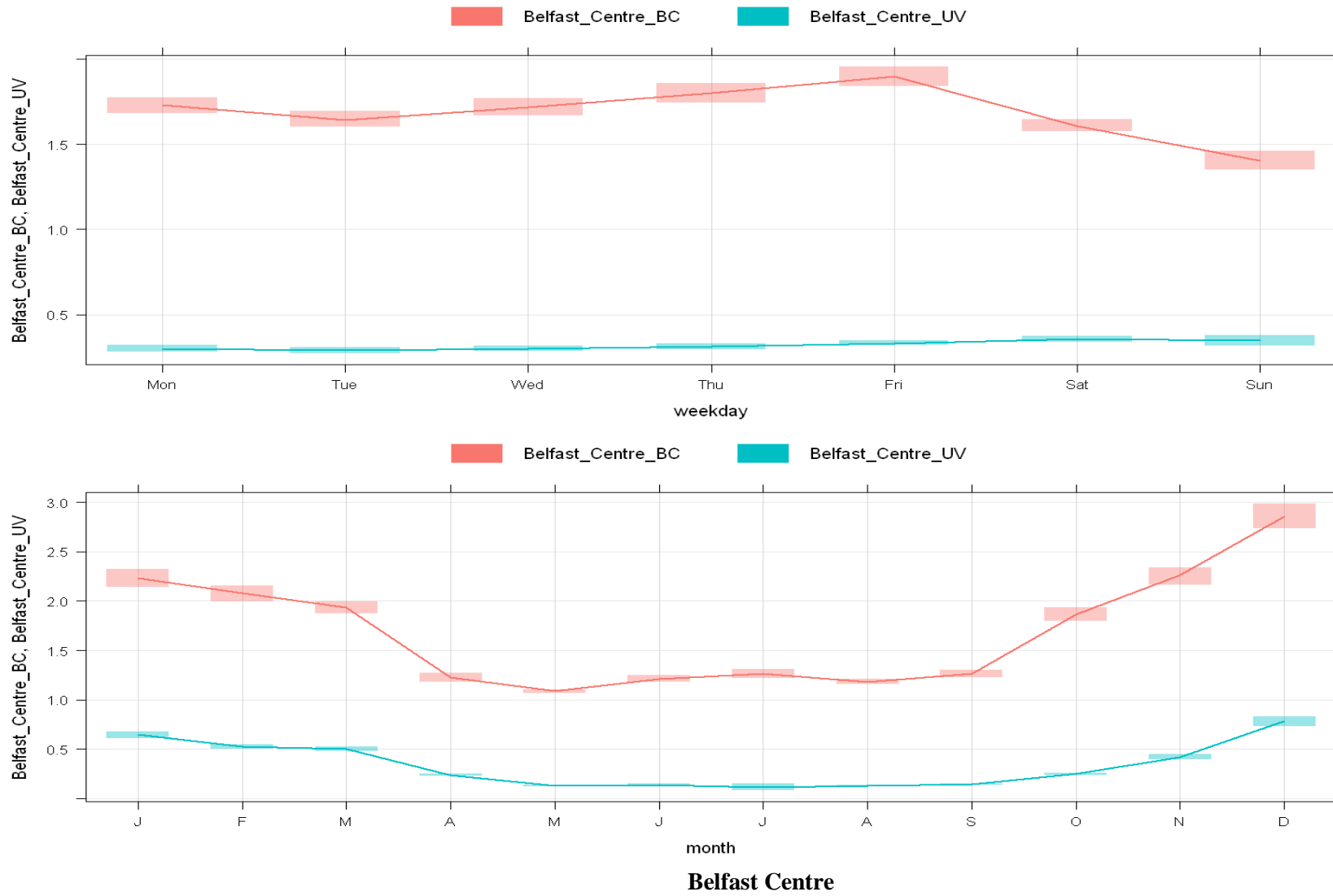
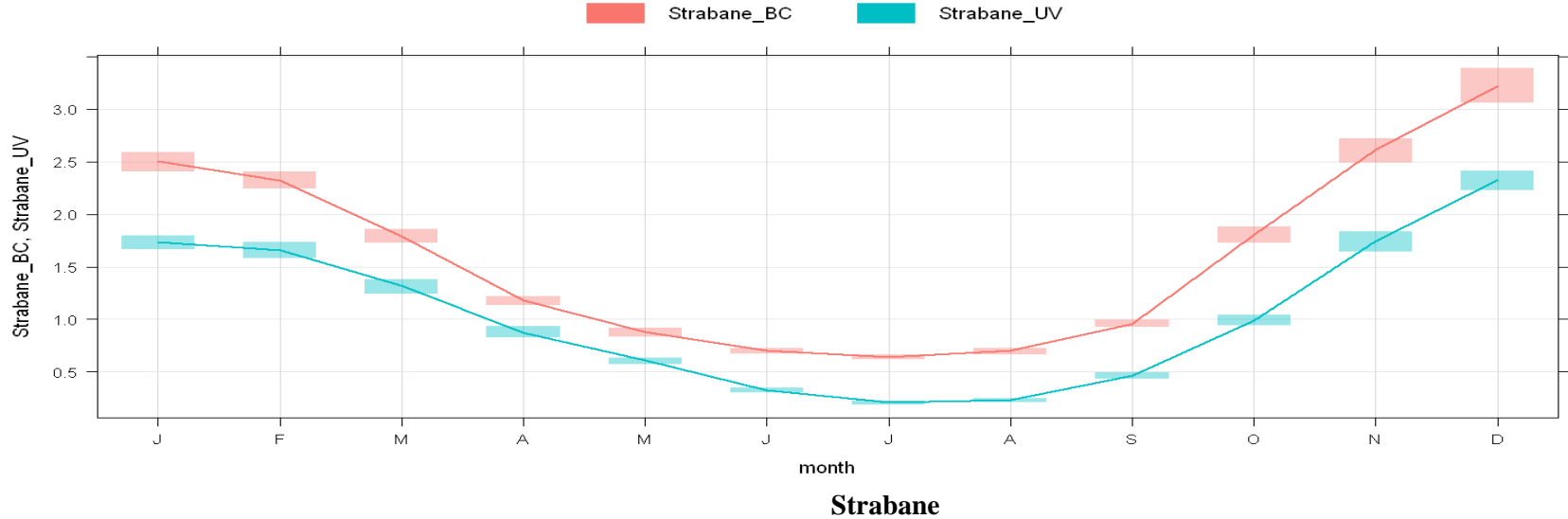
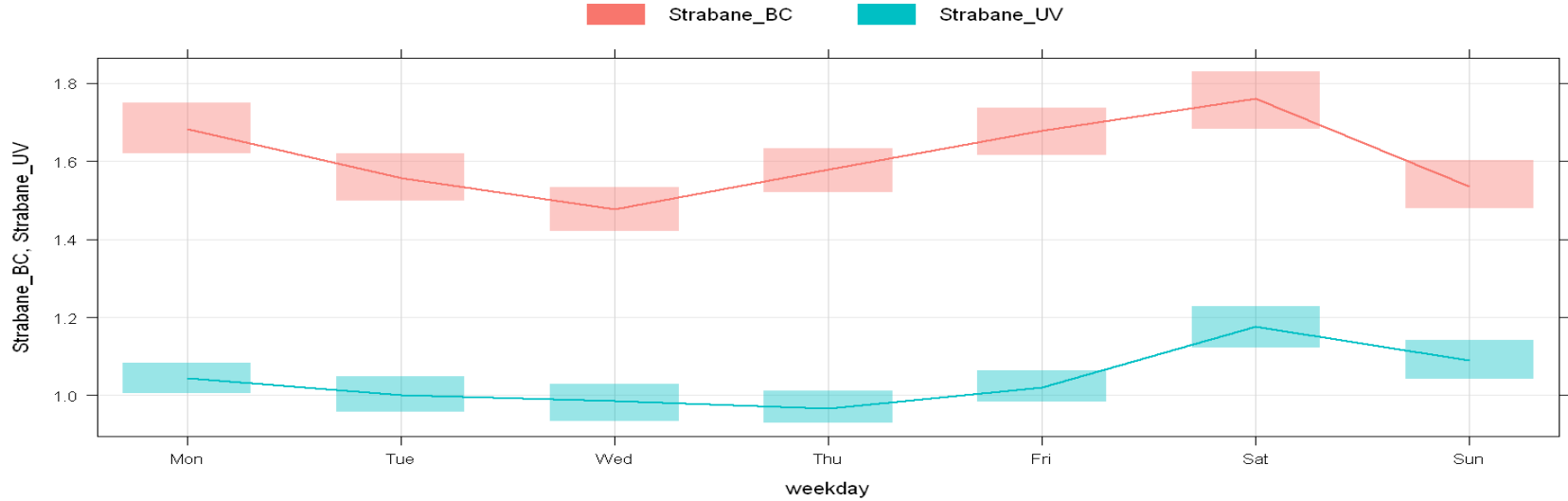
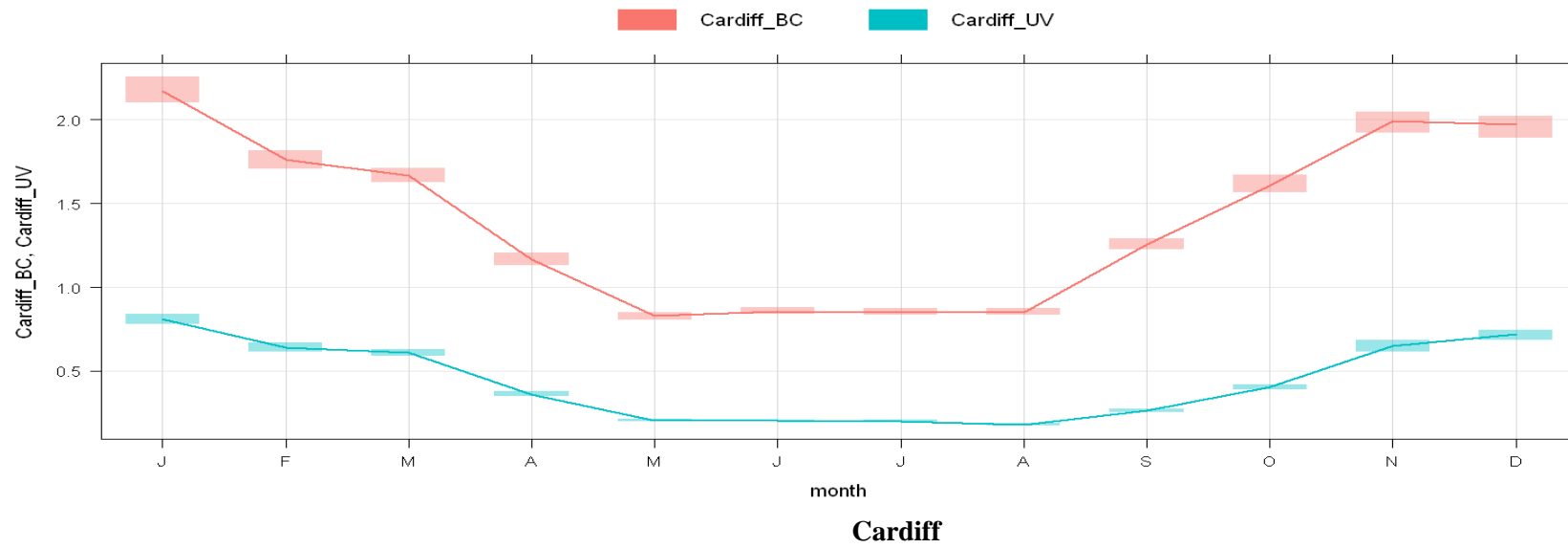
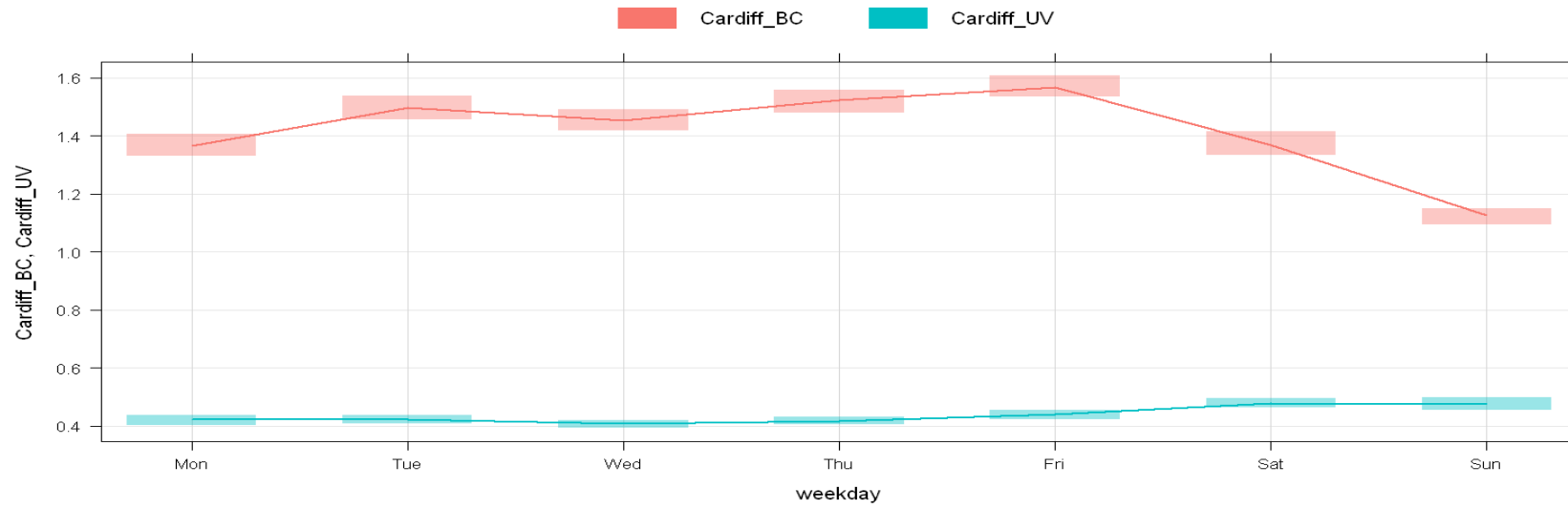
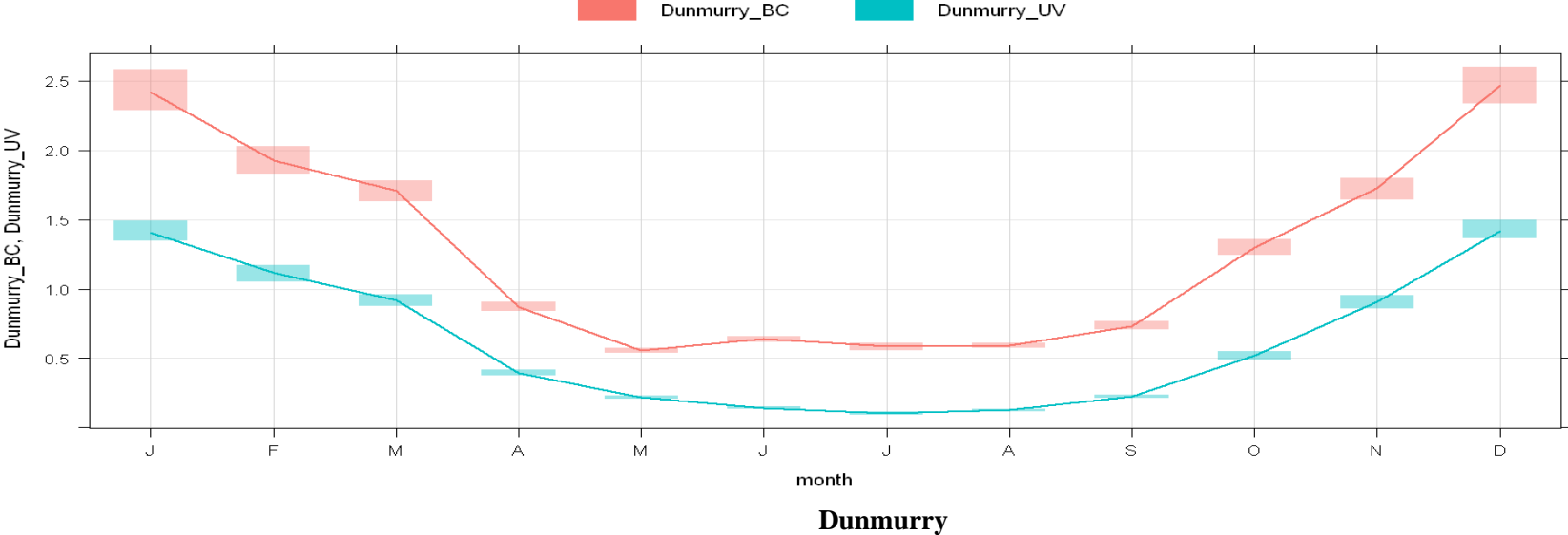
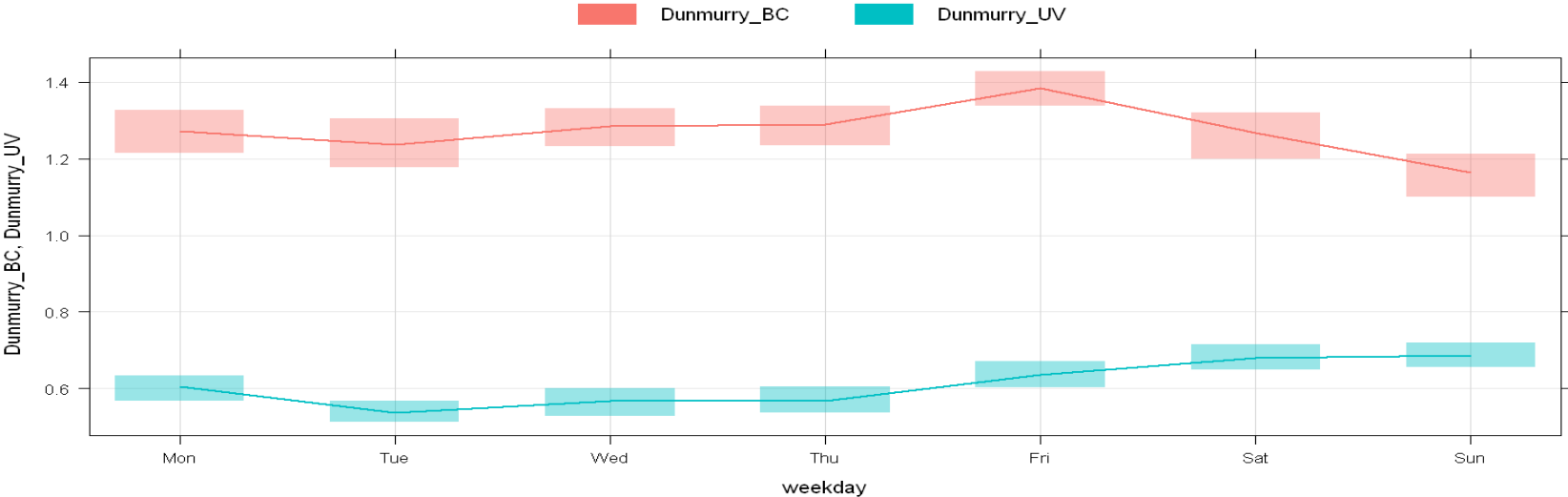
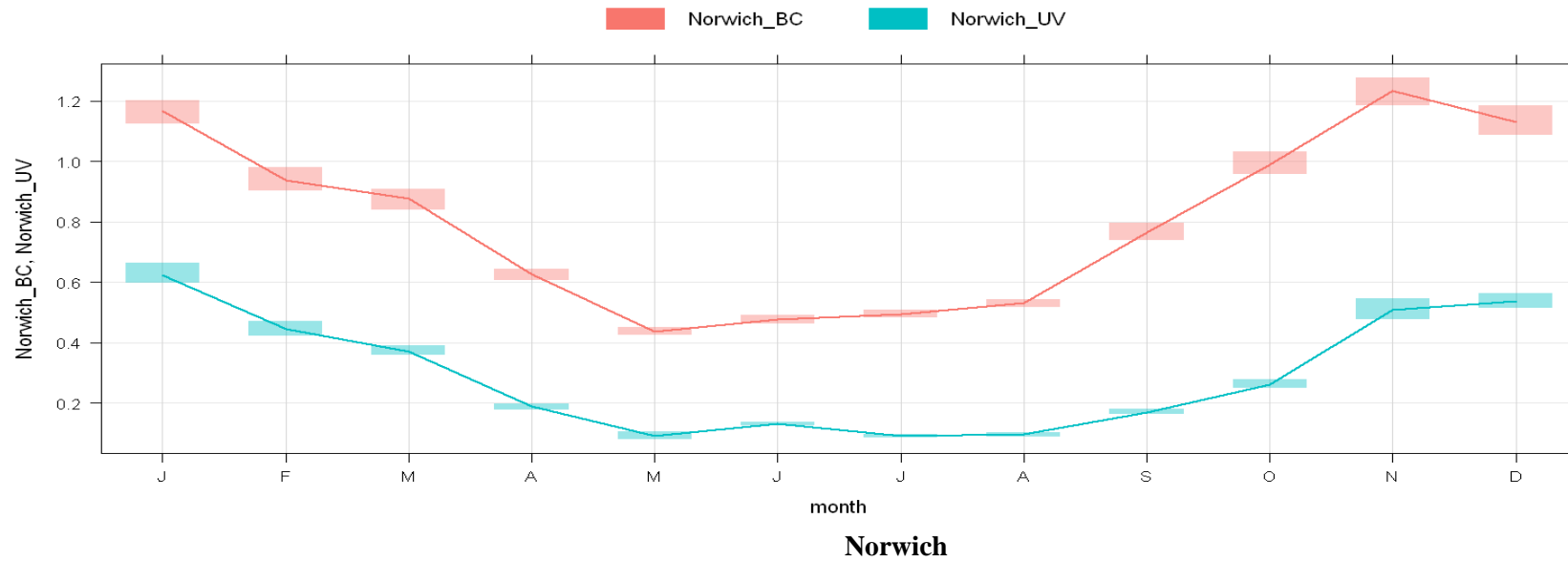
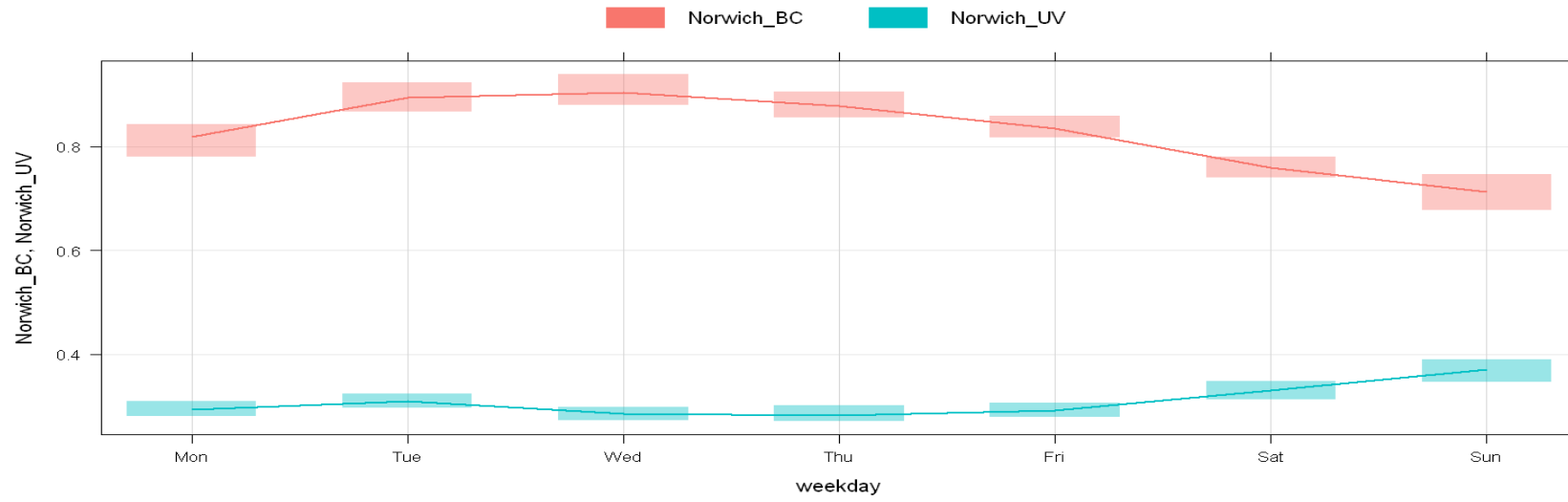


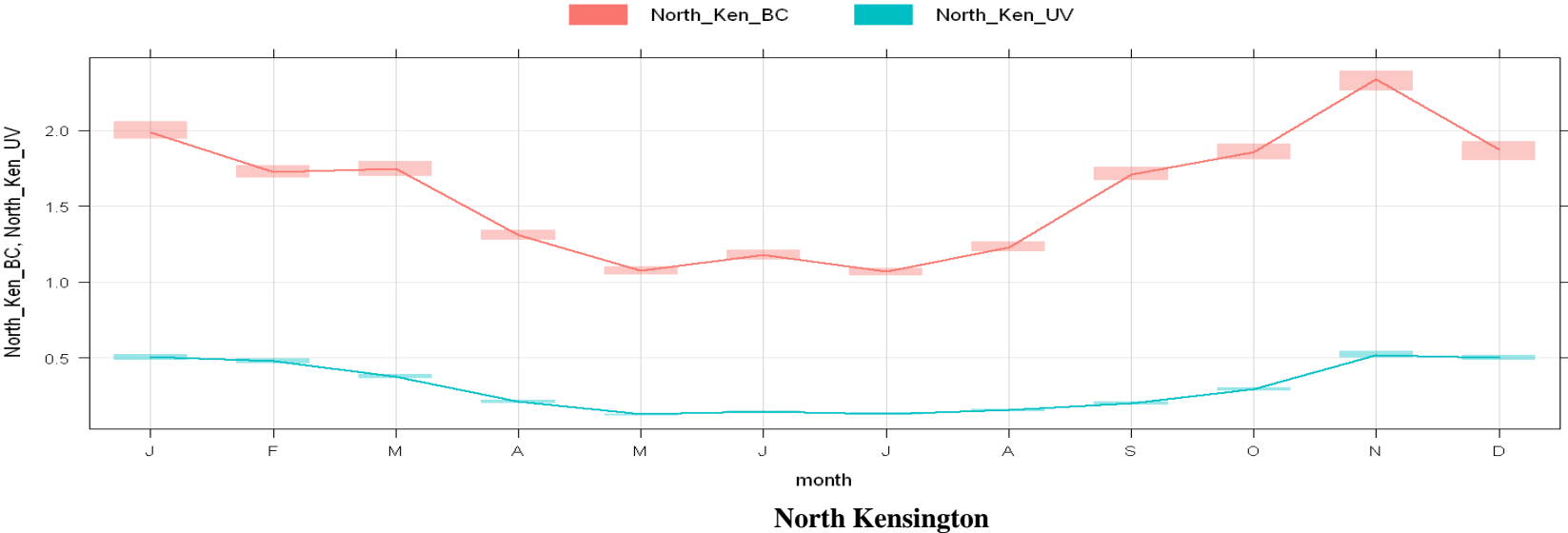
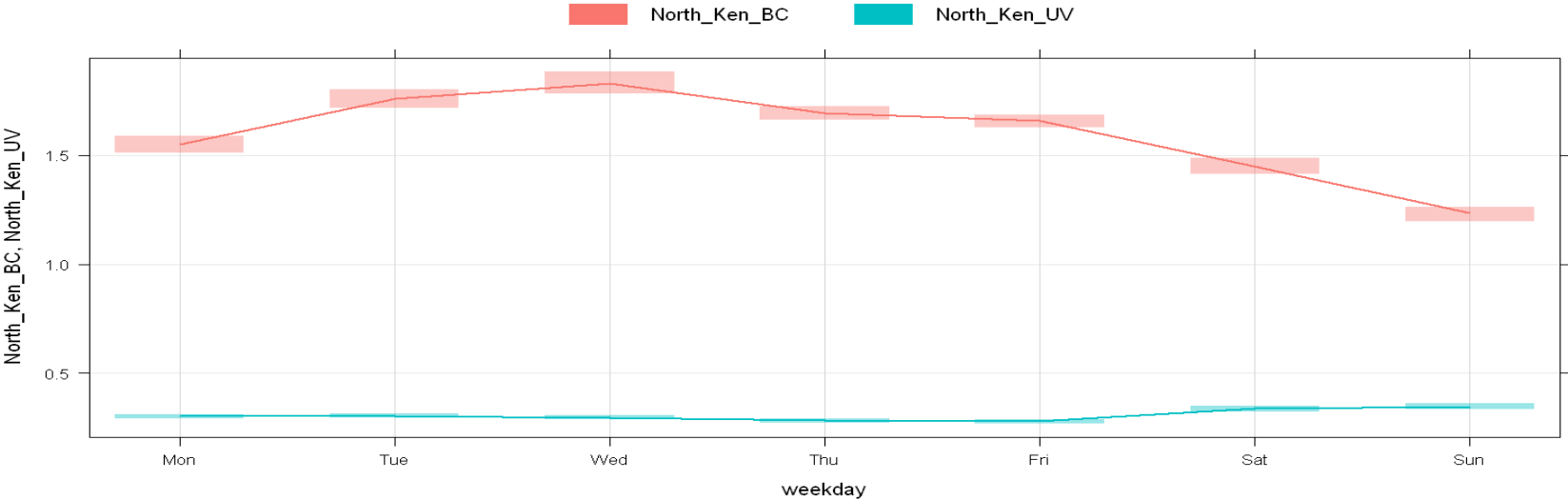
Figure 21 2009 – 2013 Urban Centre sites



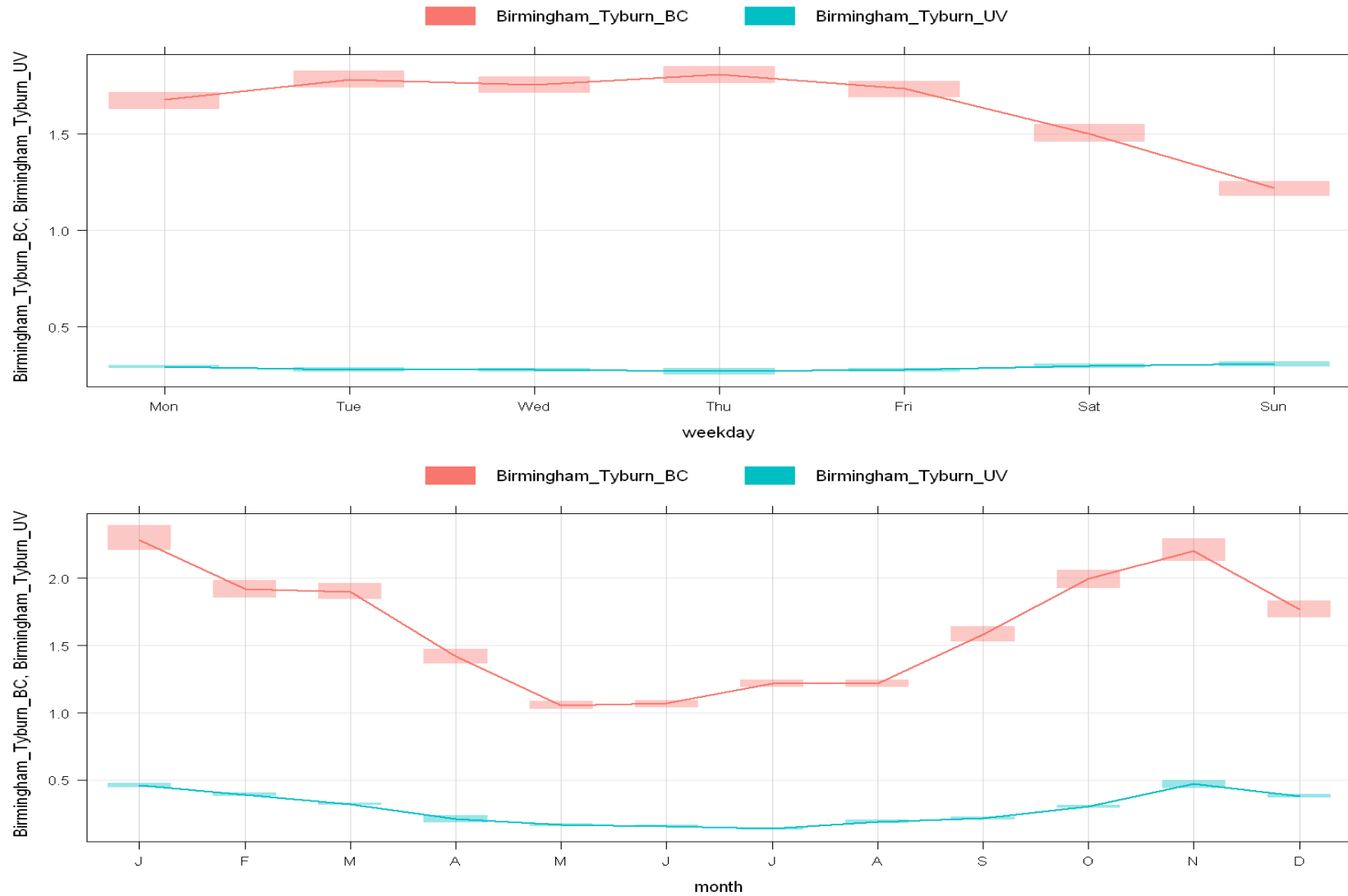








North Kensington



Birmingham Tyburn Background

Figure 22 2009 – 2013 Urban Background sites

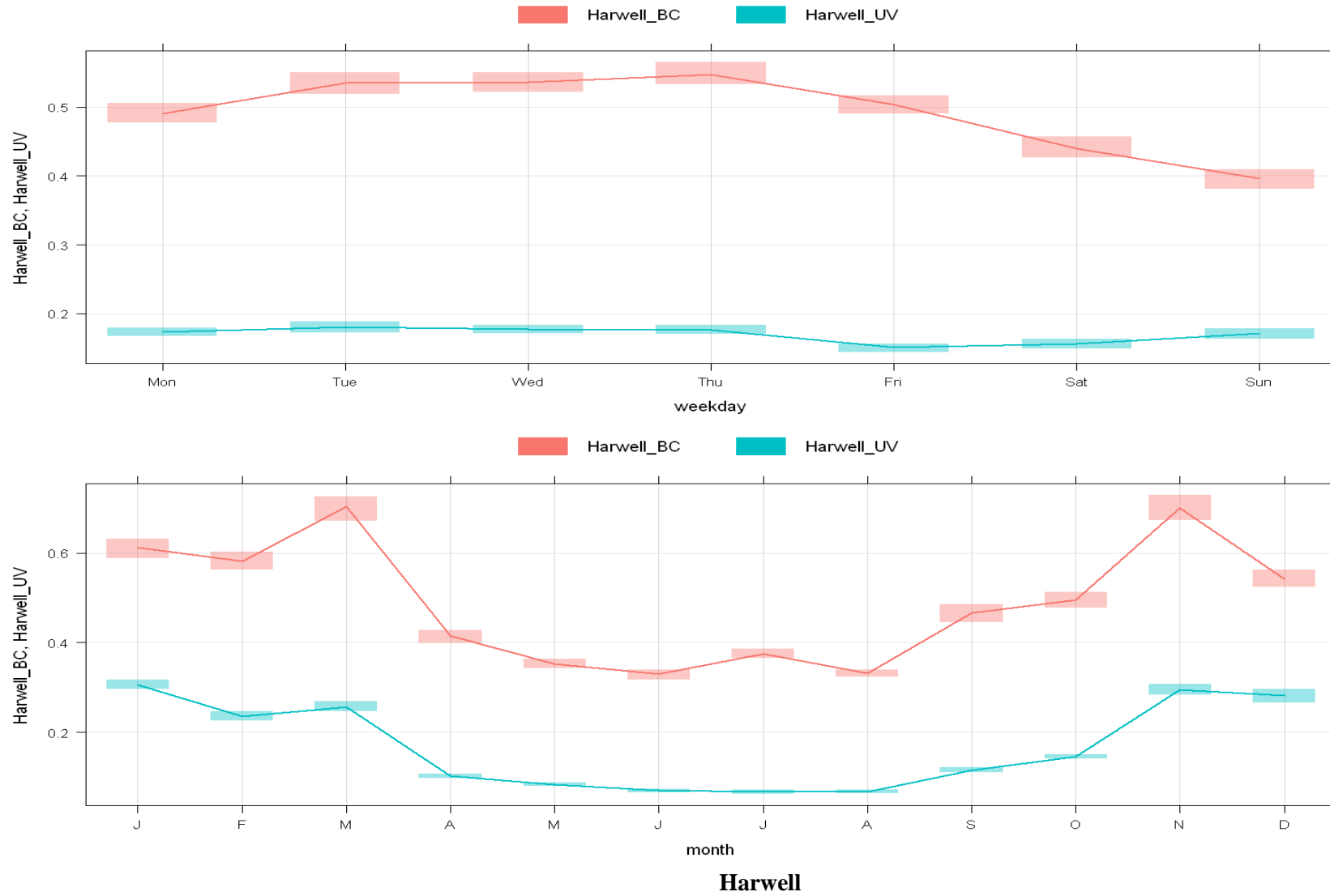


Figure 23 2009 -2013 Rural Background sites

Roadside sites

It can be seen that the Black Carbon concentrations at the roadside sites follow the expected profile for traffic movements through the day, with raised concentrations in the morning and evening rush hours.

Marylebone Road shows a decrease in Black Carbon concentrations at the weekend, when compared with weekdays, which is in line with the reduced traffic and change in vehicle fleet over the weekend. There is a considerable drop in the number of heavy goods vehicles at the Marylebone Road site over the weekend.

Birmingham Tyburn shows more of the traditional morning rush hour peak of Black Carbon in the morning, followed by a drop in concentrations during the day, with a smaller peak for the evening rush hour. Weekend concentrations are also a lot lower and relatively flat during the day. As with Marylebone Road there is little UV component signature, however Birmingham shows a slight increase in the evenings of the week end, indicating possible local solid fuel / wood burning secondary heating.

Marylebone Road shows no significant seasonal trend in either Black Carbon or UV component concentrations due to the consistent traffic flows being the dominant emission source of black carbon.

Urban Centre sites

Belfast centre also shows an increase in Black Carbon concentrations coinciding with the morning rush hour, but with concentrations remaining fairly constant from 17:00 hours until midnight. Black Carbon concentrations also show an increase late on Saturday evening / early Sunday morning, which can be assumed to be due to evening leisure journeys.

Black Carbon concentrations are generally lower at the weekends compared with working days.

Both Black Carbon and UV component concentrations show some seasonality dependence, with a decrease in concentration over the summer months and an increase in concentration in the winter months. If it is assumed that the emissions from road transport are relatively consistent through the year then normalising the concentrations should reveal any monthly variability in either the Black Carbon or UV component emissions. This can be seen in Figures 24 and 25.

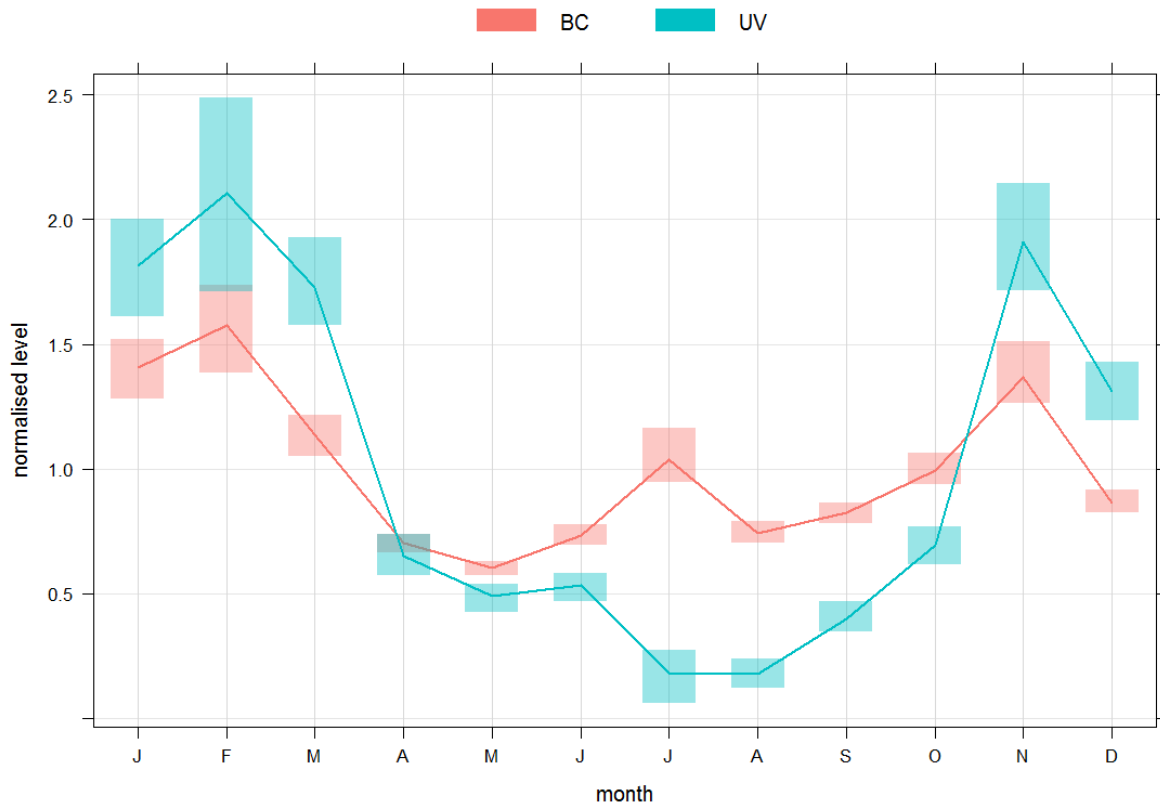


Figure 24 Normalised monthly variability at Belfast for the period 2009 - 2013

The increase in Black Carbon in July is due to localised burning around the period of 12 July.

Urban Background sites

Urban background sites can be split into three categories: those predominantly influenced by emissions from domestic heating, those away from main roads but still predominantly influenced by road traffic sources and those that are a combination of domestic and traffic sources.

Concentrations measured at Strabane, Ballymena and Dunmurry Kilmakee in Northern Ireland are dominated by emissions from domestic heating. UV component and Black Carbon concentrations follow similar hourly, daily and seasonal trends and have similar absolute concentrations. Strabane is not on the natural gas supply and domestic heating mainly comes from oil. Strabane is in a smokeless zone, however there is evidence that this is being ignored in some areas of the town and residents are burning smoky coal. Due to the large emission factors of PAHs from smoky coal compared to oil and gas, it does not take many houses burning this coal to have a big influence on ambient concentrations. At Strabane there is little evidence of traffic emissions during the rush hour periods. Ballymena and Dunmurry are on the natural gas supply and this is the predominant source of domestic heating, however coal is often used as secondary heating in the evenings. Due to the difference in emission factor discussed above this can have a significant effect on ambient concentrations. At Ballymena and Dunmurry the morning rush hour is picked up in the Black Carbon concentrations, but the evening rush hour is masked by domestic emissions. Figure 25 gives the monthly variability and Figure 26 gives the hourly variability

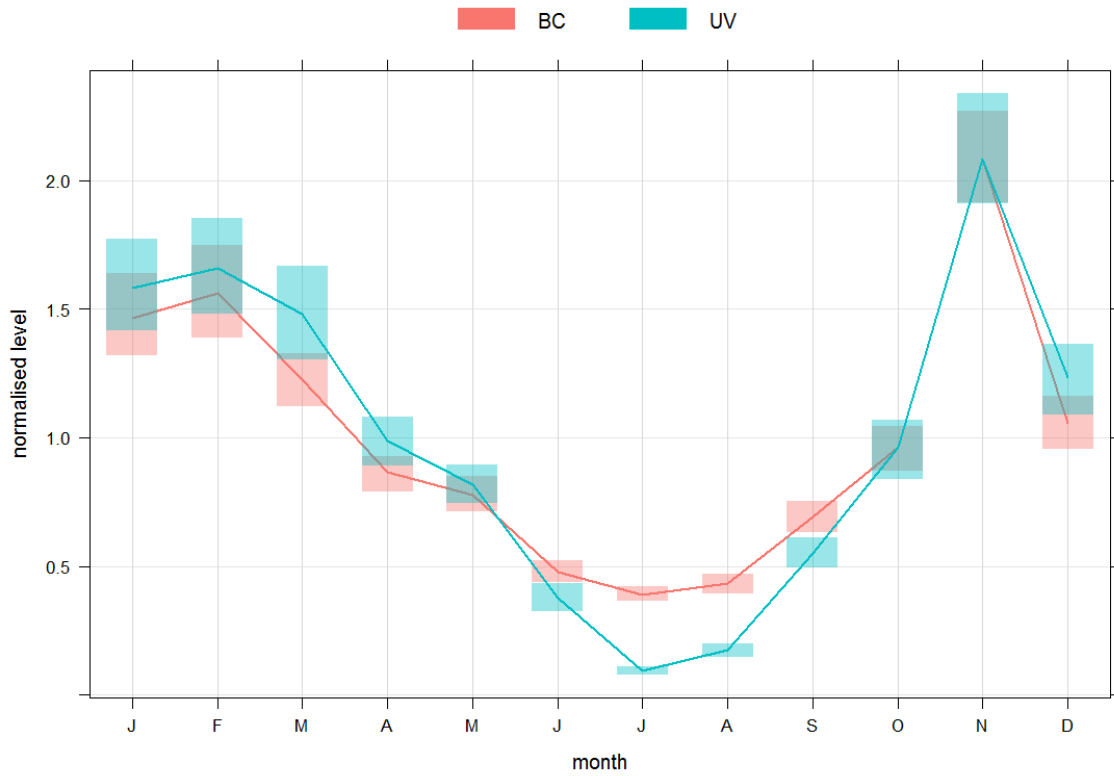


Figure 25 Normalised monthly variability at Strabane for the period 2009 - 2013

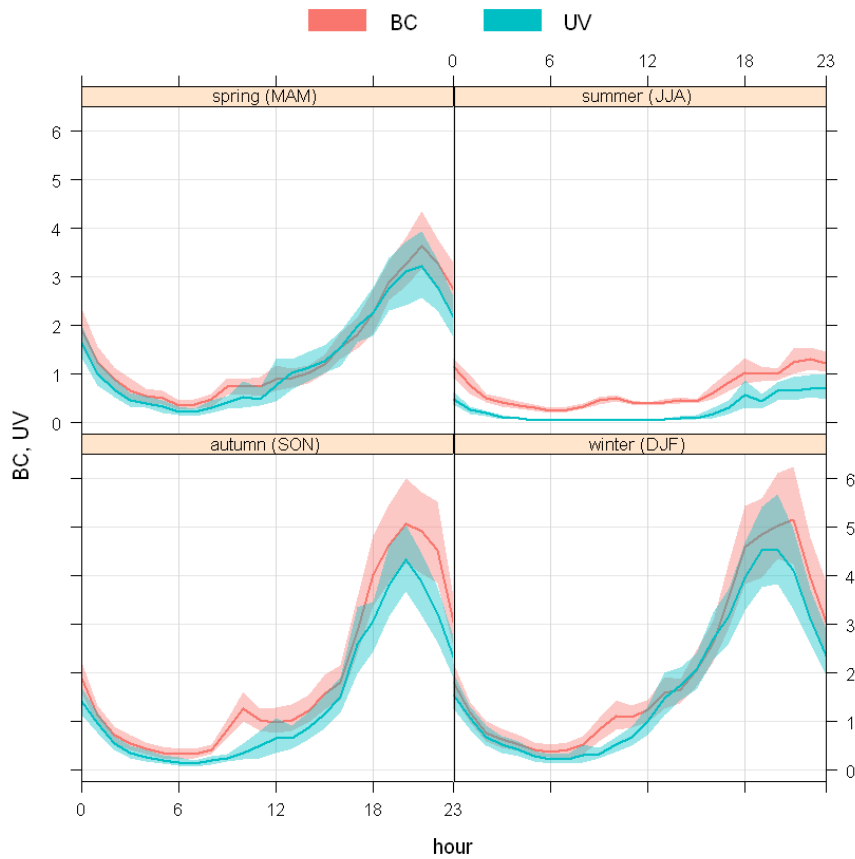


Figure 26 Seasonal Black Carbon and UV component concentrations measured at Strabane in 2013

In the above chart the shaded area on the y-axis represents the uncertainty in the mean y-value due to the spread of results, expressed with a level of confidence of 95%.

Concentrations measured at Birmingham Tyburn and North Kensington are dominated by emissions from road traffic. The morning rush hour can be clearly seen in the Black Carbon concentrations, with the evening rush hour seen to a lesser extent. The peak in evening Black Carbon concentrations continues later into the evening than at roadside sites and could be linked to small scale emissions from domestic heating. This peak in the evening Black Carbon concentrations is also seen as a small peak in the UV component. Both sites have lower Black Carbon concentrations at the weekend and slightly elevated UV component concentrations over the same time scale. Also both sites show a seasonal component in Black Carbon and UV component concentrations. Both these phenomena point to the use of solid fuel (possibly wood burning) in the urban area.

Concentrations at Cardiff and Norwich show responses to both emissions from road sources and domestic heating. The morning rush hour is detected by raised Black Carbon concentrations with little change in the UV component, but from 15:00hrs onward both Black Carbon and UV component concentrations rise quickly. Evening Black Carbon concentrations reach similar levels to morning rush hour while the UV component reaches 2 – 3 times the day time concentration. Also UV component concentrations at Norwich on Saturday and Sunday nights are considerably higher than work days. Both sites show seasonal trends associated with raised concentrations in both measurement channels in line with increased domestic fuel usage / secondary heating using coal or wood in open fires in the colder months.

Rural sites

The rural background site concentrations are lower than the other site classifications, as expected. The effect of local heating can also be seen in the concentrations at the Detling suburban site. All sites display the morning rush hour in the Black Carbon concentration to a small extent.

The spikes in Black Carbon and UV component concentrations seen at Detling on Wednesday and Goonhilly on Monday are due to localised burning episodes. At Goonhilly a large gorse fire on Monday 1st April 2013 required the attendance of the fire brigade.

Over the longer dataset, Harwell shows some seasonality in the UV component concentration associated with domestic emissions, and reduced Black Carbon concentrations at weekends associated with lower local traffic flows.

5.4 COMPARISONS WITH OTHER POLLUTANTS

Comparisons are possible between Elemental Carbon and Black Carbon concentrations at three sites, and between PAH and UV component concentrations at three (different) sites.

Comparisons were also made with particle mass measurements where these instruments were collocated with the Aethalometer.

5.4.1 Elemental Carbon

Daily Elemental Carbon (EC) measurements are made at the North Kensington, Marylebone Road and Harwell sites by the Particle Number and Speciation Network⁵. Aethalometer concentrations (BC) at these sites have been averaged into daily measurements and plotted as scatter plots against the elemental carbon (EC) concentrations in Figures 27 to 29.

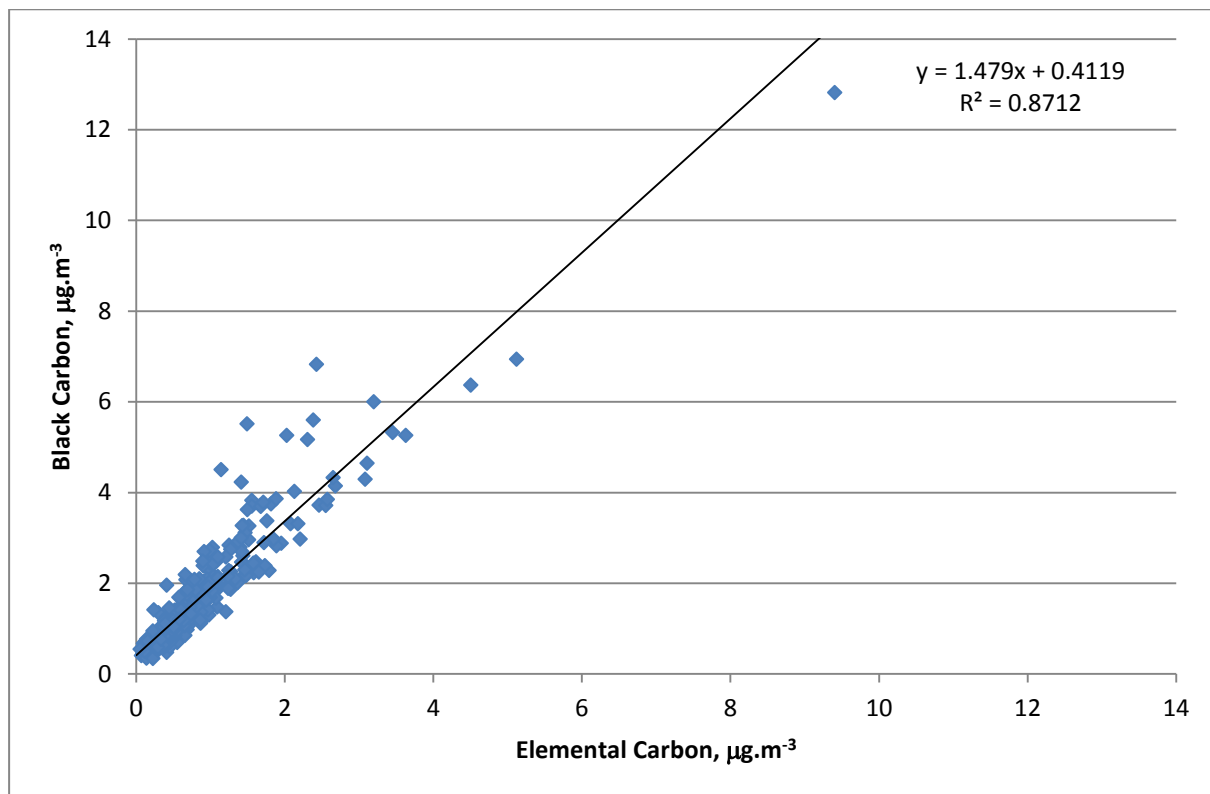


Figure 27 2013 EC and BC Measurements at North Kensington

5 S Beccaceci et al, Draft NPL REPORT, 2013 Annual Report for Airborne Particulate Concentrations and Numbers in the United Kingdom (phase 3), April 2014

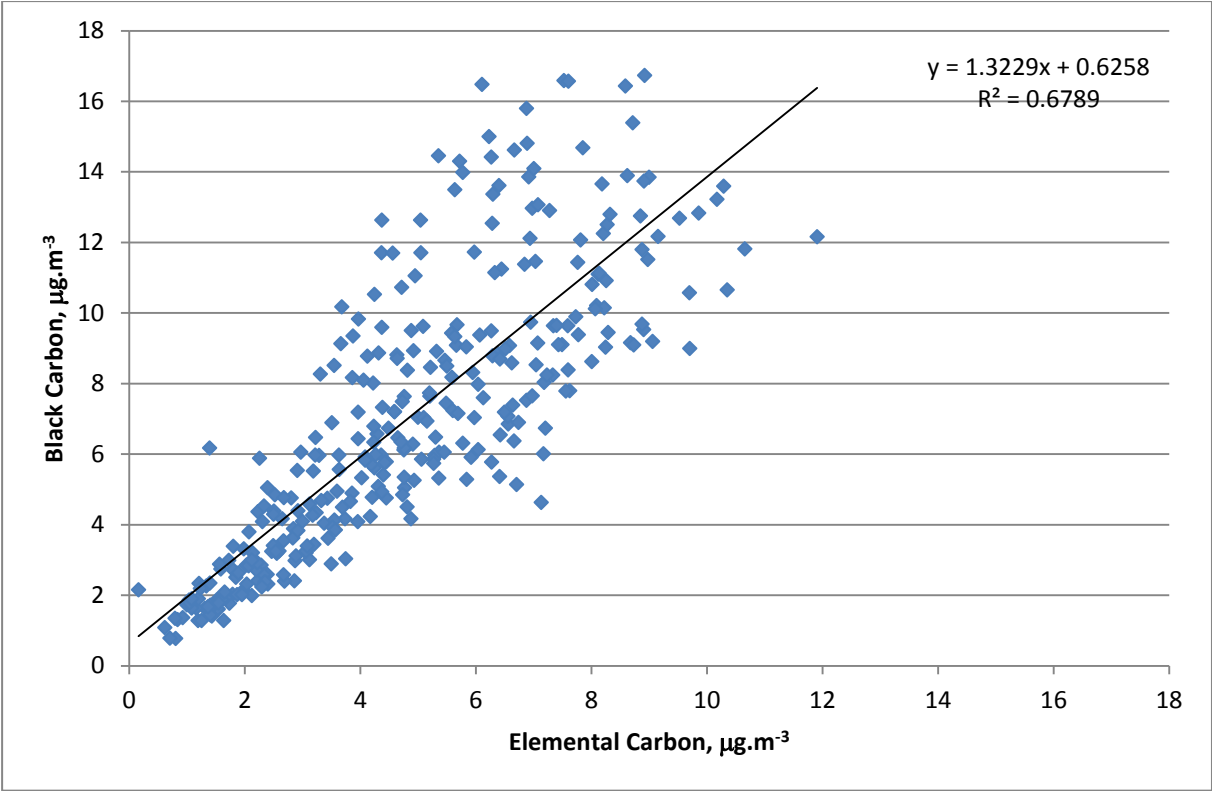


Figure 28 2013 EC and BC Measurements at Marylebone Road

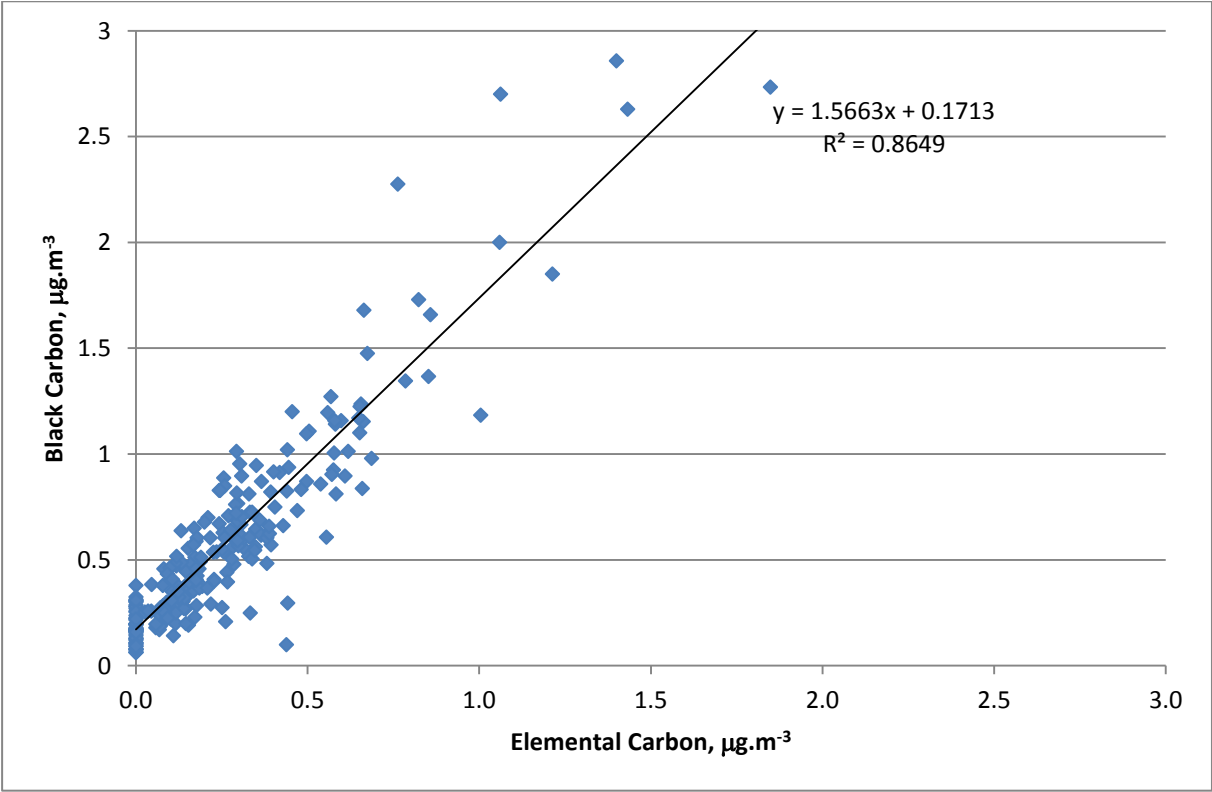


Figure 29 2013 EC and BC Measurements at Harwell

It can be seen that there is a good linear relationship between the Elemental Carbon and Black Carbon concentrations measured at the North Kensington and Harwell sites, while at Marylebone Road there is more scatter. The relationship between Black Carbon and Elemental Carbon has been quite variable year on year. This variability is shown in Table 14.

Year	Harwell		North Kensington		Marylebone Road	
	Relationship	R ²	Relationship	R ²	Relationship	R ²
2009	N/A	N/A	0.97x + 0.33	0.858	1.20x + 0.56	0.776
2010	0.89x + 0.20	0.555	1.16x + 0.03	0.734	1.26x + 0.40	0.946
2011	1.42x + 0.16	0.844	1.16x + 0.14	0.810	1.36x + 0.95	0.924
2012	1.69x + 0.11	0.908	1.32x + 0.30	0.906	1.33x + 0.70	0.898
2013	1.57x + 0.17	0.865	1.48x + 0.41	0.871	1.32x + 0.63	0.679

Note There is not enough BC data collected at Harwell to form a reliable relationship as the Aethalometer was only installed in November 2009.

Table 14 Relationships between Black Carbon and Elemental Carbon over the period 2009 – 2013

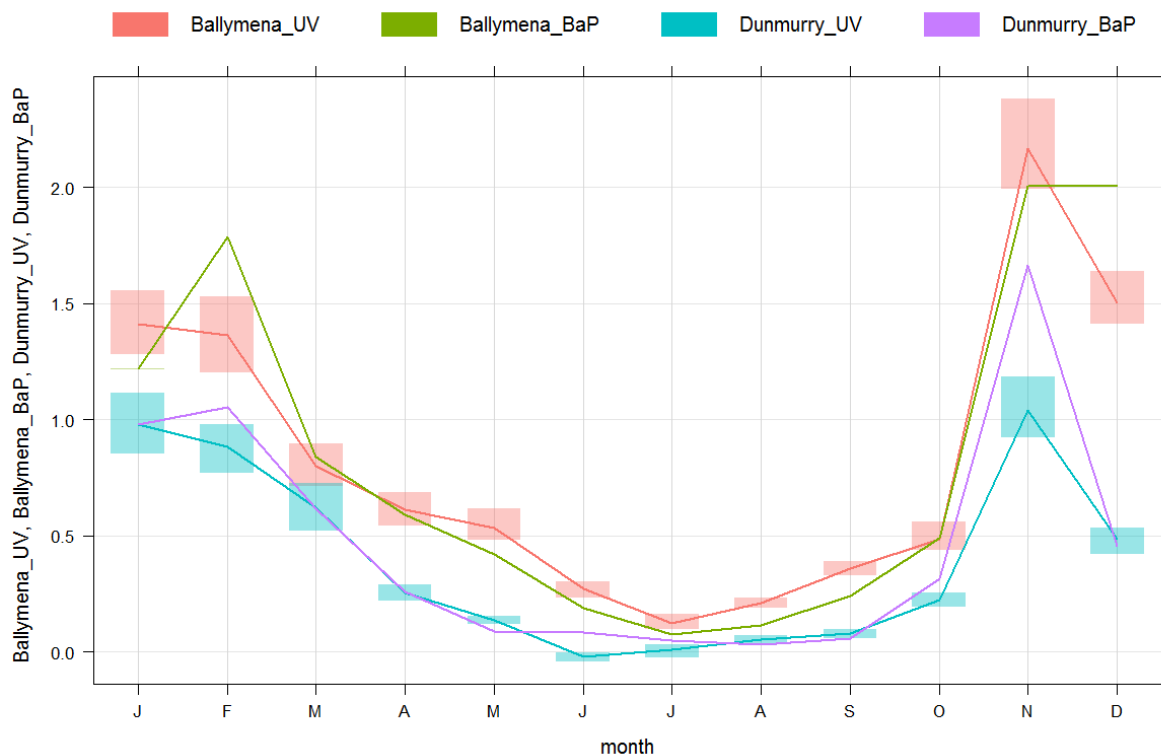
It can be seen that the slopes, i.e. the amounts of Black Carbon relative to Elemental Carbon, have generally increased year on year at all sites, with the effect being larger at North Kensington and Harwell, where the slopes have increased from about 1 to about 1.5, than at Marylebone Road, where the slopes have increased from about 1.2 to 1.35.

There are three possible general explanations for this. (1) The method used for Black Carbon measurement has changed such that reported concentrations are now relatively higher than they used to be; (2) the method used for Elemental Carbon measurement has changed such that reported concentrations are now relatively lower than they used to be; and (3) the nature of the soot-like pollution has changed in a way that causes larger differences between the methods. As the methods have all been under the control of NPL over this period, and no significant changes to them have been made, the first two explanations seem less likely than the third.

5.4.2 Polycyclic Aromatic Hydrocarbons (PAH)

Monthly concentrations of benzo[a]pyrene are measured at Dunmurry Kilmakee, Ballymena and Birmingham Tyburn Background under the UK PAH Network⁶. The concentrations at Birmingham are much lower than at the other two sites, and are not presented here graphically. Aethalometer concentrations (UV component) have been averaged into monthly measurements and plotted as a time series with the benzo[a]pyrene concentration for Dunmurry and Ballymena in Figure 30.

⁶ Reference for 2013 Annual Report for UK PAH Network



Note: The units are different for the two quantities, the PAH measurements are represented in ng.m^{-3} while the UV component is represented as $\mu\text{g.m}^{-3}$, however it can be seen that there is good temporal agreement between the two species due to similar emission sources such as solid fuel burning.

Figure 30 UV component and Benzo[a]pyrene concentrations measured at Ballymena and Dunmurry in 2013

The shaded area on the y-axis represents the uncertainty in the mean y-value due to the spread of the results over that averaging period, expressed with a level of confidence of 95%. It is not the overall measurement uncertainty. As the PAH measurements are a monthly average there is no spread in the result over the month and therefore no uncertainty displayed in the y-value. The shaded area on the x-axis is for display purposes only, to allow the uncertainty in the mean value to be seen more clearly.

Looking at data over the full period of Network operation, the move of the site from Dunmurry High School to Dunmurry Kilmakee has had an effect on the concentrations measured. Concentrations over the period 2009 – 2013 are shown in Table 15.

Year	BaP, ng.m^{-3}		UV component, $\mu\text{g.m}^{-3}$	
	Dunmurry High School	Dunmurry Kilmakee	Dunmurry High School	Dunmurry Kilmakee
2009	0.83		0.7	
2010	1.44		0.9	
2011	0.86		0.6	
2012	0.58	0.49	0.6	0.5
2013		0.46		0.4

Table 15 Annual Mean BaP and UV component concentrations measured in Dunmurry over the period 2009 – 2013

The reduction in concentration of both pollutants may be due to different topography between the two sites or due to differences in local fuel use.

Table 16 compares the ratio of the BaP and UV component concentrations at Dunmurry, Ballymena and Birmingham Tyburn Background sites for the period 2011 to 2013.

Year	Dunmurry			Ballymena			Birmingham Tyburn Background		
	UV Component, $\mu\text{g.m}^{-3}$	BaP, ng.m^{-3}	BaP / UV	UV Component, $\mu\text{g.m}^{-3}$	BaP, ng.m^{-3}	BaP / UV	UV Component, $\mu\text{g.m}^{-3}$	BaP, ng.m^{-3}	BaP / UV
2011	0.60	0.86	1.438	N/A	N/A		0.30	0.199	0.663
2012	0.55	0.53	0.968	2.10	2.00	0.952	0.30	0.254	0.847
2013	0.40	0.46	1.15	0.80	0.82	1.025	0.20	0.200	1.000

Table 16 UV component and BaP concentrations at Dunmurry, Ballymena and Birmingham Tyburn Background sites for the period 2011 – 2013.

The mean ratio of BaP/UV for the three sites is 1.004. This ratio can then be used at Strabane, where the Aethalometer shows a raised UV component concentration, to give an indication of the probable BaP concentration. Table 17 gives the predicted BaP concentrations at Strabane for the last 5 years.

Year	UV component concentration	Predicted BaP concentration ng.m^{-3}
2009	0.9	0.9
2010	1.3	1.3
2011	0.8	0.8
2012	1.1	1.1
2013	1.2	1.2

Table 17 Predicted BaP concentrations from UV component concentration at Strabane for the period 2009 to 2013.

For 3 out of the 5 five years the predicted BaP concentration is above the 1.0 ng.m^{-3} target value in the EC Directive 2004/107/EC⁷ relating to ambient BaP concentrations. The average concentration over the last 5 years is 1.06 ng.m^{-3} .

A detailed analysis of PAH concentrations in Northern Ireland is contained in an NPL Report for the Department for the Environment Northern Ireland⁸.

7 DIRECTIVE 2004/107/EC OF THE EUROPEAN PARLIAMENT AND OF THE COUNCIL, relating to arsenic, cadmium, mercury, nickel and polycyclic aromatic hydrocarbons in ambient air, 15 December 2004.

8 D Butterfield, R Brown, NPL REPORT AS66, Polycyclic Aromatic Hydrocarbons in Northern Ireland, February 2012.

5.4.3 Particulate Mass

The annual average particulate mass concentration was compared with the Black Carbon concentration at collocated sites where automatic particulate mass instrumentation was installed. Two different types of instruments provide particulate mass concentrations across the Black Carbon Network: TEOM FDMS Model CB drier, and MetOne BAM – reference equivalent. As different automatic methods provide different results, the concentrations reported by these automatic instruments have been separated into instrument type and the results shown in Table 18.

Site	BC $\mu\text{g.m}^{-3}$	TEOM FDMS PM ₁₀ $\mu\text{g.m}^{-3}$	TEOM FDMS PM _{2.5} $\mu\text{g.m}^{-3}$	MetOne BAM PM ₁₀ $\mu\text{g.m}^{-3}$	Percent BC PM ₁₀ %	Percent BC PM _{2.5} %
Marylebone Road	6.9	29.2	20.1		24	34
Birmingham Tyburn Roadside	3.1	19.4	16.0		16	19
Birmingham Tyburn Background	1.4	17.9	14.2		8	10
North Kensington	1.7	23.1*	14.7		7*	12
Norwich	0.7	19.1*	19.5		4*	4
Belfast Centre	1.4	17.7*	11.6		8*	12
Harwell	0.5	16.4*	12.7*		3*	4*
Auchencorth Moss	0.2	10.2	4.9		2	4
Detling	0.7	20.7			3	
Strabane	1.5			23.6	6	

Note: * indicates averages where the data capture is below the required 75% required for a valid average.

MetOne data is taken from the Air Quality Northern Ireland web site where it is stated to be Reference Equivalent.

Grey shaded cells indicate no measurements were made.

Table 18 Comparison of Annual Black Carbon and Particulate Mass Concentrations

The particulate mass measurements made by the bottom two sites are made by Local Authorities and may not have the same QA/QC procedures applied to the data as that applied to the AURN sites.

The fact that the PM_{2.5} mass concentration at Norwich is larger than the PM₁₀ mass concentration, is because the data capture for the PM₁₀ instrument is not good enough for a valid average to be calculated.

It can be seen that the PM₁₀ and PM_{2.5} mass concentration measured at Marylebone Road and Birmingham Tyburn roadside sites have a much higher percentage of Black Carbon than the other sites. Black Carbon represents a significant proportion of the total particulate mass at sites influenced by road traffic emissions. Any significant reduction in Black Carbon emissions from road traffic will lead to a measurable reduction in PM₁₀ and PM_{2.5} mass concentrations.

The same comparison cannot be done for the UV component as this is not a real physical or chemical material, but a parameter based on UV absorption due to the mix of organic compounds measured at this wavelength.

5.5 TRENDS

Short and long-term trends in Black Carbon and UV component concentrations are given below.

5.5.1 Short-Term Trends by Site

Figures 31 to 42 show the trend in Black Carbon and UV component concentrations as monthly averages over the full calendar years 2009 to 2013. The Theil-Sen method in OpenAir^{4,5} was used to calculate the regression parameters including slope and uncertainty in the slope.

The Theil-Sen method chooses the median slope among all lines through pairs of two-dimensional sample points. The Theil-Sen estimator tends to yield accurate confidence intervals even with non-normal data and heteroscedasticity (non-constant error variance). It is also resistant to outliers.

Bootstrap resampling provides the confidence interval for the regression slope. For these analyses the 2.5th and 97.5th percentile slopes are taken from all possible slopes.

5.5.1.1 Black Carbon

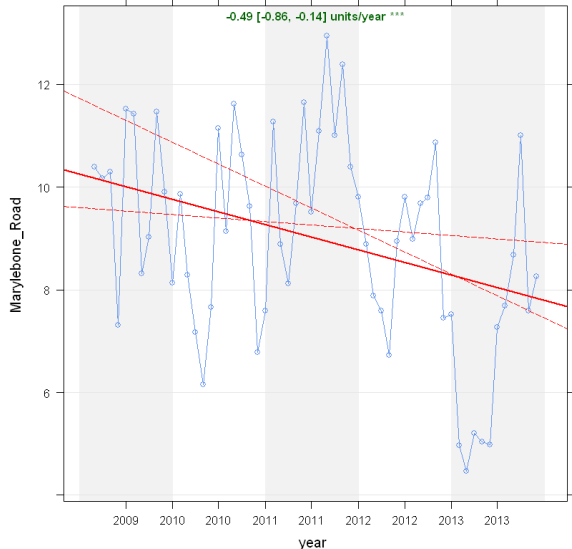


Figure 31 Black Carbon concentrations measured at roadside sites, 2009 – 2013

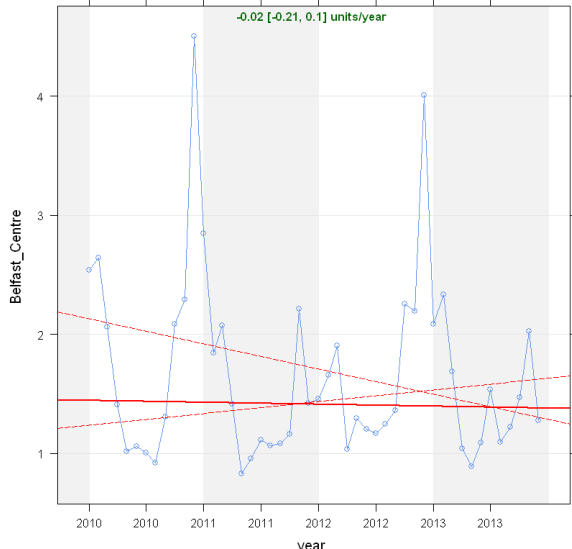


Figure 32 Black Carbon concentrations measured at urban centre sites, 2009 – 2013

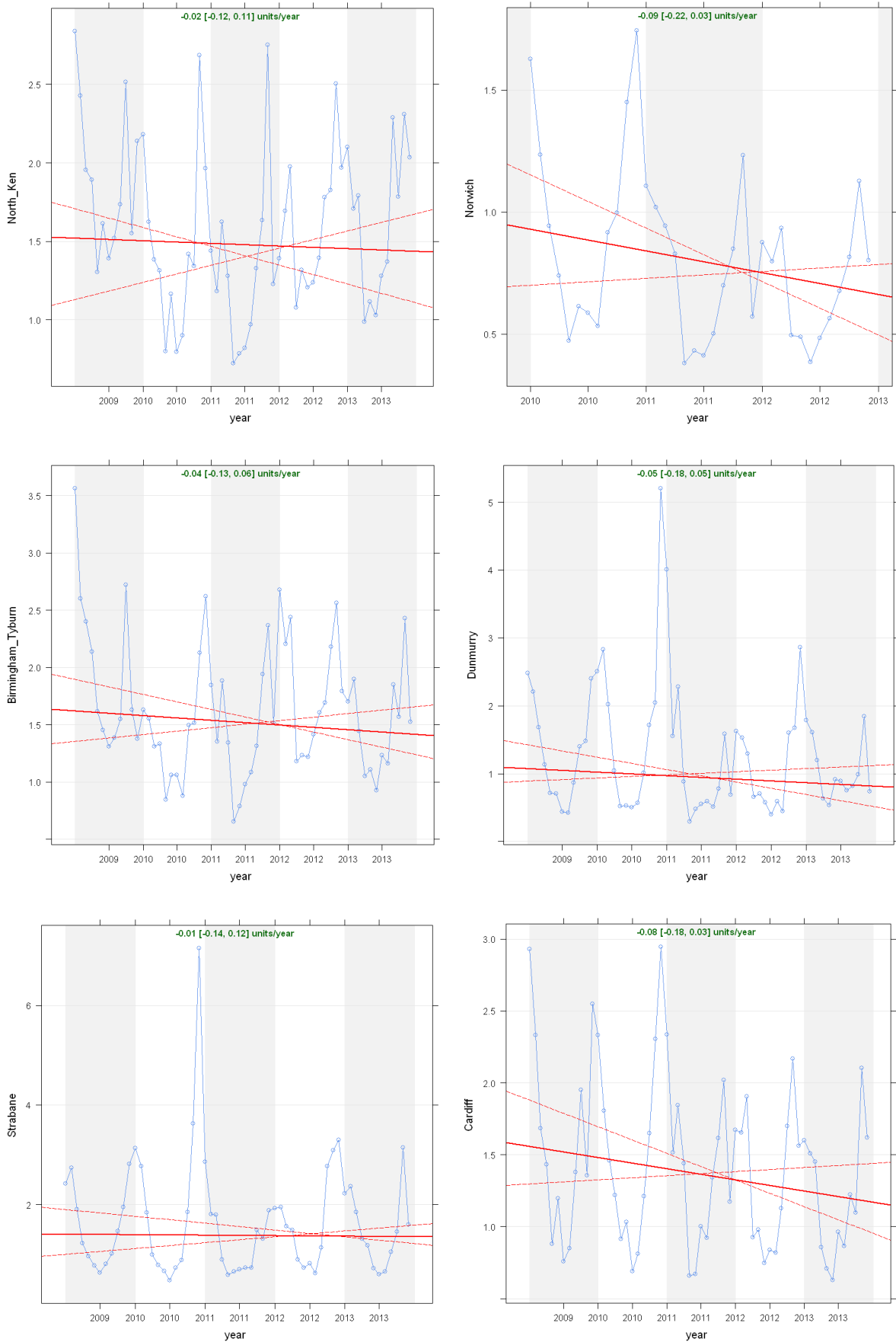


Figure 33 Black Carbon concentrations measured at urban background sites, 2009 – 2013

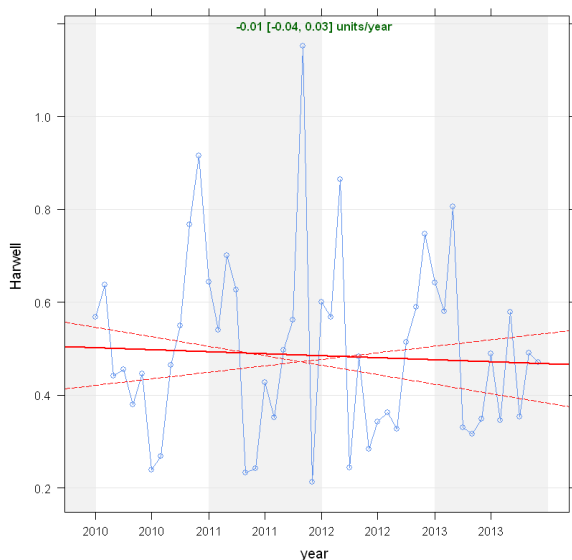


Figure 34 Black Carbon concentrations measured at rural background sites, 2009 – 2013

Site	Slope	Lower limit	Upper limit	Slope significant
Roadside				
Marylebone Road	-0.49	-0.86	-0.14	Y
Urban Centre				
Belfast Centre	-0.02	-0.21	0.10	N
Urban Background				
North Kensington	-0.02	-0.12	0.11	N
Norwich	-0.09	-0.22	0.03	N
Birmingham Tyburn	-0.04	-0.13	0.06	N
Dunmurry	-0.05	-0.18	0.05	N
Strabane	-0.01	-0.14	0.12	N
Cardiff	-0.08	-0.18	0.03	N
Rural				
Harwell	-0.01	-0.04	0.03	N

Table 19 Summary of Black Carbon trends

Over the period 2009 - 2013 the only site that has a significant slope in Black Carbon concentration is Marylebone Road. This should be treated with caution as it may be unduly influenced by the lower than normal concentrations in February to June 2013. This early 2013 drop in concentration is real, as it is also seen in the oxides of nitrogen (NO_x) concentrations (Figure 35). NO_x concentrations are a good tracer for changes in the local primary emissions. Not all PM_{2.5} is due to local primary emissions whereas NO_x is more dominant

Figure 35 shows the 2009-2013 time series for Marylebone Road, with data averaged into monthly values to remove daily fluctuations in concentration. The peak in the March 2013 PM_{2.5} concentration could be due to secondary aerosol formation during a long-range transport episode. These particles will contain little black carbon compared to their mass.

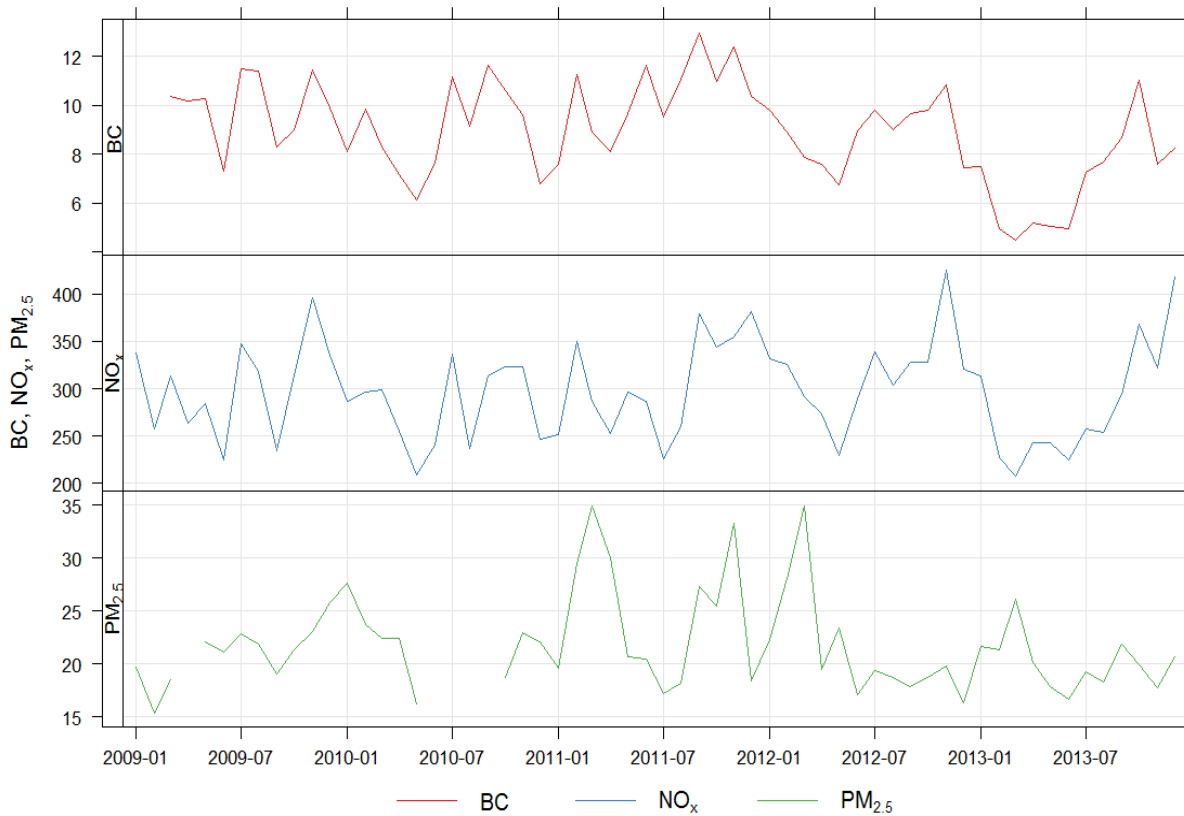


Figure 35 Monthly averages of Black Carbon, PM_{2.5} and Oxides of Nitrogen (NO_x) at Marylebone Road over the period 2009 - 2013

In addition to correlations with other pollutants, daily average traffic flow at Marylebone Road have been compared to Black Carbon and Elemental Carbon concentrations over the period 2009 – 2013, as shown in Figure 36.

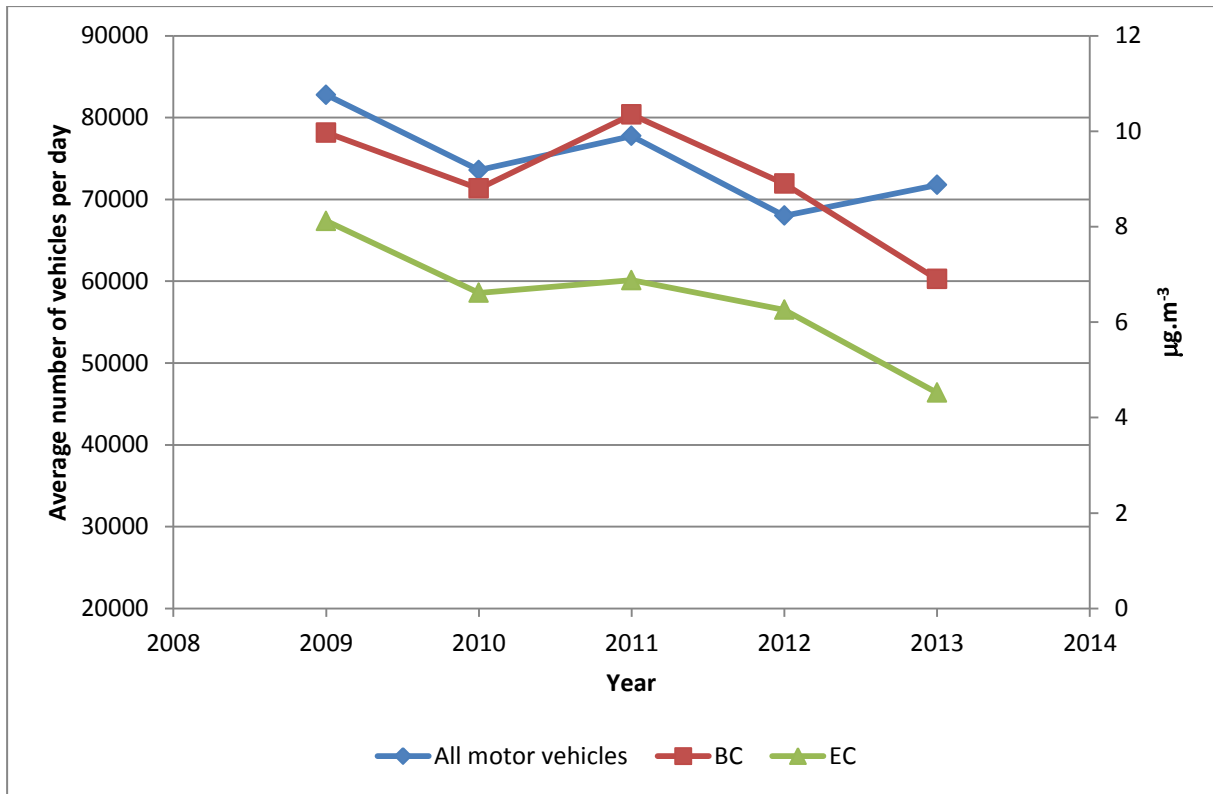


Figure 36 Annual Average Black Carbon, Elemental Carbon and Motor Vehicles per Day at Marylebone Road for the period 2009 – 2013

It can be seen that the changes in Black Carbon and Elemental Carbon concentrations follow changes in the total traffic flow for the years 2009 to 2012, but not for 2013.

5.5.1.2 UV Component

Figures 37 to 40 show the UV component trends.

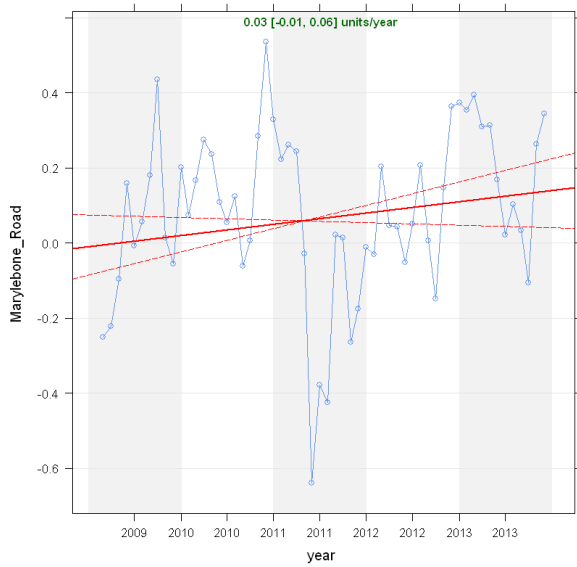


Figure 37 UV component concentrations measured at roadside sites, 2009 – 2013

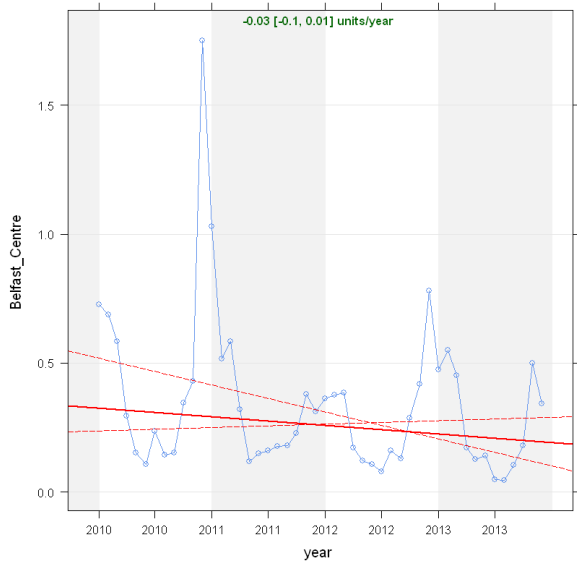


Figure 38 UV component concentrations measured at urban centre sites, 2009 – 2013

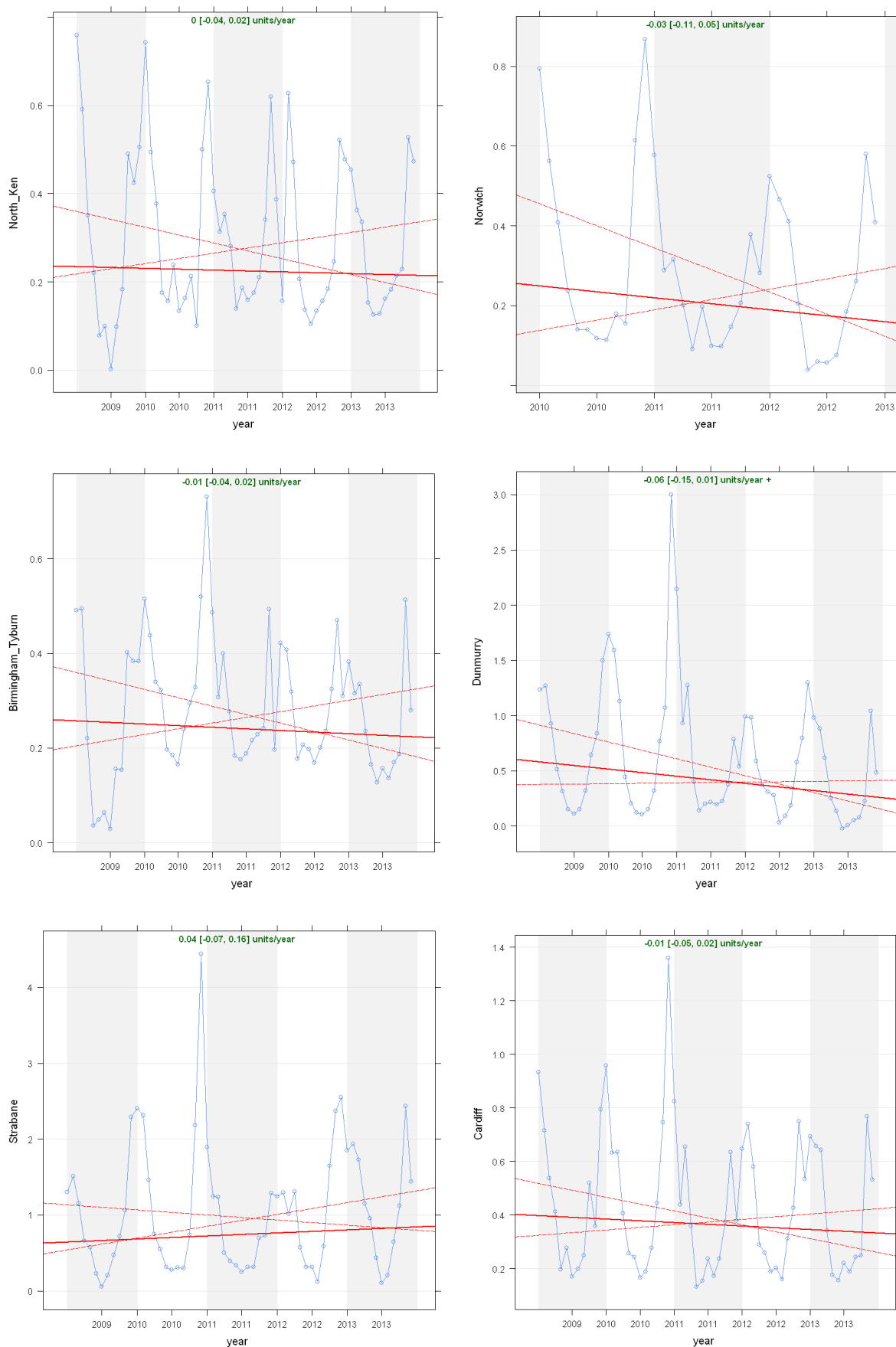


Figure 39 UV component concentrations measured at urban background sites, 2009 – 2013

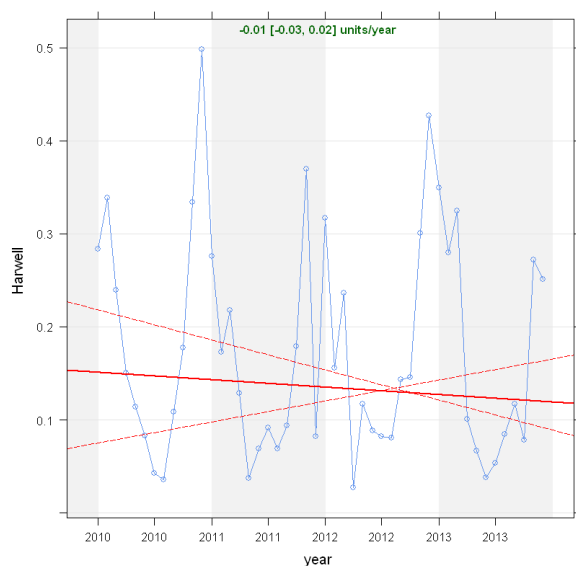


Figure 40 UV component concentrations measured at rural background sites, 2009 – 2013

Site	Slope	Lower limit	Upper limit	Slope significant
Roadside				
Marylebone Road	0.03	-0.01	0.06	N
Urban Centre				
Belfast Centre	-0.03	-0.10	0.01	N
Urban Background				
North Kensington	0.00	-0.04	0.02	N
Norwich	-0.03	-0.11	0.05	N
Birmingham Tyburn	-0.01	-0.04	0.02	N
Dunmurry	-0.06	-0.15	0.01	N
Strabane	0.04	-0.07	0.16	N
Cardiff	-0.01	-0.05	0.02	N
Rural				
Harwell	-0.01	-0.03	0.02	N

Table 20 Summary of UV component trends

No sites show a significant trend in UV component concentration between 2009 and 2013.

To show how pollutant concentrations can depend strongly on the weather, the 2009-2013 UV component concentrations at Strabane, which are strongly affected by domestic solid fuel use, are plotted in Figure 38, along with average temperature for same period. Temperature measurements from Armagh have been used as this is the nearest Met Office site with a long time series.

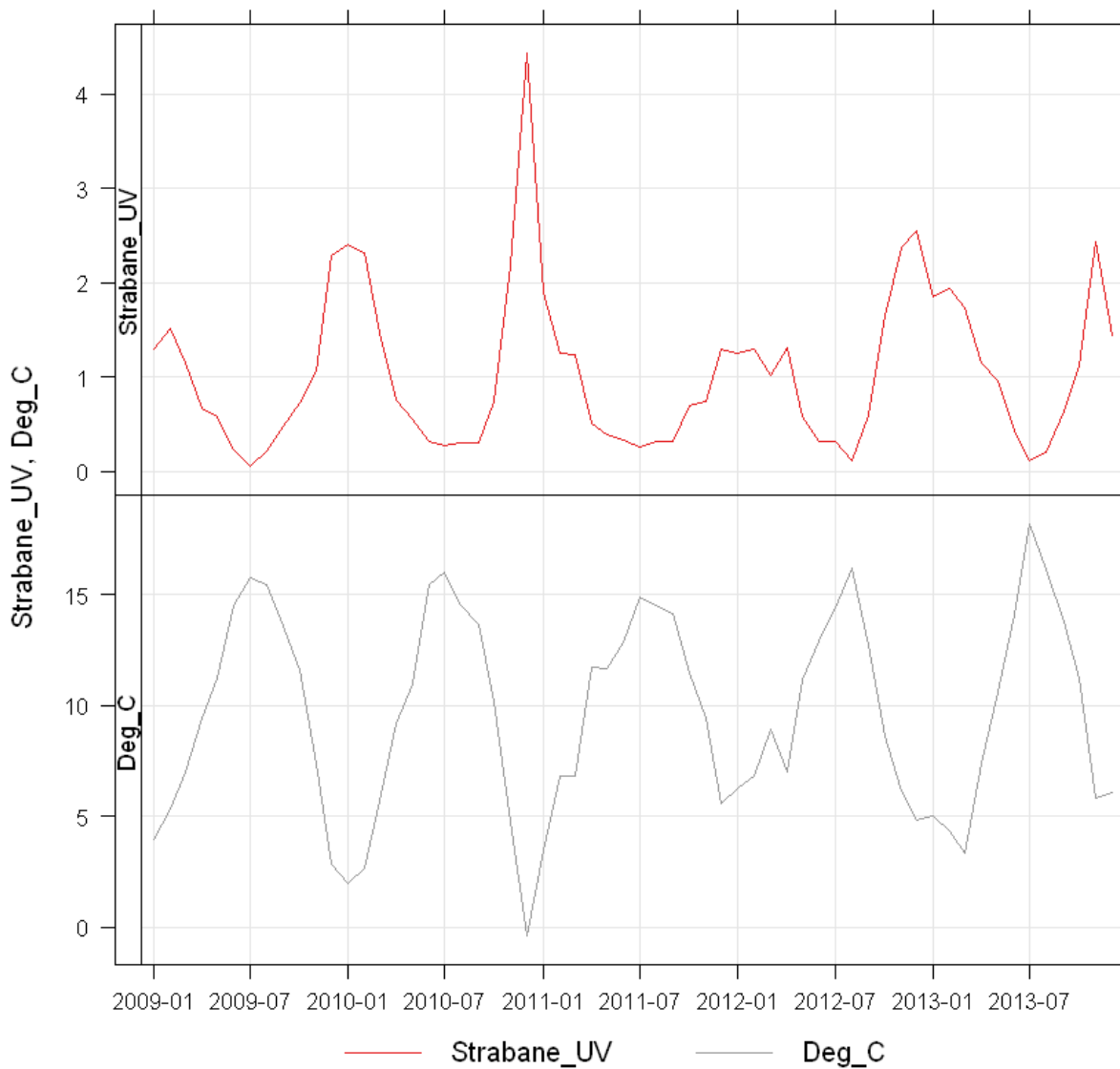


Figure 41 Strabane monthly UV component concentration and average ambient temperature for 2009 – 2013

It can be seen that the UV component concentration is inversely proportional to the average ambient temperature. This is due to the main source of UV component emissions being local domestic heating in Strabane. This relationship is also shown in Figure 39 as a scatter plot.

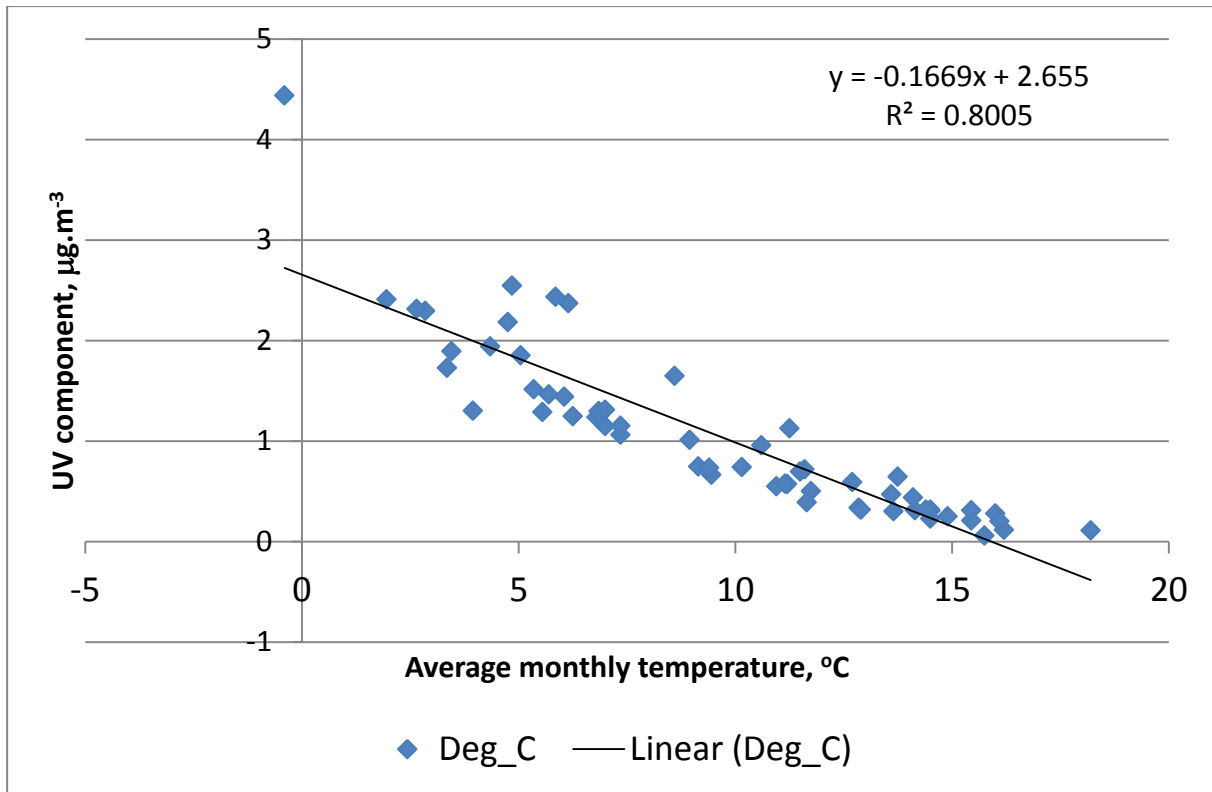


Figure 42 Scatter plot of monthly UV component concentration versus ambient temperature at Strabane over the period 2009 -2013

It can be seen that there is a clear linear relationship between increased UV component concentrations with a drop in ambient temperature, due to the increase in fuel usage in cold weather periods.

5.5.2 Network Short-Term Trends

Figures 43 and 44 show the Network annual mean and median concentrations for Black Carbon and UV component for the subset of sites that have been continuously running since 2009. The median concentration is shown to remove the influence of large changes in a single site that would skew the overall result for the Network.

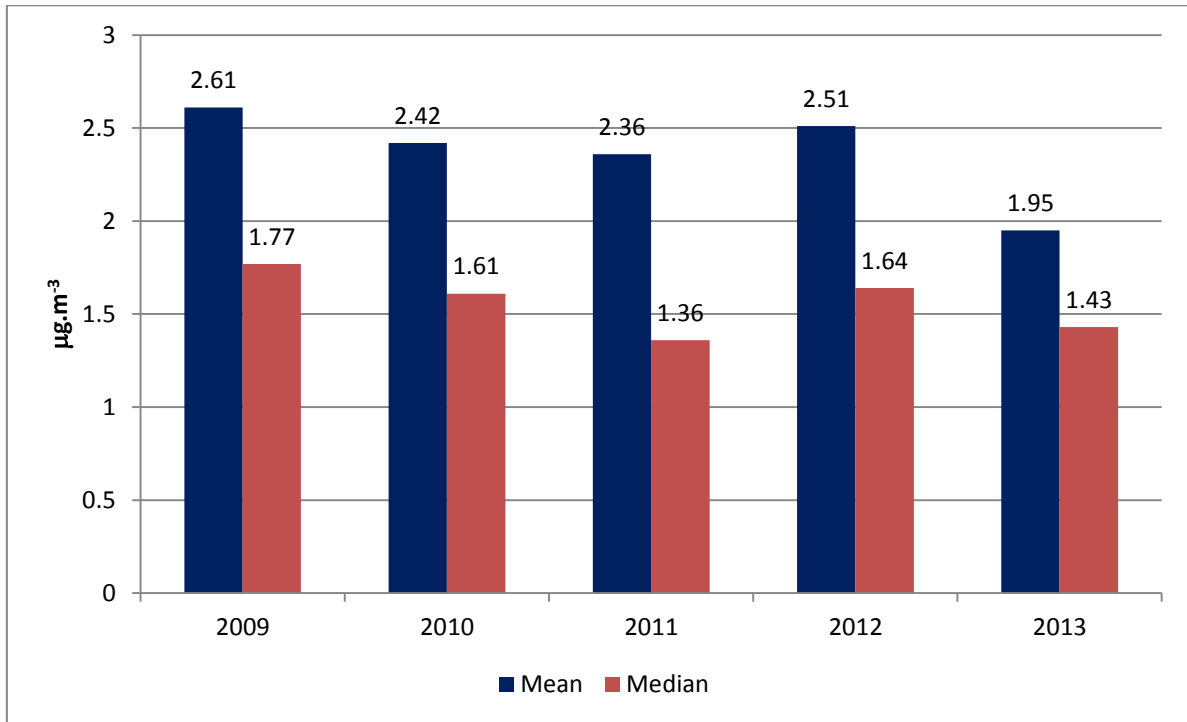


Figure 43 Network annual average Black Carbon concentrations

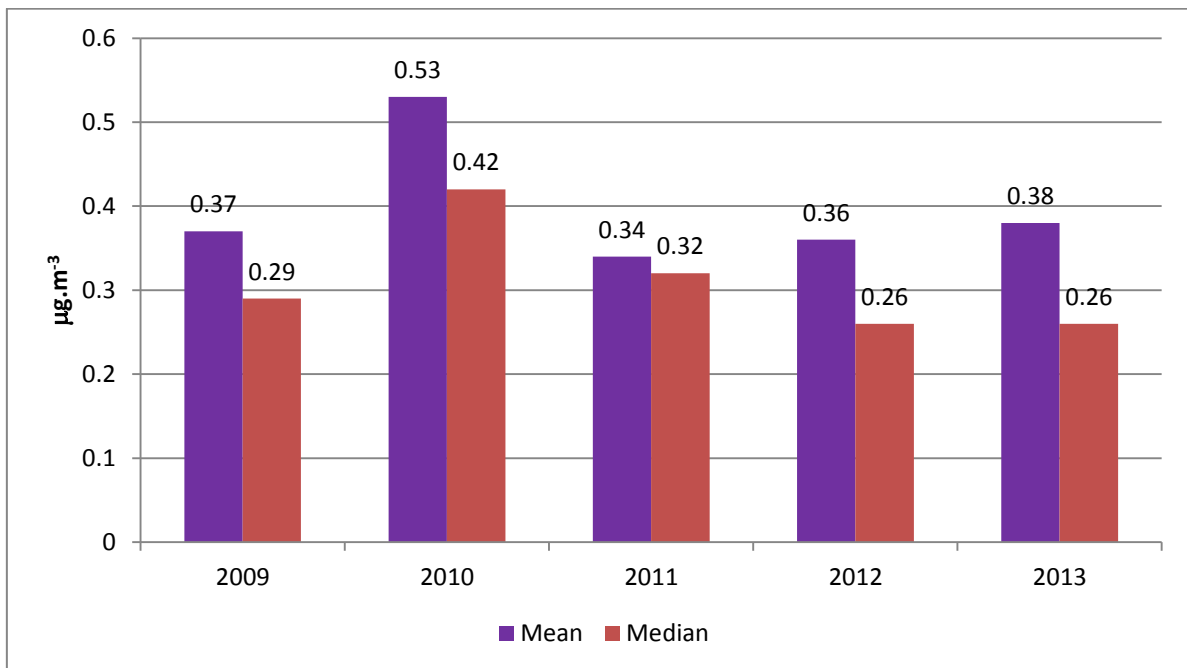


Figure 44 Network annual average UV component concentrations

It can be seen that there has been a considerable drop in the Network Annual mean for Black Carbon in 2013, while the median concentration is similar to previous years. The drop in average concentration is mainly driven by the 2013 drop in concentration at Marylebone Road. Concentrations for previous years are relatively similar. The drop in the median concentration in 2011 is due, with the exception of Marylebone Road, to most sites showing lower than average concentration in 2013.

Both the annual average and median UV component concentrations vary from year to year with maximum concentrations in 2010. 2010 had very cold and snowy periods at the start and end of the year, especially in Northern Ireland, as shown in Figure 41, so elevated concentrations are probably due to domestic emissions. Temperatures in 2012 and 2013 were in general warmer than those between 2009 and 2011.

5.5.2.2 Long term Black Carbon concentrations

Trends in annual average Aethalometer concentrations over the period 2009 to 2013 are given along with the long-term trend in Black Carbon concentrations by converting historical Black Smoke concentrations into Black Carbon concentrations. The relationship, described by Equation 1 below⁹, was used to convert Black Smoke Index measured between 2000 and 2008 into Black Carbon concentrations. For sites still continuing into 2013, these Black Carbon concentrations are given in Figure 45.

$$C_{BC} = \sqrt{4.18I_{BS} + 59.6} - 7.72 \quad \text{Eq 1}$$

where:

C_{BC} =Black Carbon concentration in $\mu\text{g.m}^{-3}$
 I_{BS} =Black Smoke Index

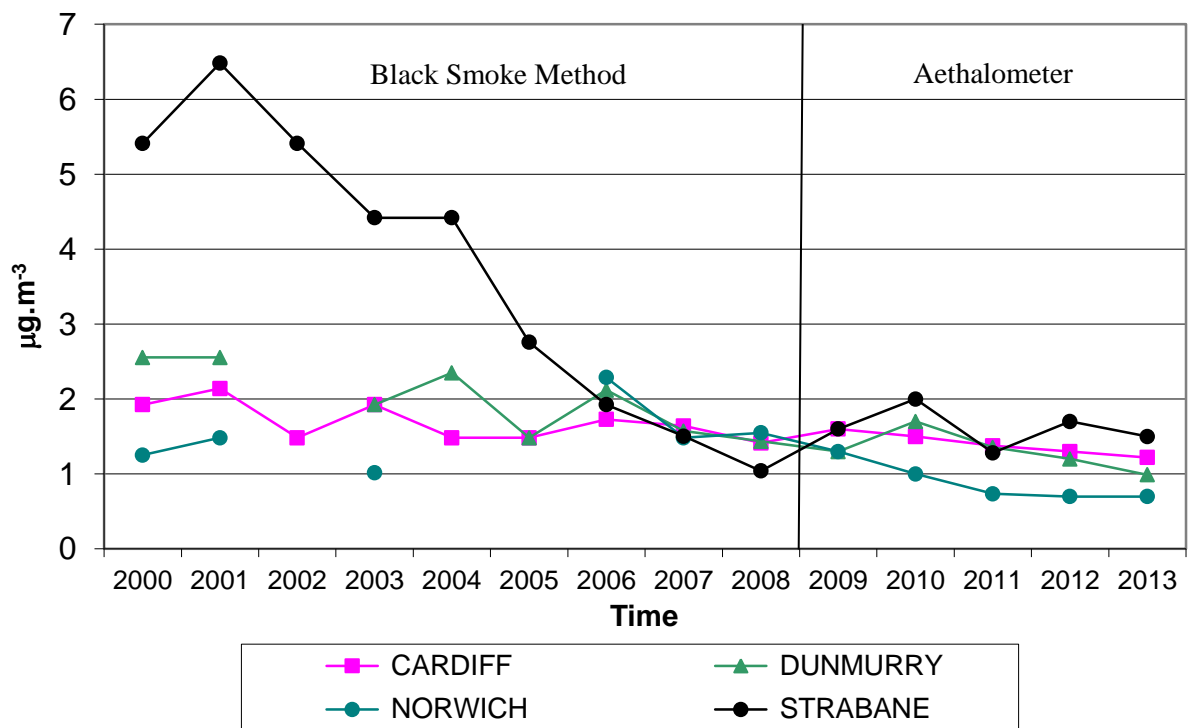


Figure 45 Trends in Black Carbon Concentrations 2000 to 2013

It can be seen that there is no significant discontinuity in results between the two methods and, apart from Strabane, there is no obvious long-term trend in the Black Carbon concentrations. In 2004 Strabane Council replaced the central heating systems in the local Council-controlled housing from coal to oil.

⁹ P Quincey, D Butterfield, D Green, G Fuller, Black Smoke and Black Carbon: further investigation of the relationship between these ambient air metrics. Atmospheric Environment, 2011, 45, (21), 3528-3534

6.0 EUROPEAN STANDARDISATION

The European standardisation body CEN is formulating a European Standard on the measurement of Elemental Carbon and Organic Carbon deposited on filters. Technical Report EN/TR 16243 has been published¹⁰. The standardisation working group are now performing validation work on the procedures and requirements laid out in this technical report. The work packages being performed are:

- WP1: Literature review
- WP2: Lab tests
- WP3: Field tests
- WP4: Statistical evaluation

The current progress is as follows:

WP1 The final draft of the literature review has been approved by the working group and publication is imminent.

WP2 Two inter-laboratory comparisons have been performed. The first demonstrated the requirement to perform regular instrument calibrations to ensure satisfactory performance from the thermal / optical method. The second intercomparison showed good agreement between the 4 laboratories.

WP3 Under this work package filter sampling is being performed at 6 field locations along with parallel sampling using automatic analysers. The automatic analysers being used are the Magee AE33 dual spot, 7-wavelength Aethalometer, and a Thermo Scientific Multi Angle Absorption Photometer (MAAP). The AE33 differs from the AE22 Aethalometer used on the Network in 2 ways. Firstly the AE22 is a 2 wavelength analyser. Secondly the AE33 does an internal correction for tape loading, while for the AE22 this has to be performed post sampling in the ratification process. Part of the remit of the working group is to comment on the comparability between the proposed reference method and automatic analyser methods. So far, five of the field validation campaigns has been completed.

WP4 The first interim report on the working group's progress has been approved by the working group and sent to CEN.

10 CEN/TR 16243: Ambient air quality - Guide for the measurement of elemental carbon (EC) and organic carbon (OC) deposited on filters, August 2011.

7.0 CONCLUSIONS

Black Carbon concentrations measured at most sites in 2013 were similar to those in previous years, with the exception of Marylebone Road, which showed a notable decrease. UV component concentrations in 2013 were very similar to previous years. The Network mean Black Carbon concentration in 2013 was $1.5 \mu\text{g.m}^{-3}$ whereas the 2012 concentration was $1.9 \mu\text{g.m}^{-3}$. These years are not totally comparable as the sites making up the Network were different. From looking at the trend in Black Carbon concentrations for only those sites that have been open from 2009 – 2013, the concentration was stable from 2009 to 2012, with a significant drop in average concentration from $2.51 \mu\text{g.m}^{-3}$ in 2012 to $1.95 \mu\text{g.m}^{-3}$ in 2013 (22%), while the median concentration has dropped from $1.64 \mu\text{g.m}^{-3}$ to $1.43 \mu\text{g.m}^{-3}$ (13%). The median concentration is much less susceptible to changes at one site. The 2013 UV component concentration was $0.3 \mu\text{g.m}^{-3}$, which is the same as 2012. The range of concentrations between roadside and rural background were also similar.

At individual sites, short term trends in Black Carbon concentrations measured by the Aethalometer (2009 – 2013) are insignificant at all stations except Marylebone Road. Black Carbon concentrations at this site have dropped over the last 2 years in line with changes in total traffic volume.

There are no significant trends in UV component concentration at any of the sites, or in the annual average or median concentration across the history of the Network.

The new network design implemented in early 2012 allows urban increments in Black Carbon and UV component concentrations to be determined for London, Birmingham and Glasgow, and roadside increments for London and Birmingham. The urban increment for Black Carbon was similar for all locations while the roadside increment was roughly proportional to road traffic volumes. There was no significant urban or roadside increment in UV component concentration.

Diurnal average concentrations of Black Carbon show that the dominant emission sources are road traffic and domestic heating using non-smokeless fuel. Diurnal average concentrations of UV component show that the dominant emission source is domestic heating using non-smokeless fuel, with little to no impact from road traffic.

APPENDIX 1 MEASUREMENTS BY THE CARBOTRAF PROJECT IN GLASGOW

An Aethalometer was installed in the Byres Road monitoring site in Glasgow on 7th November 2012 as part of the European CARBOTRAF Project, in association with Air Monitors Ltd. This project will run to the end of 2014. Details on the project can be found at the following URL:

<http://www.carbotraf.eu/>

The following project aim is taken from their website:

“Realize a method, system and tools for adaptively influencing traffic in real-time to reduce carbon dioxide CO₂ and black carbon (BC) emissions caused by road transport in urban and inter-urban areas. The inter-relationships between traffic states and CO₂ and BC emissions will be investigated. In particular a model linking traffic states to emission levels will be established on the basis of existing and new simulation methods and tools. A decision support system for online prediction of emission levels will use real-time and simulated traffic and air-quality data. Based on this prediction a low emission traffic scenario will be achieved by imposing ITS measures (re-routing, adjustment of traffic light sequences)”.

The Byres Road monitoring site is classified as a roadside site and is approximately 5m from the road and 25m from the nearest road junction. Details of the site can be found at:

<http://www.scottishairquality.co.uk/>

The Byres Road site is not part of the Network. The Aethalometer is run on a time base of 1 minute instead of the Network standard of 5 minutes. The 1 minute data has been averaged into 15 minute data and then ratified using the same procedures as the Network, except that no QA/QC audit of the analyser has been performed by NPL. Figures 46 and 47 show the time series of the black carbon and UV component concentrations from November 2012 to the end of 2013.

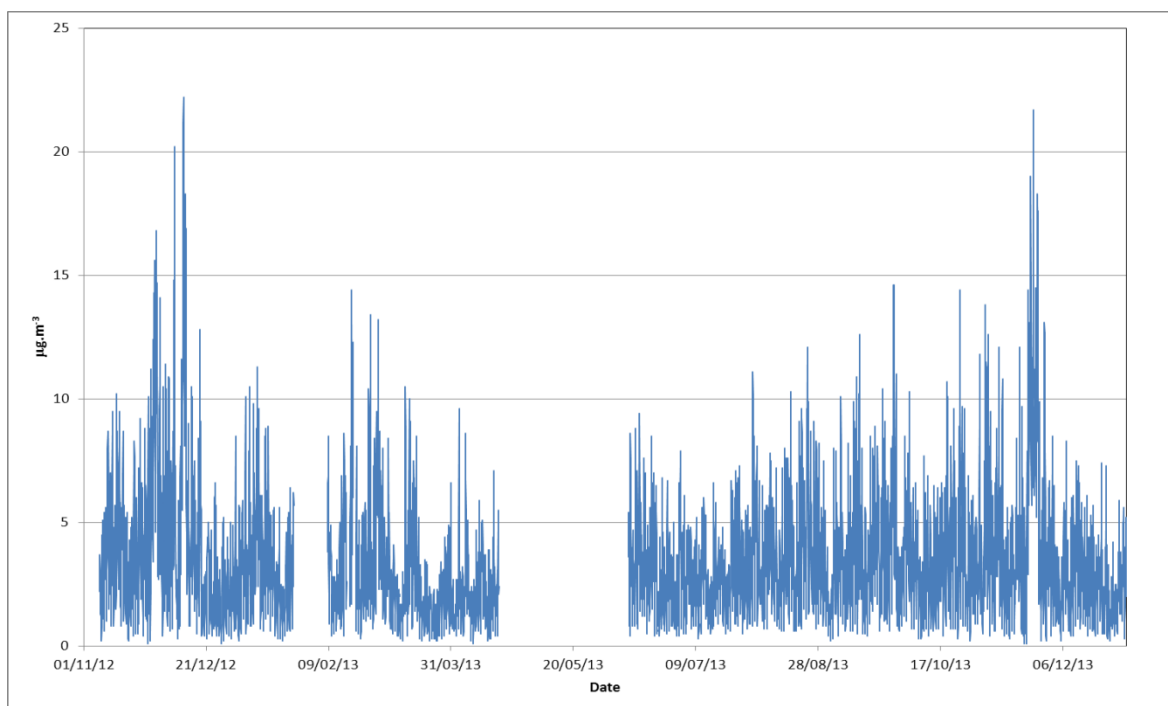


Figure 46 Black Carbon concentrations at Byres road for the period November 2012 to December 2013

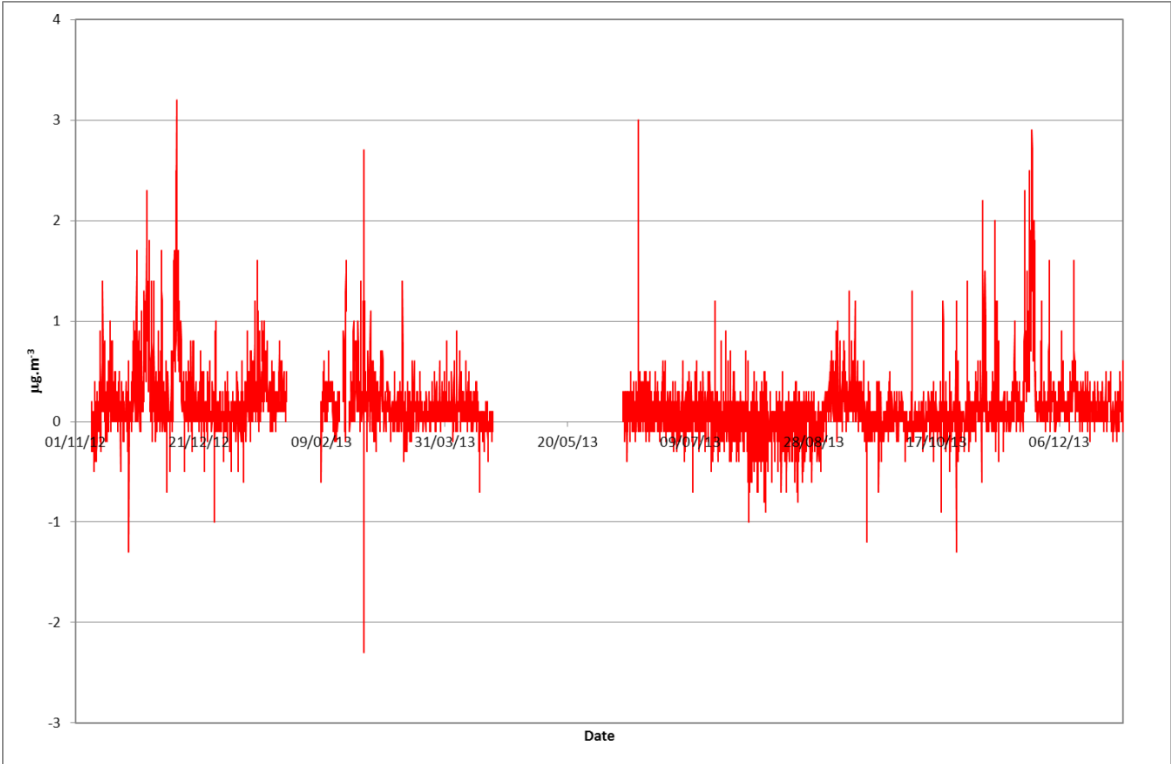


Figure 47 UV component concentrations at Byres road for the period November 2012 to December 2013

The analyser suffered some reliability problems which resulted in large data gaps from April to June 2013, giving a data capture for the complete period of 83%.

The average black carbon concentration over the period is $3.2 \mu\text{g.m}^{-3}$ and the average UV component concentration is $0.1 \mu\text{g.m}^{-3}$, these are very similar to the values recorded at Birmingham Roadside during 2013 ($3.2 \mu\text{g.m}^{-3}$ and $0.2 \mu\text{g.m}^{-3}$ respectively).

Figure 48 shows the temporal variation of the black carbon and UV component concentrations.

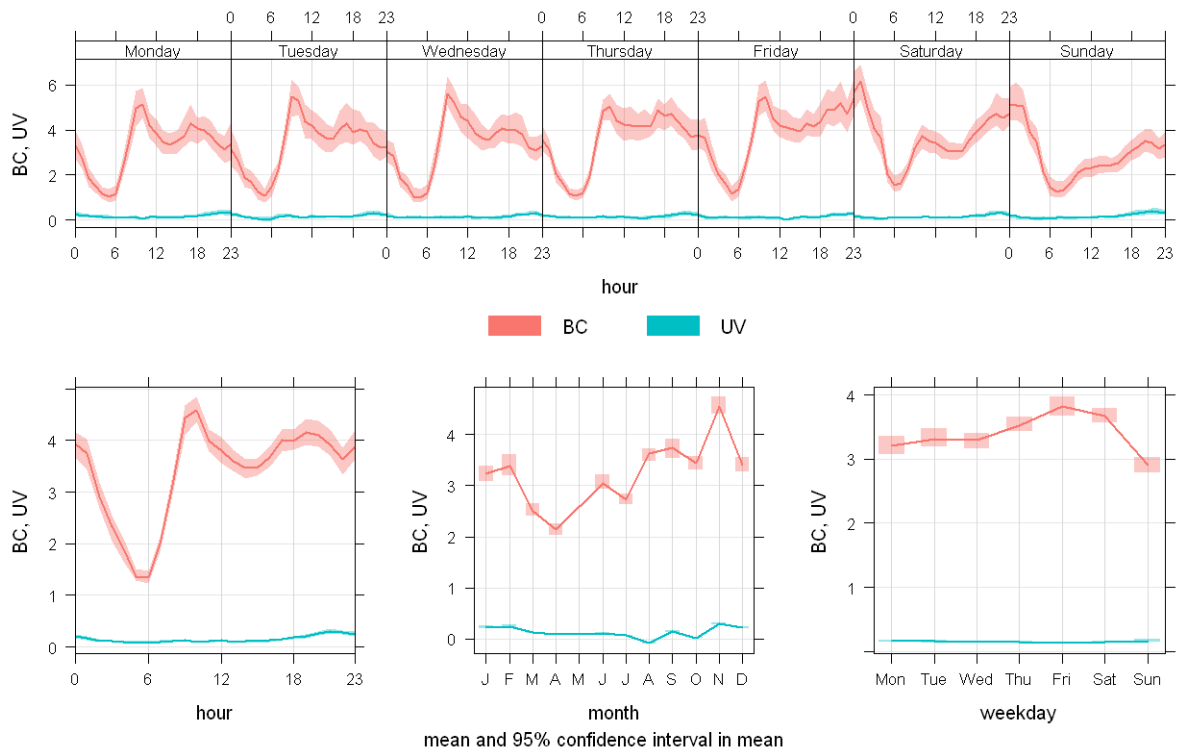


Figure 48 Temporal variations in black carbon and UV component at Byres Road

The time profiles of the Byres Road site are very similar to roadside sites on the Network.

Comparisons with Traffic Data

There are two “smart-eye” traffic cameras situated at the junction of Byres Road and Highburgh Road, about 25m from the monitoring station. These cameras have an uncertainty of ~10% when compared to manual count data. Traffic data has been collected from 1st October to 31st December 2013. Figures 49 to 52 show the time variation of Black Carbon concentrations, hourly total traffic flow for all vehicles, and hourly traffic flow from the individual roads during this period.

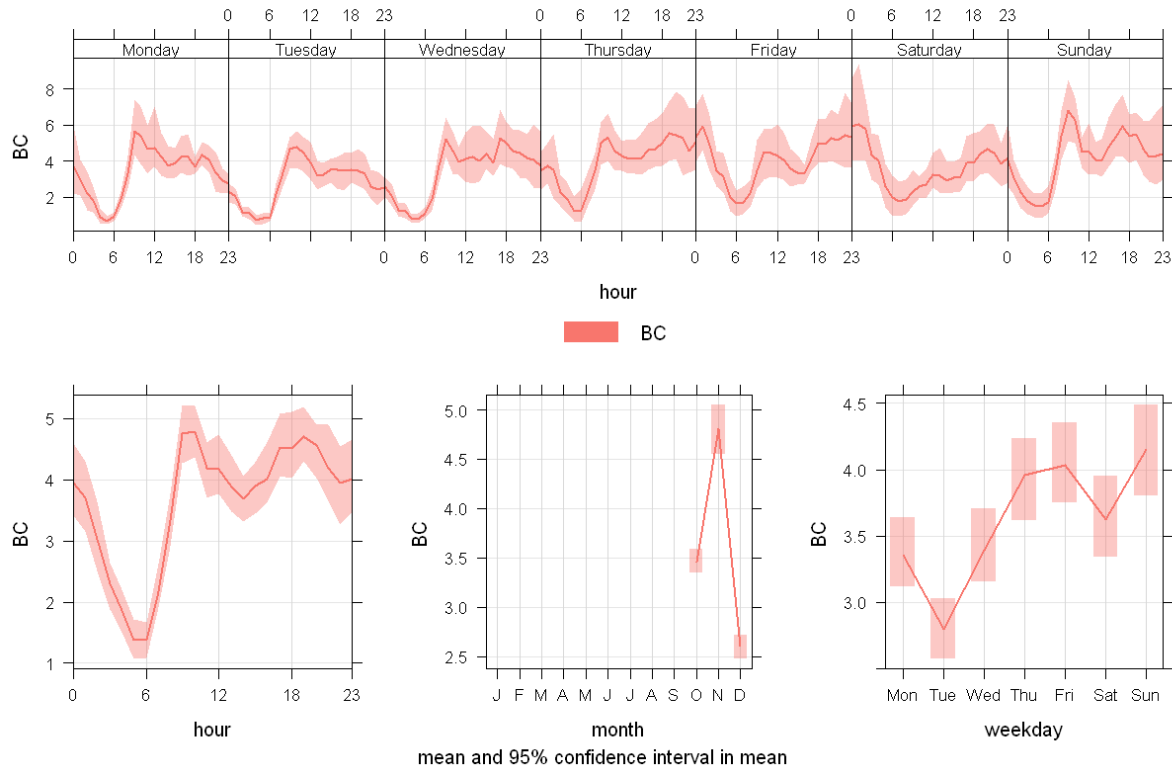


Figure 49 **Variability in Black Carbon concentrations**

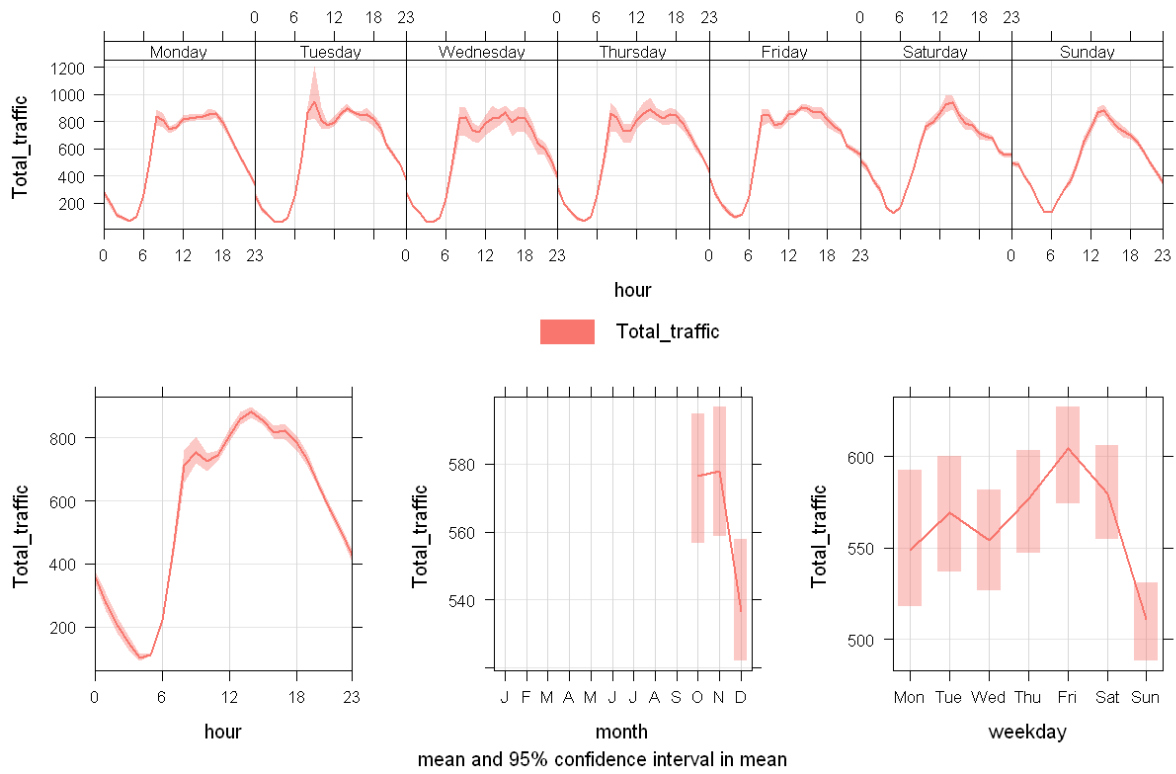


Figure 50 Variability in the sum of both counters for all vehicles at the junction

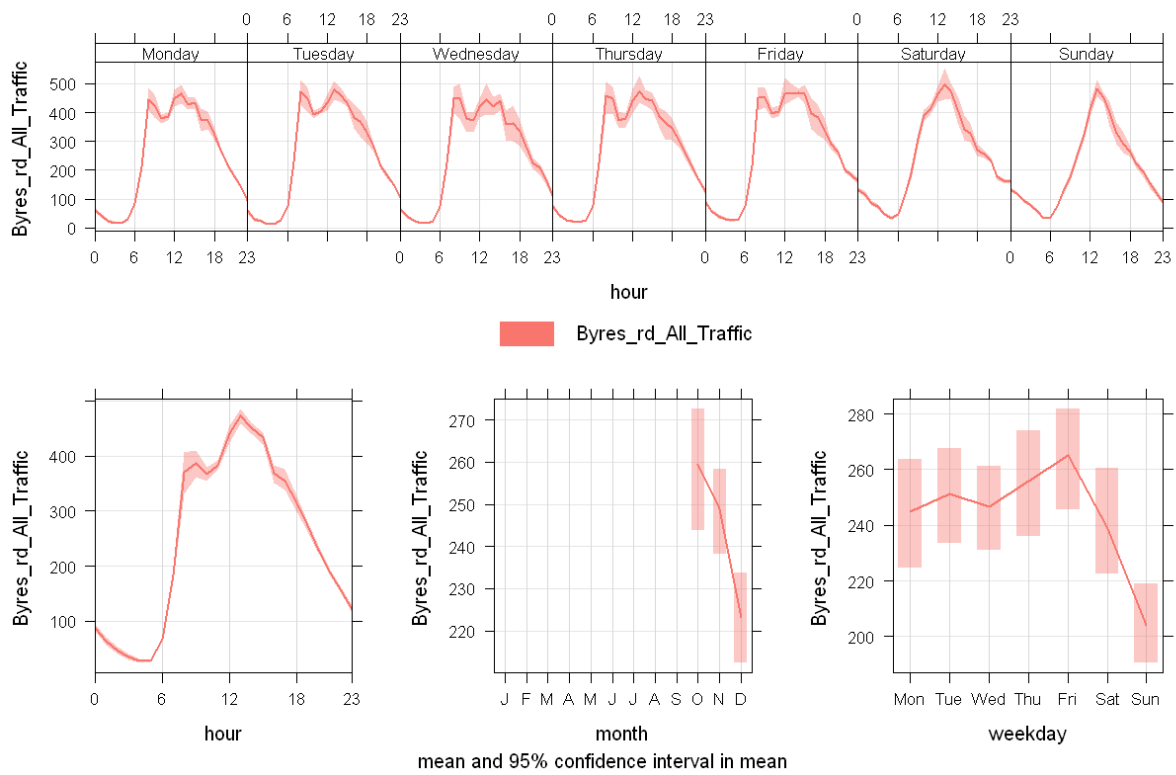


Figure 51 Variability in traffic numbers for all vehicles on Byres Rd

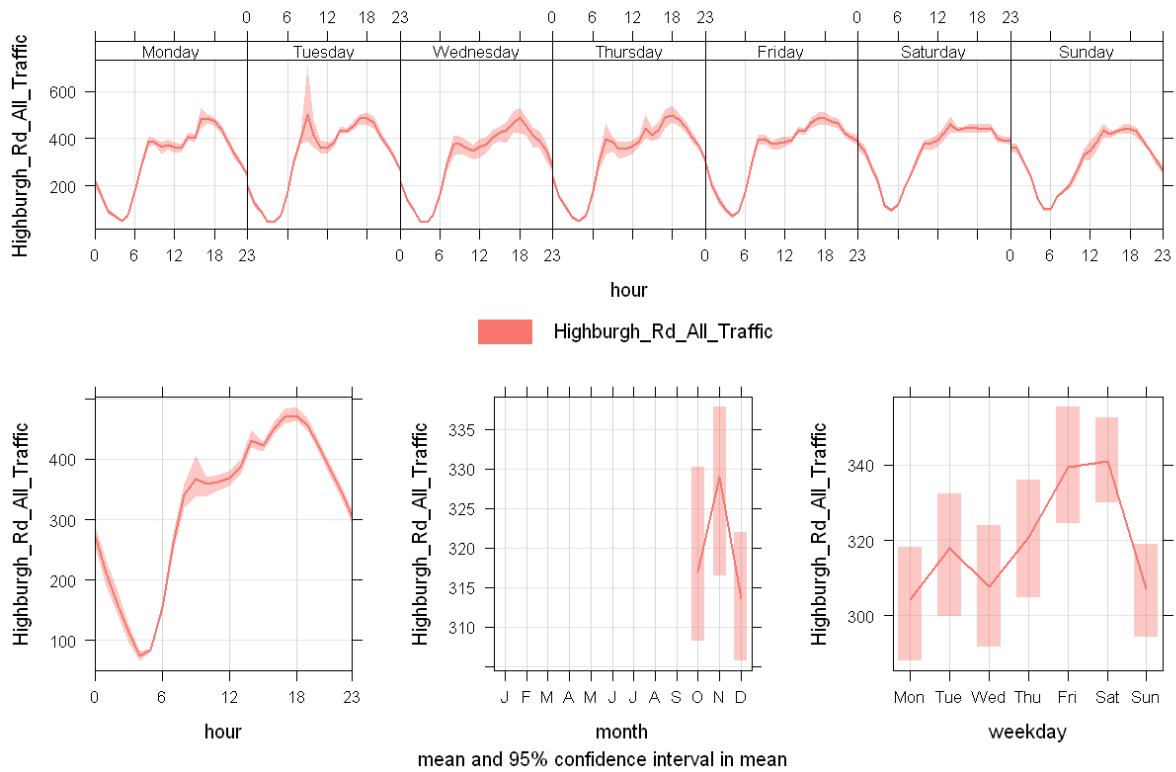


Figure 52 Variability in traffic numbers for all vehicles on Highburgh Rd

It can be seen that the traffic flow on Byres Road follows the profile expected of the daily morning and evening commute for week days, with a different profile at the weekend with peak traffic flow around midday and less vehicles per hour. The Highburgh Road has a morning commuter signature, but has peak vehicle flow around 18:00 hours and higher sustained vehicle flow into the late evening compared to Byres Road.

The combination of the two traffic patterns gives elevated black carbon concentrations from the morning commute until well into the evening.