

Air Pollution Trends Report 2023: Critical Load and Critical Level exceedances in the UK

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Executive Summary

This contract provides key information to track the effects on ecosystems of policies aimed at meeting national and international air pollution targets, e.g. under the UK Government's Clean Air Strategy 2019 (CAS), the National Emission Ceilings Regulations (NECR), and the United Nations-Economic Commission for Europe Convention on Long-range Transboundary Air Pollution (CLRTAP). It also provides the means to develop targeted action for emission reduction policies to get the maximum improvement in air quality.

Exceedance of critical load or critical level indicates that an ecosystem is at risk from potential harmful effects. When pollution subsequently decreases to below the critical load or level, there may be delays to recovery, but the risk of harm is reduced. Pollution control measures aimed at meeting CLRTAP targets have reduced the extent and magnitude of critical load and level exceedances across the UK. A target for reactive nitrogen (N) deposition in England (a percentage reduction compared to 2016) was included in the CAS to further reduce impacts on ecosystems in response to the UK's new emission reduction targets for ammonia and oxides of nitrogen.

Critical loads for nutrient N were reviewed and revised in 2022, and new mapping values have been applied in this latest Air Pollution Trends Report. Values were recalculated for previous years using the new critical load values, to allow time series to be reported with consistent methodology. Effects of these changes on the values of the reported exceedance metrics are discussed and illustrated. The values of metrics have changed, in some cases considerably, but the trends in air pollution pressure over time are similar using the old and new methods.

- UK habitats at risk from acidification: The area of acid-sensitive habitats with exceedance of acidity critical loads has fallen from 65.6% in 2003, to 44.5% (40,252 km²) in 2020, due mainly to decreases in NO_x and sulphur deposition. Of the total acidifying pollution onto UK woodland in 2020, 70 % was reduced N (ammonia and ammonium), 21 % was NO_x, and 9 % was sulphur.
- UK habitats at risk from excess N (eutrophication): The area of N-sensitive habitats with exceedance of nutrient N critical loads fell from 93.8% (87,954 km²) in 2003, to 85.9% (80,528 km²) in 2020. This decline was largely driven by changes in Scotland – in England, Wales and Northern Ireland there was little change in area exceeded over this period.
- UK habitats at risk of exceeding the critical levels for gaseous ammonia: In 2020, 2.6% (6,386 km²) of the UK land area was exposed to ammonia concentrations above the critical level set to protect higher plants (3 µg m⁻³), and 64.3% (157,367 km²) exposed to ammonia at concentrations above the critical level set to protect lichens and mosses (1 µg m⁻³). The area where the critical level for higher plants is exceeded has increased by 0.9% of UK land area since 2003. The area where the critical level for lichens and mosses is exceeded has increased by 7.6% of UK land area since 2003.

Technical Summary

Critical loads define the rates of acid or nitrogen (N) deposition (e.g. in kiloequivalents per hectare per year, $\text{keq ha}^{-1} \text{yr}^{-1}$) below which significant harmful effects are not expected to occur in sensitive habitats. Critical load exceedance is the amount by which acid or N deposition exceeds the critical load.

Critical levels are concentrations of pollutants (e.g. in micrograms per cubic metre, $\mu\text{g m}^{-3}$) in the atmosphere below which direct adverse effects on receptors, such as human beings, plants, ecosystems or materials, are not expected to occur according to present knowledge. Critical level exceedance is the amount by which the concentration exceeds the critical level.

Critical load and critical level exceedances are calculated using rolling 3-year mean data sets for deposition rates and gaseous concentrations, which are updated annually. A 3-year mean is used to smooth out inter-annual variability due to the influence of weather on atmospheric chemistry. This report describes critical loads and their exceedances, and critical levels and their exceedances, for the period from 2003 to 2020. In previous Trends Reports, ammonia concentrations were only available up until the year preceding the deposition data, but a change of methodology (see Section 4.2) has allowed the concentration data to catch up with the deposition data.

This report presents trends in: a) acidity critical load exceedances for UK habitats at risk from acidification by excess sulphur (S) and/or N; b) nutrient-nitrogen critical load exceedances for UK habitats at risk of eutrophication by excess N; c) critical load exceedances for acid- and N-sensitive habitat features of UK designated sites (Special Areas of Conservation: SACs; Special Protected Areas: SPAs; Sites of Special Scientific Interest: SSSIs); and d) exceedances of ammonia critical levels across the UK.

Summary statistics monitor progress in the area of habitats in the UK at risk from acidification and eutrophication from air pollution over time, and are reported in the annual “UK Biodiversity Indicators” summary produced by the Joint Nature Conservation Committee (<https://jncc.gov.uk/our-work/ukbi-b5a-air-pollution/>; indicator B5a for assessing the pressures from air pollution).

UK habitats at risk from acidification

- The area of acid-sensitive habitats in the UK with exceedance of acidity critical loads continued to decline due mainly to decreases in oxidised N (NO_x) and sulphur (S) deposition, having fallen from 65.6% (59,407 km^2) in 2003 to 44.5% (40,253 km^2) in 2020.
- Excess acidity (expressed as the Average Accumulated Exceedance) for all UK habitats combined fell between 2003 and 2020, from 0.58 to 0.25 $\text{keq ha}^{-1} \text{year}^{-1}$.
- The largest reduction in the area of acid-sensitive habitats with exceedance of acidity critical loads was in Scotland, where it fell from 56.7% (30,896 km^2) in 2003 to 28.9% (15,753 km^2) in 2020.
- The smallest reduction in the area of acid-sensitive habitats with exceedance of acidity critical load was in Northern Ireland, falling from 79.3% (2,797 km^2) in 2003 to 72.8% (2,566 km^2) in 2020.
- Of the terrestrial acid-sensitive habitats mapped, dwarf shrub heath occupies the largest area across the UK (24%); the area of this habitat with exceedance of acidity critical loads decreased, from 34.0% (7,435 km^2) in 2003 to 13.9% (3,042 km^2) in 2020.

UK habitats at risk from eutrophication (i.e. from excessive nutrient availability)

- The area of N-sensitive habitats in the UK with exceedance of nutrient N critical loads decreased from 93.8% (87,954 km²) in 2003, to 85.9% (80,528 km²) in 2020.
- Excess Nitrogen (Average Accumulated Exceedance for nutrient N) for all UK habitats combined decreased from 10.6 kg N ha⁻¹ year⁻¹ in 2003 to 7.6 kg N ha⁻¹ year⁻¹ in 2020.
- The largest reduction in the area of N-sensitive habitats with critical load exceedance was in Scotland, falling from 89.3% (48,529km²) in 2003 to 75.8% (41,200 km²) in 2020.
- Reductions between 2003 and 2020 in the area of N-sensitive habitats with critical load exceedance in England, Wales and Northern Ireland were less than 1%.
- The nutrient N critical load was exceeded for more than 90% of the areas of seven N-sensitive habitats in all years: calcareous grassland, montane habitats, beech woodland, acidophilous oak woodland, other broadleaved woodland, scots pine woodland, and mixed woodland.
- Although the decline between 2003 and 2020 in area exceeded was small for most habitats (e.g. from 100.0% to 99.5% of beech woodland), the magnitude of exceedance (Excess Nitrogen) for beech woodland decreased more substantially, e.g. from 18.6 kg N ha⁻¹ year⁻¹ in 2003 to 17.2 kg N ha⁻¹ year⁻¹ in 2020.

Total N deposition onto protected sensitive habitats

- Since 2020, the Trends Report includes a metric against which progress towards the UK Government's Clean Air Strategy target (Defra, 2019) can be measured, i.e. "to reduce damaging deposition of reactive forms of nitrogen by 17% over England's protected priority sensitive habitats by 2030".
- The mean N deposition rate onto priority habitats in England was 22.7 kg N ha⁻¹ yr⁻¹ in 2016, and 20.6 kg N ha⁻¹ yr⁻¹ in 2020, representing a 9.4 % decrease.

Designated sites with acid-sensitive feature habitats

- The percentage of SACs and SPAs in the UK with exceedance of acidity critical loads for one or more features decreased from 80 % and 73.7% (SACs: 389 sites, SPAs: 129 sites) in 2003 to 67.7% (SACs: 330 sites) and 57.1% (SPAs: 100 sites) in 2020.
- The percentage of SSSIs with exceedance of at least one sensitive feature fell from 64.5% (3,021 sites) in 2003 to 49.2% (2,305 sites) in 2020.
- Scotland had the largest reductions in the percentage of designated sites with exceedance of acidity critical loads between 2003 and 2020, e.g. 20.3% fewer SACs.
- In England, 48.1 of SSSIs had exceedance of acidity critical loads in 2020, as did 63.3 % of SSSIs in Wales, 37.8 % of SSSIs in Scotland, and 71.9 % of ASSIs in Northern Ireland.

Designated sites with nitrogen-sensitive feature habitats

- There were small reductions in the percentage of designated sites in the UK with exceedance of nutrient N critical loads (for one or more features) between 2003 and 2020 (2.2 % for SACs, 1.8% for SSSIs, 4.9% for SPAs). This reflects the downward trend in N deposition during 2020.
- Scotland had the largest reduction (8.9%) in the percentage of SPAs with exceedance of nutrient N critical loads between 2003 and 2020.
- Of nature conservation sites in England, Wales and Northern Ireland, 98–100 % had exceedance of nutrient N critical loads for one or more features in 2020. Proportionally fewer sites in Scotland (76.6-89.1%, depending on designation) were exceeded in this way.

Exceedance of ammonia critical levels

The trends in ammonia critical levels exceedance are available for the period 2003 to 2020; there have only been small changes in ammonia concentrations over this time period.

UK land area

- 64.3% of the UK currently was exposed to ammonia concentrations above the critical level set to protect lichens and bryophytes ($1 \mu\text{g m}^{-3}$) in 2020; this represents 90.5% of England, 59.7% of Wales, 15.8% of Scotland and 97.1% of Northern Ireland.
- There was an increase in the UK land area with ammonia concentrations above $1 \mu\text{g m}^{-3}$, from 56.7% in 2003 to 64.3% in 2020.
- About 2.6% of the UK was exposed to ammonia concentrations above the critical level set to protect higher plants ($3 \mu\text{g m}^{-3}$) in 2020; this ranged from 0% of Scotland to 16.7% of Northern Ireland.
- The UK land area with ammonia concentrations above $3 \mu\text{g m}^{-3}$ increased, from 1.7% in 2003 to 2.6% in 2020.

Nitrogen-sensitive habitats

- 30.1% of the mapped area of N sensitive habitats in the UK was exposed to ammonia concentrations above the critical level of $1 \mu\text{g m}^{-3}$ in 2020; the percentage area exceeded ranged from 5.1% for Scotland to 93.2% for Northern Ireland.
- 84.7% of the area of calcareous grassland, and ~ 84 - 93 % of the area of some woodland habitats (beech and broadleaved woodland) were in areas exposed to ammonia concentrations above the critical level of $1 \mu\text{g m}^{-3}$ in 2020. Only 13.2 % of dwarf shrub heath was in areas exceeding this critical level, but this equates to 2,822 km².
- 0.5 % of the area of N-sensitive habitats in the UK was exposed to ammonia concentrations above $3 \mu\text{g m}^{-3}$ in 2020; this ranged from 0 % in Scotland to 5.7 % in Northern Ireland.

Designated sites (SACs, SPAs and SSSIs or ASSIs)

- 46.4% (115) of SPAs, 60% (368) of SACs and 72 % (4,896) of SSSIs (ASSIs in Northern Ireland) in the UK were exposed to ammonia concentrations above $1 \mu\text{g m}^{-3}$ on at least part of the site in 2020.
- 89-95% of the designated sites in England, 99-100% of sites in Northern Ireland, 60-75% of sites in Wales, and 12-21% of sites in Scotland, were exposed to ammonia concentrations above $1 \mu\text{g m}^{-3}$ in 2020.
- 3.3% (20) of SACs, 2.8% (7) of SPAs and 2% (137) of SSSIs in the UK were exposed to ammonia concentrations above $3 \mu\text{g m}^{-3}$ on at least part of the site in 2020. The percentage of designated sites with exceedance of this critical level rose by 0.7 to 2.0 % between 2003 and 2020, depending on the designation.
- Only 0.1% of SSSIs in Scotland (and no SACs or SPAs) were exposed to ammonia concentrations above the critical level of $3 \mu\text{g m}^{-3}$ on at least part of the site, compared with 0.9 % of SSSIs in Wales (1.2 % of SACs and no SPAs, 2.2 % of SSSIs in England (3.5 % of SACs and 5.2 % of SPAs) and 12.8 % of ASSIs in Northern Ireland (14.8 % of SACs and 14.3 % of SPAs).

Report structure

Section 1 of this report provides an overview of critical loads for acidity and for nutrient N, deposition data, and exceedance calculations and metrics for habitats across the whole area of the UK. This is followed in Section 2 by summaries of these trends for specific habitats and countries, i.e. England, Scotland, Wales, and Northern Ireland. Section 3 focuses on designated sites, the application of “site-relevant critical loads” (SRCL) to these sites, and trends in their exceedances. Section 4 addresses critical levels for ammonia and their exceedances. Finally, Section 5 focuses on N deposition onto sensitive habitats, which is the basis of a target in the UK Government’s Clean Air Strategy (CAS) (Defra, 2019).

Notes on rounded numbers and percentages

Numbers in tables are shown to one decimal place. Numbers reported in the main text were rounded after calculation, so may not always equate precisely to the differences between the numbers in the table. For example, if there is a change of +0.08 units from 72.06 to 72.14, the real change is $72.14 - 72.06 = +0.08$, which rounds to +0.1, and the rounded numbers would show as $72.1 - 72.1 = 0.1$. This is not an error.

Changes in the area of habitat where critical loads or critical levels are exceeded are always expressed in terms of absolute percentage of the total habitat area, not as a relative percentage change from the previous value, nor as an absolute percentage of total land area for the country. For example, if a habitat occupies 10% of the UK, a change from 40% to 30% of the habitat area being exceeded would be expressed as a decrease of 10% (i.e. 40 minus 30), not as a decrease of 25% (i.e. $100 \times (40 - 30) / 40$), nor as a decrease of 1% (i.e. 4% of UK area minus 3% of UK area).

Section 1. Pollutant deposition, critical loads, and exceedances

1.1 Overview of deposition and critical loads

Pollutant deposition rates, expressed for example in $\text{kg ha}^{-1} \text{yr}^{-1}$, are useful indicators of pressure on ecosystems (Rowe et al., 2017). However, ecosystems are considered able to withstand a certain amount of pollution, and this amount is expressed as the critical load. Critical loads are thus thresholds for effects from atmospheric deposition and are defined as “a quantitative estimate of the exposure to one or more pollutants below which significant harmful effects on specified sensitive elements of the environment do not occur according to present knowledge” (Nilsson and Grennfelt, 1988).

The methods used to calculate and map UK critical loads are described in detail in Hall et al. (2015). Critical loads are calculated and mapped for UK habitats sensitive to acidification and/or eutrophication (Table 1.1). Here and in Section 2, results are presented for the entire mapped extent of habitats; results for designated sites are presented in Section 3. The critical load methods applied in the UK are based on methods approved at international workshops held under the Convention on Long-Range Transboundary Air Pollution (CLRTAP) and published in the “Mapping Manual” (CLRTAP, 2017).

A metric has been included to illustrate progress towards the aim of the UK Government’s CAS (Defra, 2019), to develop a target to achieve a “reduction of damaging deposition of reactive forms of nitrogen by 17% over England’s protected, priority, sensitive habitats by 2030”. The metric is here defined as “total deposition of reactive N onto nutrient-N sensitive priority habitat”, i.e. including priority habitat that is not within protected sites. Total N deposition does not take into account critical load, but is a readily-understood indicator of overall pressure on sensitive ecosystems. The baseline year against which the 17% reduction will be assessed is 2016. The derivation of this metric is discussed further in section 1.4.

The Devolved Administrations are considering whether to develop nitrogen deposition reduction targets for their own countries. Progress towards country-specific targets may be presented in future Trends Reports.

1.1.1 Habitat mapping

Until the Trends Report 2021 (Rowe et al., 2021), habitat distribution maps for exceedance calculations were based principally on the UKCEH Land Cover Map 2000 (LCM2000: Fuller et al., 2002) and additional data sets such as species distribution data and altitude. The habitat maps have now been updated and are now based principally on a more recent Land Cover Map, LCM2019, as described in Carnell et al. (2022). Montane habitat was not mapped in LCM2019, so the original map, based on altitude and latitude, was used for this habitat. A recent map of UK peat distribution (Evans et al., 2017) was used to apply distinct methods for woodland on peat and non-peat soils (see Section 1.1.2) and to refine the bog map. Areas mapped as bog but not occurring on peat were reassigned as wet heath. The update has resulted in some changes to the habitat categories used:

- a) Managed and unmanaged broadleaved woodland are no longer distinguished, because the National Forest Inventory (GB) and Agri-Food and Biosciences Institute woodland register (Northern Ireland) no longer map these separately. However, even commercial forestry in the

UK is increasingly managed for multiple objectives. Exceedance of critical loads is likely to limit nature conservation value of any woodland, so it was considered appropriate to include all broadleaved woodland within the sensitive area.

- b) Broadleaved woodland was mapped as priority habitat if it occurred within 10 km x 10 km grid cells that according to the National Vegetation Classification (NVC) dataset contain H9190 or H91A0 woodland types (i.e. acidophilous oak woodland); or W12, W14 or W15 woodland types (i.e. beech woodland). For NVC squares that contained both beech woodland and oak woodland, it was assumed that 50 % of the Broadleaved woodland area was oak woodland and 50 % was beech woodland
- c) Coniferous woodland in areas where scots pine is native was considered to be unmanaged. All coniferous woodland within 10 km x 10 km grid cells in Scotland that according to the National Vegetation Classification (NVC) dataset contain W18 (*Pinus sylvestris–Hylocomium splendens* woodland) was mapped as unmanaged priority-habitat woodland, and assessed for exceedance of nutrient-N critical load separately from managed coniferous woodland. This approach presumably overestimates the area of scots pine woodland, but coniferous woodland within the native range of scots pine has the potential to become priority habitat and was treated as such. All coniferous woodland in grid cells where W18 does not occur was considered to be managed. Different critical loads for nutrient N were assigned to managed coniferous woodland and scots pine woodland. The same approach to calculating acidity critical load was used for both types of coniferous woodland.
- d) Wet and dry heath are no longer mapped separately, since they have the same critical loads for acidity and for nutrient-N.

The changes to habitat mapping have increased the total mapped area of habitats sensitive to nutrient-N in the UK by 28 % (Figure 1.1).

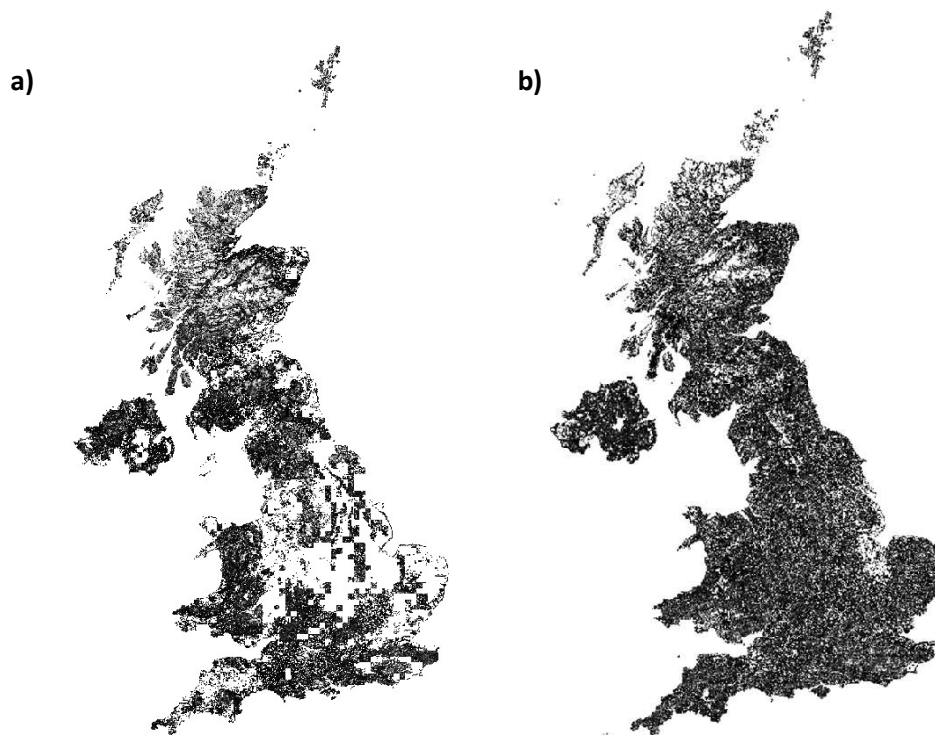


Figure 1.1. Areas of habitat sensitive to nutrient-nitrogen in the UK, as mapped using a) mainly LCM2000 data, as used for Trends Reports until 2021; and b) mainly LCM2019 data, as used for the current report.

It should be noted that the habitat distribution maps and areas used to calculate exceedances: a) only include areas where data exist for the calculation or derivation of critical loads; and b) may differ from other national habitat distribution maps or estimates of habitat areas.

Published correspondence tables (available from: <https://hub.jncc.gov.uk/assets/9e70531b-5467-4136-88f6-3b3dd905b56d>) are used to relate broad habitats to the European Nature Information System (EUNIS: Moss and Davies, 2002) hierarchical habitat classification scheme, developed for pan-European applications. The codes used for EUNIS habitats are those outlined in 2012, and will be updated to use the 2021 codes in a future report.

Table 1.1: Habitat distributions mapped for acidity and for nutrient nitrogen critical loads. See section 1.4 for definition of N_{sens} .

Habitat	EUNIS habitat class(es) assigned ¹	Mapped for acidity	Mapped for nutrient-N	Included in N_{sens} calculation
Acid grassland	R372 & R1M (E1.7 & E3.52)	Yes	Yes	Yes
Calcareous grassland	R1A (E1.26)	Yes	Yes	Yes
Dwarf shrub heath	S411 & S42 (F4.11 & F4.2)	Yes	Yes	Yes
Montane	E4.2 ³	Yes	Yes	Yes
Bog	Q1 (D1)	Yes	Yes	Yes
Scots Pine woodland	T35 (G3.4)	Yes	Yes	Yes
Managed coniferous woodland	T31 (G3)	Yes	Yes	No
Beech woodland	T17 (G1.6)	Yes	Yes	Yes
Acidophilous oak woodland	T1B (G1.8)	Yes	Yes	Yes
Other broadleaved woodland	T1 (G1)	Yes	Yes	Yes
Mixed woodland	G4 ⁴	Yes	Yes	Yes
Freshwaters ²	C1 & C2	Yes	No	No
Dune grassland	N15 (B1.4)	No	Yes	Yes
Saltmarsh	MA223/MA224/MA225 (A2.53/54/55)	No	Yes	Yes

¹ EUNIS class (2023 EUNIS codes, with 2012 EUNIS codes in brackets) closest to broad habitat and critical loads habitat; class used for assigning empirical nutrient nitrogen critical loads and for classifying UK critical loads data for submission to the CCE.

² Critical loads are calculated for 1752 freshwater sites across the UK (see Section 1.1.2); habitat areas are based on the catchment areas of these sites.

³ The 2023 revision of EUNIS codes does not include a class for montane habitats (formerly moss summits). The revision notes that the most similar new class is R42 (boreal and arctic acidophilous alpine grassland) but does not give a critical load for this class, hence the critical load for E4.2 has been retained.

⁴ Mixed woodland is not distinguished in the 2023 revision of EUNIS codes.

1.1.2 Acidity critical loads

Two methods are used in the UK for calculating acidity critical loads for terrestrial habitats: the empirical approach is used to provide estimates for non-woodland habitats on mineral and organomineral soils; and simple-mass-balance equations are used for woodland habitats and for habitats that occur on peat. The CLmaxS, CLmaxN and CLminN values that define the critical load function (see Figure 1.3) are derived from the acidity critical load value, taking into account the deposition of anthropogenic base cations and chloride, and inputs and losses of nitrogen (Hall et al., 2015).

In the empirical approach used for non-woodland habitats on mineral and organo-mineral soils, critical loads for acidity are assigned to each 1 km x 1 km grid square of the UK based on the amount of acid deposition that could be neutralised by the base cations produced by mineral weathering of the dominant soil type in the grid square (Hornung et al., 1995).

For woodland habitats a simple mass balance (SMB) equation, based on balancing the acidic inputs to and outputs from the ecosystem, is used to derive a critical load that ensures that a specified critical chemical limit is not exceeded (Sverdrup and De Vries, 1994; Sverdrup et al., 1990). In the UK, the limit used is a calcium to aluminium molar ratio of 1.0 (Hall et al., 2015). Below this value, aluminium has significant toxic effects. Critical loads are calculated for both managed (productive) and unmanaged woodlands to protect the long-term ecosystem function of these woodland habitats; this approach also aims to protect the land under managed coniferous forest for possible future non-forest use and reversion to semi-natural land uses.

A simpler mass balance approach is used for habitats on acid peat soils, because of the absence of inputs of alkalinity and of aluminium from mineral weathering (Gammack et al., 1995; Smith et al., 1992). Critical loads of acidity for peat soils are set to the value corresponding to the amount of acid deposition that would give rise to an effective rain pH value of 4.4 (Calver, 2003; Calver et al., 2004; Skiba and Cresser, 1989); the choice of threshold pH value reflects the buffering effects of organic acids on the pH of peat drainage water. This method is applicable to acid peat soils, but not to peats in lowland fen areas that receive mineral inputs from groundwater and are less sensitive to acidification, where a relatively high critical load of 4.0 keq ha⁻¹ yr⁻¹ is applied (Hall et al., 2015).

Acidity critical loads for freshwaters are calculated using the catchment-based First-Order Acidity Balance (FAB: Henriksen and Posch, 2001) model. FAB is currently applied to 1752 sites across the UK, comprising a mixture of mainly upland, lakes, reservoirs and first-order streams (i.e. streams that feed into other larger streams, but do not have any other streams draining into them). The critical load calculations are based on the water chemistry of samples collected in the 1990s to provide an estimate of the annual mean water chemistry. The FAB model generates the acidity critical load values CL_{maxS}, CL_{minN} and CL_{maxN} (see Part 2).

1.1.3 Nutrient nitrogen critical loads

Both empirical and mass-balance methods can be used for calculating critical loads for eutrophication (i.e. an excess of nutrients, in this case N). Empirical critical loads are based on experimental or field evidence of thresholds for changes in species composition, plant growth, plant tissue chemistry or soil processes. The empirical approach is suited to semi-natural communities for which the long-term protection of biodiversity and/or ecosystem function is the key concern. In the UK, the empirical approach is applied to natural and semi-natural habitats, based on critical load values agreed at international workshops (Bobbink and Hettelingh, 2011; Hall et al., 2015). For Trends Reports up until 2022, values agreed in 2011 (Bobbink and Hettelingh, 2011) were used. Empirical critical loads for N (CL_{empN}) are presented as a range for each habitat, to represent the variation in sensitivity within the habitat, and also uncertainty in both ecosystem response and N addition rate in the studies that

were considered. For Trends Reports, exceedances are calculated in relation to a single value within each range, the “mapping value”. For Trends Reports up until 2022, the middle of the range was used as the mapping value for most habitats.

A new review and revision of empirical critical loads for N (CLEmpN) was published in 2022 (Bobbink et al., 2022) and was used as the basis for exceedances calculated in the current report. New evidence resulted in changes (mainly decreases) to the upper and lower ends of the ranges for many habitats. An expert group was convened in autumn 2022 to consider what mapping values should be used, and concluded that it is most appropriate to apply the lower end of the range for all habitats, with no modifiers such as soil pH or annual precipitation. The discussion is described in Rowe & Hina (2023). The new CLEmpN mapping values have decreased considerably, resulting in substantial increases in the exceedance statistics compared to the 2022 Trends Report. In the current report, results for previous years have been updated, and exceedances are calculated using the same mapping values (and other methods) throughout.

Empirical critical loads have been applied to all habitats apart from managed coniferous woodland, to which a mass-balance approach was applied. This approach takes into account the long-term inputs and outputs of N from the ecosystem, with the critical load being exceeded when a specified critical rate of N leaching is exceeded. This approach is suitable for managed ecosystems with relatively low biodiversity, in which the inputs and outputs can be quantified with some confidence and in which a key concern is nitrate leaching. As with acidity, in the UK this approach is applied to managed coniferous woodlands to ensure that long-term ecosystem functions (e.g. of soils, soil biological resources, trees, or linked aquatic systems) are protected.

1.2 Deposition data and trends

The S, N and base cation deposition data used in the UK calculations of critical loads and their exceedances are based on the “Concentration Based Estimated Deposition” (CBED) methodology (RoTAP, 2012). Site-based measurements of air concentrations of sulphur and nitrogen, as gases and particulates (Tang et al., 2018a; Tang et al., 2018b), are interpolated to generate 5 km x 5 km maps of concentrations for the UK. Ion concentrations in precipitation from the UK Eutrophying and Acidifying Pollutants (UKEAP) network (Braban et al., 2021; Conolly et al., 2018) are combined with the Met Office precipitation map to generate maps of wet deposition. The wet deposition values include direct deposition of cloud droplets to vegetation (known as “occult” deposition), and orographic enhancement to take account of the “seeder-feeder” effect in upland regions (Fowler et al., 1988). A map of gaseous ammonia concentration is combined with spatially distributed estimates of vegetation-specific deposition velocities (Smith et al., 2000) to generate dry deposition. The spatial distribution of gaseous ammonia was derived from the UK version of the European Monitoring and Evaluation Programme (EMEP4UK) model, driven by the most recent emissions and meteorology data available, as discussed in Section 4.2. Combining these data sets produces 5 km x 5 km maps of total (wet + cloud + dry) deposition of S (non-marine), oxidised N (NO_y) and reduced N (NH_x). Two different sets of deposition values are calculated for critical load and exceedance applications: i) deposition onto “moorland”, i.e. unfertilised open (non-wooded) habitats; and ii) deposition onto “woodland”, which has greater deposition velocities due to the greater roughness of the land surface.

Significant inter-annual variations in deposition can occur due to the natural variability in annual precipitation (which influences wet deposition) as well as the general circulation of air which can increase or decrease the amount of polluted air imported from the European continent. The CBED deposition data used to calculate critical load exceedances is therefore averaged over a three-year period; this has been demonstrated to be a suitable time period to smooth out inter-annual variations in deposition. Figure 1.1 shows the CBED data for 2018-20.

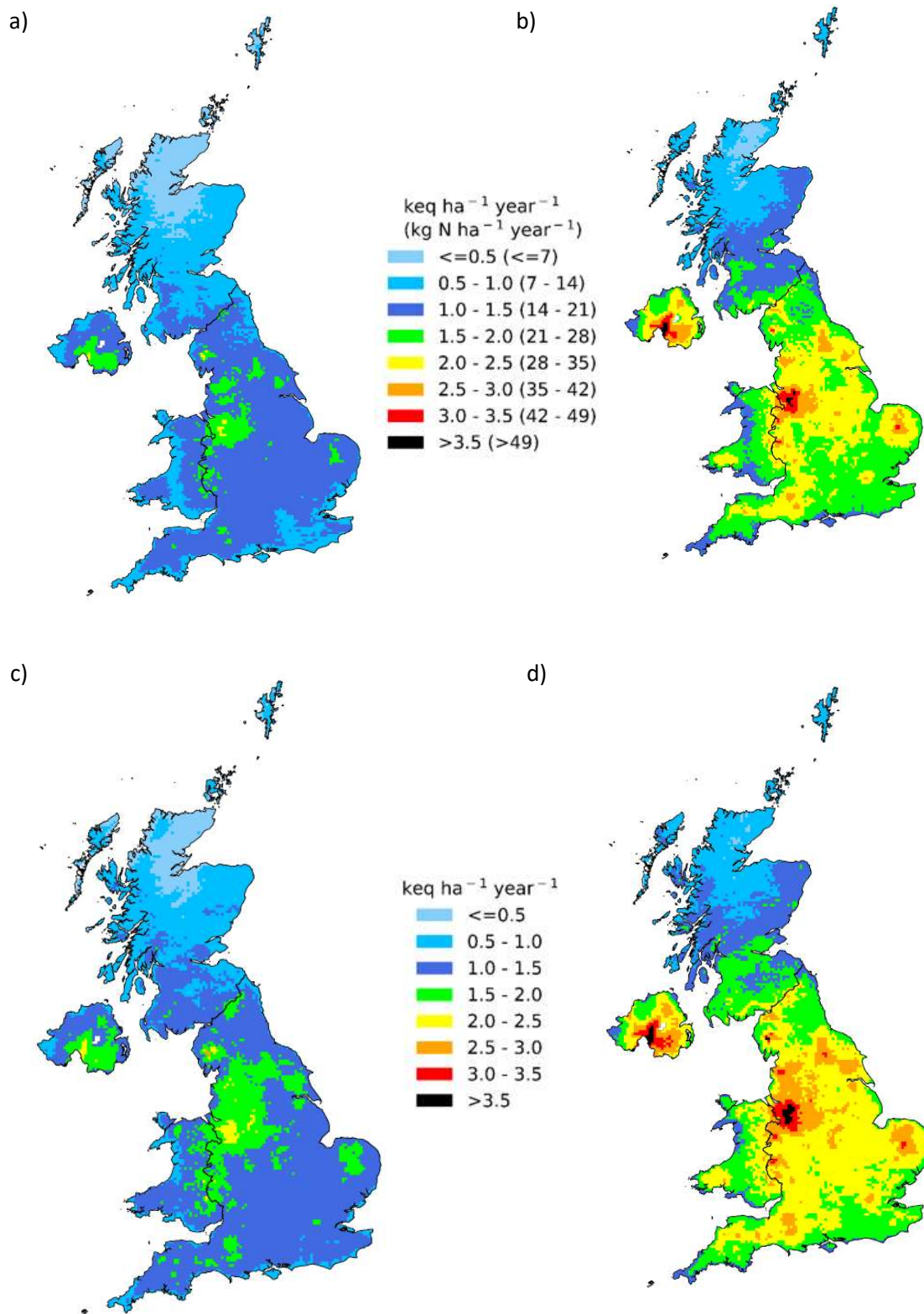


Figure 1.2: CBED deposition for 2019-2021: a) nitrogen (oxidised plus reduced) deposition to moorland; b) nitrogen (oxidised plus reduced) deposition to woodland; c) acid (sulphur + nitrogen) deposition to moorland; d) acid (sulphur + nitrogen) deposition to woodland. Deposition is mapped for N and acidity using the same units ($\text{keq ha}^{-1} \text{ yr}^{-1}$) and class intervals. For the N maps, deposition rates are also shown in $\text{kg N ha}^{-1} \text{ yr}^{-1}$.

Since critical loads for terrestrial habitats are mapped on a 1 km grid, for exceedance calculations deposition is assumed to be constant for all 1 km squares within each 5 km square. For freshwater exceedance calculations catchment-weighted mean S and N deposition values are calculated by overlaying land cover (moorland vs. forest) and catchment boundaries onto the 5 km deposition maps.

1.2.1 Summary of trends in CBED deposition

To understand trends in critical load exceedances, it is useful to look at trends in deposition. Deposition estimates using CBED deposition data derived from the EMEP map for ammonia exist for all 3-year rolling intervals from 2003 to 2020 (reminder: the 3-year time periods are referred to in brief using the *middle year* of the period). The trends in CBED deposition to moorland and to woodland are summarised in Figure 1.3. Deposition onto woodland is higher than that onto moorland due to greater deposition velocities of gases (e.g. NH₃ and HNO₃), as well as particulates, onto tall vegetation. Between 2003 and 2020, non-marine S deposition onto moorland decreased by 63.8 % and N deposition onto moorland decreased by 12.6 %. There was a small increase between 2003 and 2010 in NH_y deposition onto moorland at 1.4 %, but the latest data indicate a recent decrease.

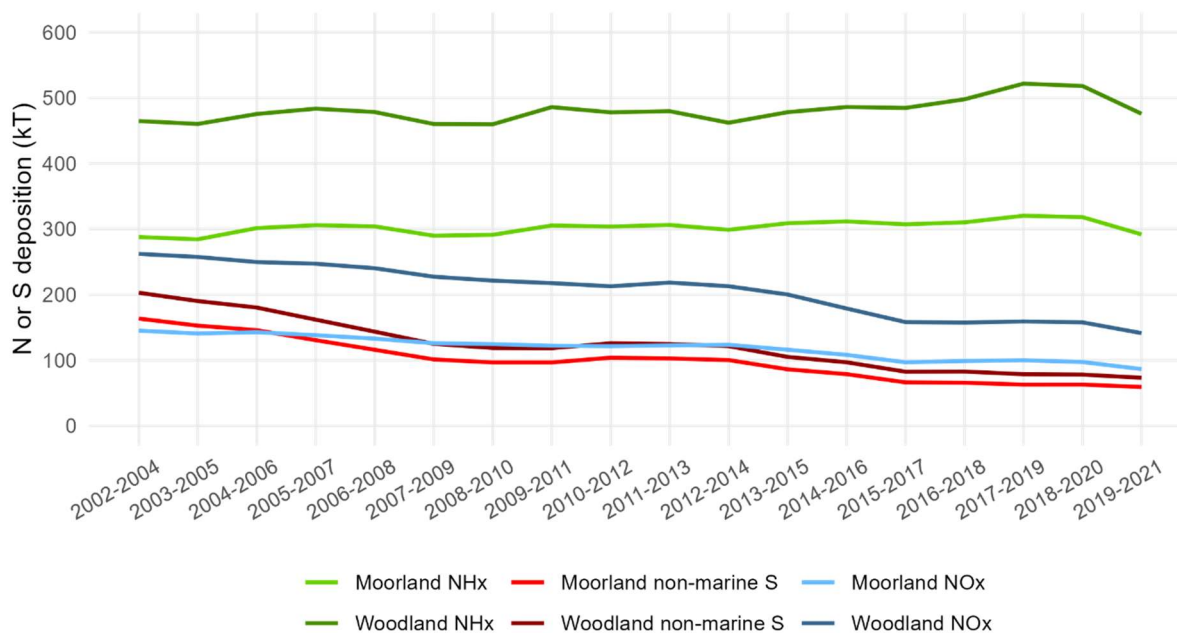


Figure 1.3: Deposition budgets (kilotonnes S or N per year) calculated by CBED using ammonia concentration maps from EMEP assuming UK area is a) woodland or b) moorland or other unfertilised open habitat. NMS = non-marine sulphur, NOx = oxidised nitrogen, NHx = reduced nitrogen.

1.3 Overview of the calculation of critical load exceedances

Critical load exceedances are the amount of excess deposition above the critical load; for nutrient N the calculation is simply total N deposition (derived from N oxides and ammonia) minus the critical

load. For acidification, deposition of both S and N compounds can contribute to the exceedance of critical loads. The Critical Load Function, developed under the UNECE CLRTAP (Hettelingh et al., 1995; Posch et al., 1999; Posch et al., 1995; Posch and Hettelingh, 1997), defines combinations of S and N deposition that will not cause harmful effects. In its simplest form, an acidity critical load can be defined graphically by a 45 degree diagonal line on a sulphur-nitrogen deposition plot (Figure 1.3a), where both types of deposition are expressed in chemical equivalents, i.e. moles of charge. The line intercepts the x-axis (representing N deposition) and y-axis (representing S deposition) at the same value in equivalents, each representing the N or S deposition equal to the critical load for acidity. Each point along the diagonal line represents the critical load in terms of some combination of S and N deposition.

To allow for the long-term N removal processes by the soil and through harvesting of vegetation, the simple diagonal line is shifted along the N axis to increase the N values across the entire Critical Load Function (Figure 1.3b). More N can then be deposited before the acidity critical load is exceeded. There are no similar removal processes that need to be considered for S.

The intercepts of the Critical Load Function on the S and N axes (Figure 1.3c) define the following terms:

1. The “maximum critical load of S” (CL_{maxS}): the critical load for acidity expressed in terms of S only, i.e. when N deposition is zero.
2. The “maximum critical load of N” (CL_{maxN}): the critical load for acidity expressed in terms of N only (when S deposition is zero).
3. The “minimum critical load of N” (CL_{minN}): represents long-term N removal processes in the soil (e.g. N uptake and immobilisation) and harvesting of vegetation.

These critical loads are calculated from the acidity critical loads described in Section 1.1 and additional soil-specific or habitat-specific data.

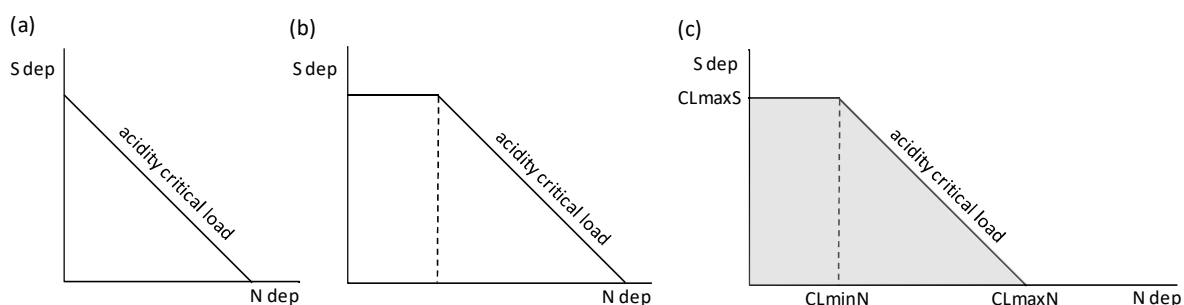


Figure 1.4: Development of the Critical Load Function: a) acidity critical load defined by equal amounts of sulphur and nitrogen deposition; b) shifting the acidity critical load diagonal line to allow for nitrogen removal processes; c) the 3 intercepts of the Critical Load Function: CL_{maxS}, CL_{minN} and CL_{maxN}. The area shown in grey represents the combinations of sulphur and nitrogen deposition that are below the critical load (i.e. critical load is not exceeded).

Exceedances are calculated by comparing the values of CL_{maxS}, CL_{minN} and CL_{maxN} to the values of S and N (oxidised + reduced) deposition. The actual calculation depends on where the deposition falls in relation to these critical load values; the Critical Load Function is divided into five different regions for this purpose (Figure 1.4). The exceedance is defined by the sum of S and N deposition as shown by

the red arrows in Figure 1.4 (i.e. not the length of the diagonal line); this is referred to as the “shortest distance” exceedance. Further details of the calculations are given in Hall et al. (2015).

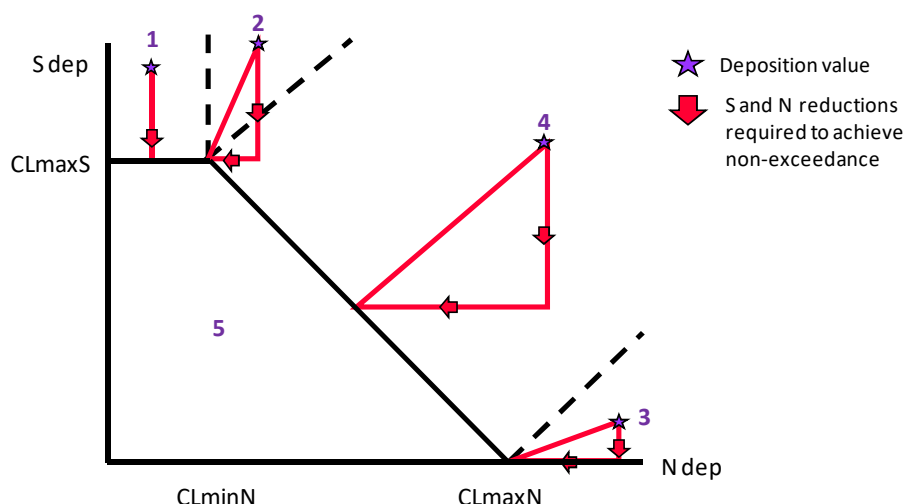


Figure 1.5: Example of S and N deposition reductions required depending on the region of the CLF. Deposition that falls in region 5 is below the critical load (i.e. critical loads not exceeded).

1.3.1 Exceedance and damage

The critical loads data on which exceedance calculations are based are derived from empirical or steady-state mass balance methods, which are used to define critical loads for the long term. Exceedance of critical loads is an indication that an ecosystem is at risk from potential harmful effects in the long-term. Therefore, exceedance is not a quantitative estimate of damage to the environment; it does not necessarily mean that harmful or adverse effects have already occurred or may be observed, but that there is a risk of damage in the long-term. Reducing deposition to below the critical load does not mean that ecosystems immediately recover. There are time lags before chemical recovery takes place, and further delays before biological recovery. Timescales for both chemical and biological recovery could be very long, particularly for the most sensitive ecosystems.

1.3.2 Critical load exceedance metrics

Critical load exceedances are calculated for each 1km square of the distributions of each terrestrial habitat, and for each catchment for freshwaters. The results are then summarised by habitat and country using the following exceedance metrics:

- (i) *Area of habitat exceeded*

For terrestrial habitats, the area values are based on habitat maps mainly derived from LCM2019 data; if the critical load for any individual habitat is exceeded, the exceeded area is set to the habitat area within the 1km square for that particular habitat. For freshwater habitats, if the FAB acidity critical load is exceeded, the whole catchment is assumed to be exceeded and the exceeded area set to the catchment area. The total exceeded areas for individual habitats are summarised by country. Freshwater catchment areas are not included when calculating total area exceeded for acidity per country, to avoid double-counting of the areas of sensitive habitat within these catchments.

(ii) *Percentage area of habitat exceeded*

This is calculated from the exceeded areas derived in (i) and the total area of each habitat mapped in each country (Section 1.1). While this is a useful metric for expressing how much habitat is at risk, it does not clearly reflect decreases in the amount of exceedance. For example, when comparing exceedance results from one year to another (or one deposition scenario to another), there may be only small changes in the percentage area of habitat exceeded, particularly in regions with high deposition. This is because the magnitude of the exceedance may have reduced, but the area exceeding the critical load remains the same; the area exceeded will only reduce when the critical load is no longer exceeded. Even when the critical load is still exceeded, decreases in the amount of exceedance are likely to have some benefits (Rowe et al., 2017).

(iii) *Accumulated Exceedance (AE)*

AE takes account of both the magnitude of exceedance and the habitat area exceeded:

$$AE \text{ (keq year}^{-1}\text{)} = \text{exceedance (keq ha}^{-1} \text{ year}^{-1}\text{)} \times \text{exceeded area (ha)}$$

AE is calculated for each 1km square for each habitat and then summarised by habitat and country. AE is set to zero where critical loads are not exceeded. This metric can be useful for comparing results for different years or scenarios, but because the results are expressed in as totals for the country (in keq year⁻¹) they are large numbers and not intuitive to understand. It should also be noted that the same AE can arise from a large exceedance and small exceeded area, or a small exceedance and a large area.

(iv) *Average Accumulated Exceedance (AAE)*

AAE averages the AE across the entire sensitive habitat area:

$$AAE \text{ (keq ha}^{-1} \text{ year}^{-1}\text{)} = AE \text{ (keq year}^{-1}\text{)} / \text{total sensitive habitat area (ha)}$$

This metric provides an exceedance value averaged across the whole habitat area. In this report we use the terms **Excess Nitrogen** as a synonym for the AAE of nutrient-N critical loads and **Excess Acidity** as a synonym for the AAE of acidity critical loads. In the summary statistics presented (Section 2) it is based on the AE for the habitat (by country) divided by the total sensitive habitat area (by country). AAE is set to zero where critical loads are not exceeded. This metric provides a more intuitive value for comparing the exceedance results for different years or scenarios, and gives an indication of the reduction in the magnitude of exceedance even if there is no change in the percentage area of habitat exceeded.

1.3.3 Overall maps of critical load exceedance

Critical load exceedances are calculated by habitat, which can make it hard to interpret maps of exceedance. The AAE (Excess Nitrogen or Excess Acidity) is an area-weighted measure of exceedances across habitats (Figure 1.5). The AAE for each 1km square is calculated as:

$$AAE = \frac{\sum \text{AE for all habitats}}{\sum \text{Area for all habitats}}$$

Where AE is the area exceeded. AE and AAE are set to zero where the critical loads are not exceeded. The latest Excess Nitrogen and Excess Acidity maps (Figure 1.5) show lower exceedances in Scotland. The greatest Excess Acidity values are seen in upland areas of central and NW England, and smaller areas in E and SW England, parts of Wales, southern Scotland, and Northern Ireland. Exceedances of nutrient N critical loads are seen across most of England, Wales and Northern Ireland, with many areas having exceedances above 14 kg N ha⁻¹ year⁻¹ (1 keq ha⁻¹ year⁻¹).

1.4 Calculation of N deposition onto protected sensitive habitats in England

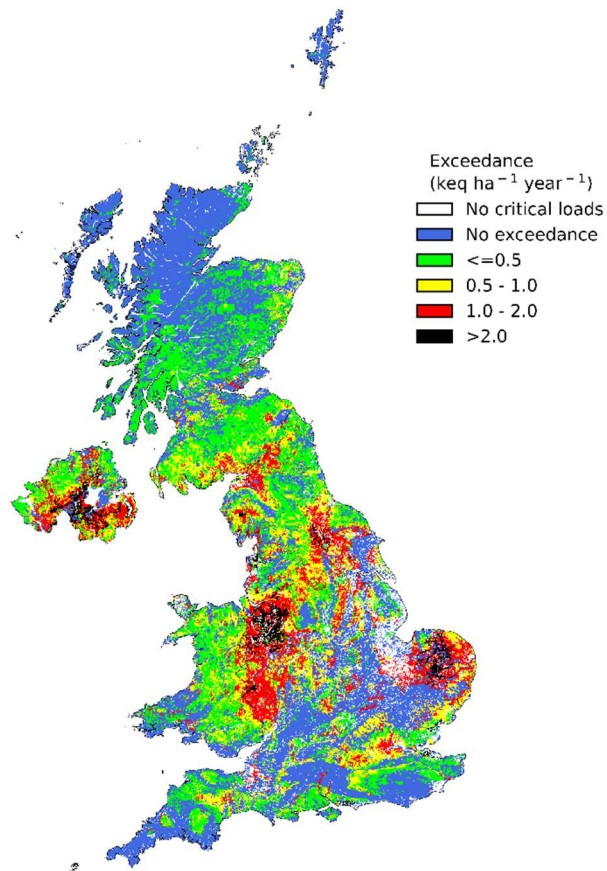
The Defra Clean Air Strategy (Defra, 2019) includes an indicator for England, “total deposition of reactive N onto nutrient-N sensitive, protected, priority habitats” (N_{sens}). This metric is calculated as the area-weighted mean deposition (sum of oxidised and reduced N) in kg N ha⁻¹ yr⁻¹ (Equation 1).

$$N_{sens} = \frac{\sum_{i=1}^n A_i D_i}{\sum_{i=1}^n A_i} \quad (\text{Equation 1})$$

Where n is the number of habitats included in the calculation, A is the total area of each habitat, and D is the total N deposition onto each habitat.

The priority habitats included in this calculation are a subset of the habitats listed in Table 1.1, as indicated in that table. Deciduous woodland and mixed woodland were included, but not managed coniferous woodland (see Section 1.1).

(a) Acidity



(b) Nutrient nitrogen

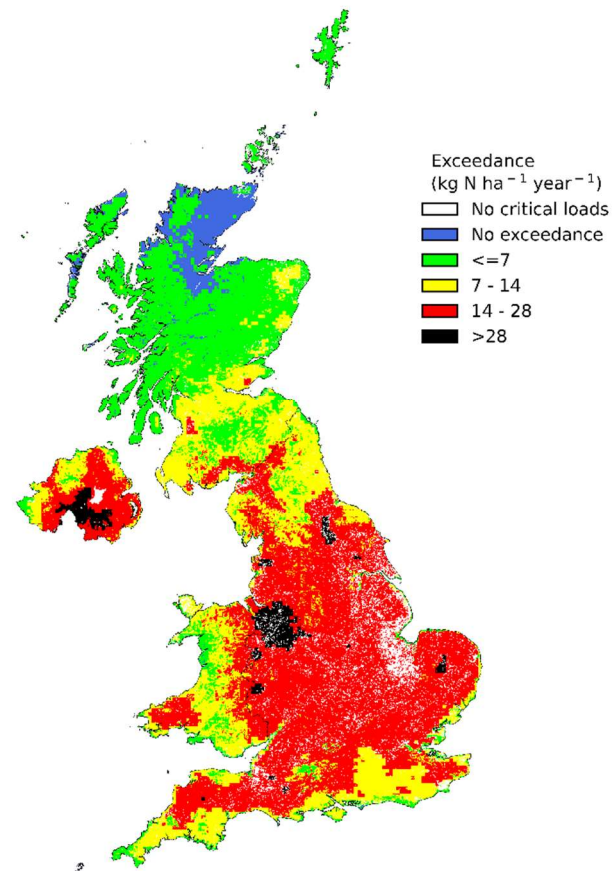


Figure 1.6: a) Excess Acidity (Average Accumulated Exceedance of acidity critical load) and b) Excess Nitrogen (Average Accumulated Exceedance of nutrient-nitrogen critical load) in 2020. Although the legends for the two maps are given in different units, the class intervals are the same (e.g. $7 \text{ kg N ha}^{-1} \text{ year}^{-1}$ is equal to $0.5 \text{ keq ha}^{-1} \text{ year}^{-1}$).

Section 2: Trends in critical loads exceedance by habitat and country

The data used for the trends analysis are described in Section 1. Acidity and nutrient N exceedances by habitat and country are updated annually using the latest three-year rolling mean deposition data. The summary statistics as described in Section 1.3.1 are made available to Defra and the Devolved Administrations and JNCC. Trends in the percentage area of habitats exceeded are also included in the annual summary of UK Biodiversity Indicators (<https://jncc.gov.uk/our-work/ukbi-b5a-air-pollution>).

The trends results are shown as both tables and simple plots. It is worth noting that the percentage area exceeded for some habitats changes only gradually from one year to another, whereas AAE values (Excess Nitrogen and Excess Acidity) are more responsive to changes in deposition.

2.1 Trends by country

Table 2.1 shows the total land area by country and the area of habitats sensitive to acidification and eutrophication to which critical loads have been applied; 36% of the UK land area has habitats mapped for acidity critical loads, and 38% for nutrient N. Freshwater habitats are also mapped for acidity, but statistics are reported separately.

Table 2.1: Total land area and terrestrial habitat areas mapped for critical loads by country.

Country	Land area (km ²) [#]	Terrestrial habitat areas mapped for acidity (km ²)	Area mapped for acidity, as % of country	Terrestrial habitat areas mapped for nutrient nitrogen (km ²)	Area mapped for nutrient nitrogen, as % of country
England	132,938	23,429	18	26,038	20
Wales	21,225	9,075	43	9,412	44
Scotland	80,239	54,512	68	54,333	68
NI	14,130	3,527	25	3,990	28
UK	248,532	90,543	36	93,774	38

[#]The UK and its countries: facts and figures. Office for National Statistics.

2.1.1 Acidity results

The results for acidity (Table 2.2, Figure 2.1) show that the total percentage area of terrestrial habitats exceeding critical loads in the UK declined by 21.2%, from 65.6% in 2003 to 44.5% in 2020. However, the area exceeded varies between countries (Table 2.2, Figure 2.2), due to: a) geographic location of different sensitive habitats across the country (see Section 2.2); b) variability in critical load values across the country – lower critical loads associated with habitats on more acid soils; and c) higher deposition found in central and south-west England, parts of Wales and Northern Ireland and south-west Scotland (Figure 1.1). The percentage area of habitats exceeded is lowest in Scotland in all years;

however as shown in Table 2.1, 68% of Scotland has habitats mapped for acidity critical loads, and that means the actual areas exceeded are larger than in the other countries (e.g. 15,753 km² exceeded by 2020 deposition). Although only 18% of England has habitats mapped for acidity critical loads, 66.5% of their area is exceeded for 2020, equivalent to 15,569 km². The magnitude of exceedance across the UK, expressed as Excess Acidity (Table 2.3, Figure 2.1), has fallen from 0.58 keq ha⁻¹ year⁻¹ in 2003 to 0.25 keq ha⁻¹ year⁻¹ in 2020 .

Table 2.2: Acid-sensitive terrestrial habitat area and percentage and total area of habitats where acidity critical loads are exceeded, by country and deposition dataset year. Areas include freshwater catchments, which are likely to include overlaps with terrestrial acid-sensitive habitats.

Years	England	Wales	Scotland	NI	UK
	<i>Total area (km²) acid-sensitive habitats</i>				
	<i>23429</i>	<i>9075</i>	<i>54512</i>	<i>3527</i>	<i>90543</i>
2002-2004	77.3 [18,112]	83.8 [7,601]	56.7 [30,896]	79.3 [2,797]	65.6 [59,407]
2003-2005	76.9 [18,016]	83.4 [7,571]	54.4 [29,642]	77.8 [2,744]	64.0 [57,972]
2004-2006	76.6 [17,945]	84.5 [7,672]	63.8 [34,761]	79.3 [2,798]	69.8 [63,176]
2005-2007	76.2 [17,849]	85.2 [7,735]	63.0 [34,343]	79.6 [2,808]	69.3 [62,736]
2006-2008	75.2 [17,611]	84.4 [7,659]	59.3 [32,314]	79.7 [2,813]	66.7 [60,397]
2007-2009	73.2 [17,148]	79.5 [7,215]	47.0 [25,609]	78.5 [2,768]	58.2 [52,740]
2008-2010	72.6 [17,006]	77.0 [6,991]	46.9 [25,569]	79.3 [2,795]	57.8 [52,361]
2009-2011	73.2 [17,160]	77.8 [7,064]	50.2 [27,391]	80.6 [2,843]	60.1 [54,458]
2010-2012	73.3 [17,179]	77.8 [7,064]	48.6 [26,503]	80.3 [2,831]	59.2 [53,577]
2011-2013	73.8 [17,292]	79.5 [7,215]	51.1 [27,860]	81.5 [2,875]	61.0 [55,242]
2012-2014	72.7 [17,023]	77.2 [7,002]	50.3 [27,422]	80.5 [2,840]	60.0 [54,287]
2013-2015	72.1 [16,901]	77.5 [7,031]	53.3 [29,080]	79.1 [2,789]	61.6 [55,801]
2014-2016	71.0 [16,631]	77.3 [7,013]	50.3 [27,395]	77.1 [2,718]	59.4 [53,758]
2015-2017	68.8 [16,109]	75.1 [6,819]	42.8 [23,338]	74.6 [2,633]	54.0 [48,897]
2016-2018	69.3 [16,225]	74.4 [6,755]	35.6 [19,380]	77.1 [2,719]	49.8 [45,079]
2017-2019	69.6 [16,302]	74.1 [6,723]	34.6 [18,876]	78.1 [2,753]	49.3 [44,653]
2018-2020	69.9 [16,378]	73.6 [6,681]	33.7 [18,351]	77.5 [2,732]	48.8 [44,142]
2019-2021	66.5 [15,569]	70.1 [6,364]	28.9 [15,753]	72.8 [2,566]	44.5 [40,253]
Reduction 2003-2020	10.9 [2,542]	13.6 [1,238]	27.8 [15,143]	6.5 [231]	21.2 [19,154]

Table 2.3: Excess Acidity (Average Accumulated Exceedance in $\text{keq ha}^{-1} \text{ year}^{-1}$) by country and deposition dataset year.

Years	England	Wales	Scotland	NI	UK
2002-2004	1.04	0.82	0.34	0.69	0.58
2003-2005	1.00	0.78	0.31	0.63	0.55
2004-2006	0.98	0.87	0.41	0.68	0.61
2005-2007	0.92	0.87	0.37	0.70	0.57
2006-2008	0.86	0.82	0.32	0.69	0.53
2007-2009	0.78	0.65	0.22	0.65	0.42
2008-2010	0.75	0.59	0.22	0.68	0.41
2009-2011	0.79	0.58	0.27	0.75	0.45
2010-2012	0.79	0.61	0.26	0.72	0.45
2011-2013	0.81	0.67	0.29	0.76	0.48
2012-2014	0.76	0.65	0.27	0.71	0.45
2013-2015	0.75	0.67	0.30	0.65	0.47
2014-2016	0.70	0.68	0.26	0.59	0.43
2015-2017	0.61	0.58	0.19	0.53	0.35
2016-2018	0.61	0.51	0.15	0.61	0.32
2017-2019	0.62	0.46	0.14	0.68	0.32
2018-2020	0.62	0.47	0.13	0.68	0.31
2019-2021	0.50	0.38	0.10	0.55	0.25
Reduction 2003-2020	0.54	0.44	0.24	0.14	0.34

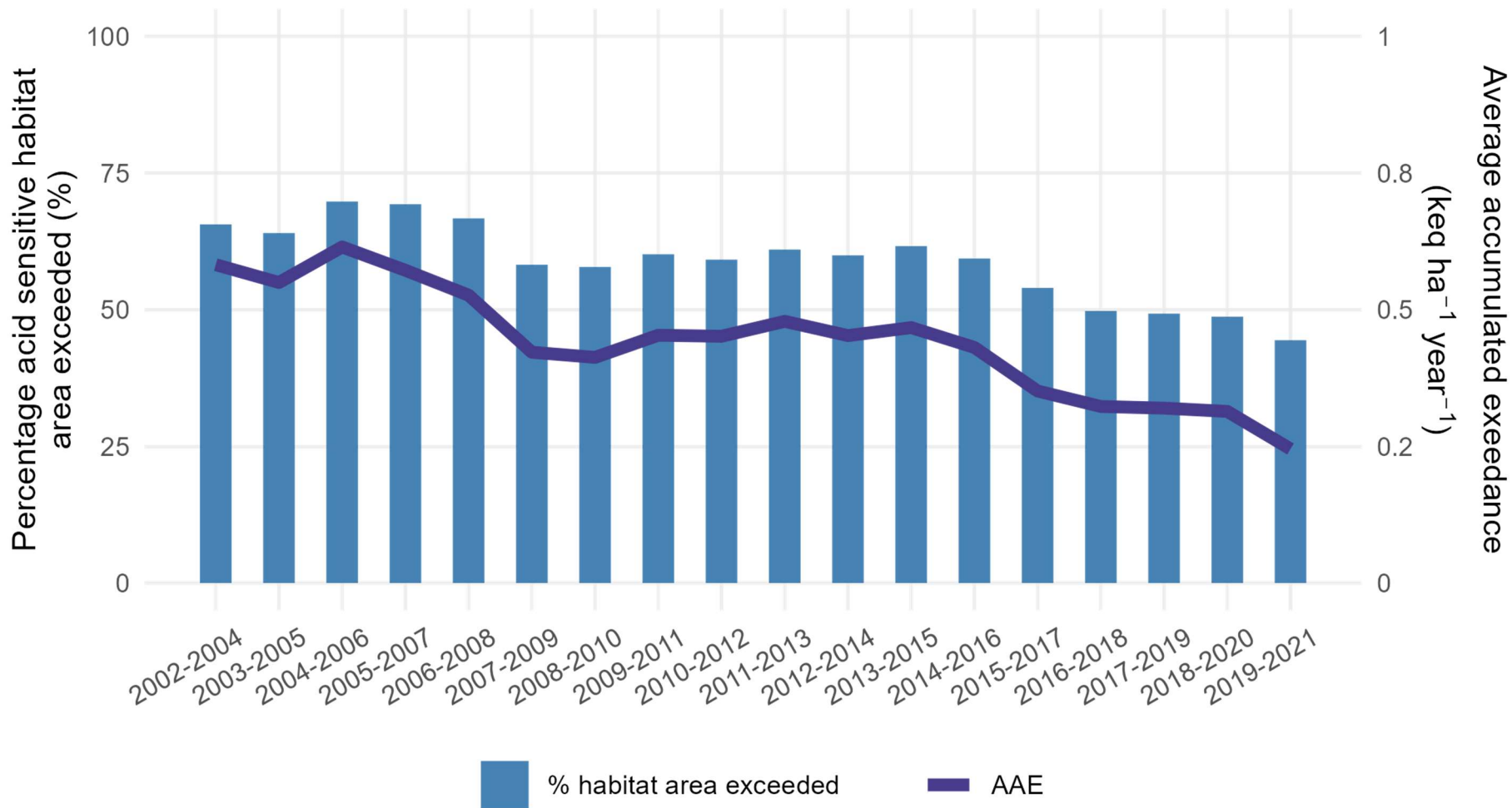


Figure 2.1: Acidity: Percentage area of acid-sensitive habitats with exceedance of acidity critical loads in the UK by year, and Average Accumulated Exceedance in keq ha⁻¹ year⁻¹.

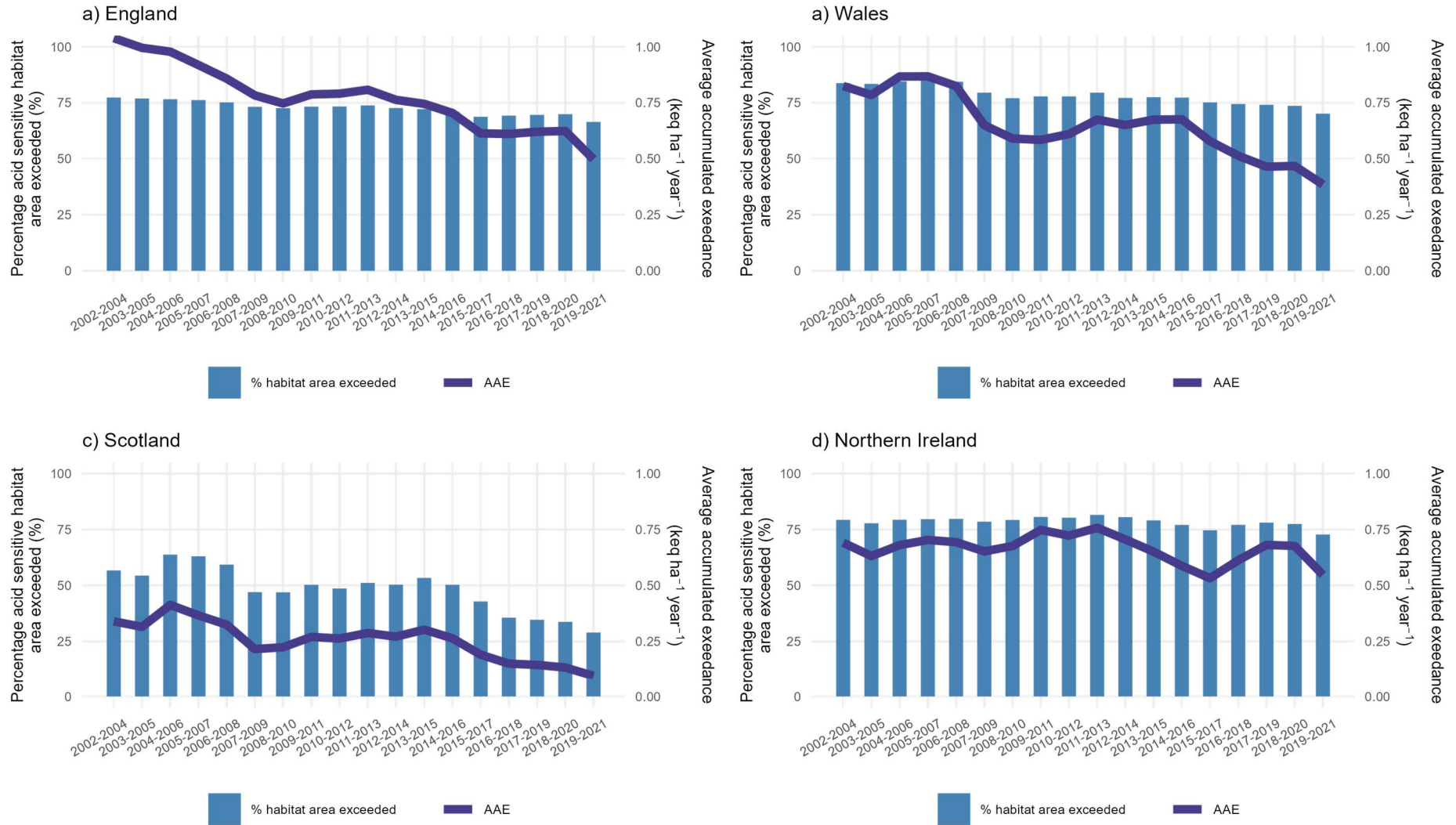


Figure 2.2: Acidity: Percentage area of acid-sensitive habitats with exceedance of acidity critical loads, by country and year, and Excess Acidity (Average Accumulated Exceedance in keq ha⁻¹ year⁻¹).

2.1.2 Nutrient nitrogen results

The results for nutrient N (Table 2.4 and Figure 2.3) show a decline in the percentage area of habitats exceeded in the UK, from 93.8% in 2003 to 85.9% in 2020. The results for England, Wales and Northern Ireland remain close to or exactly 100% exceeded over the same time period (Table 2.4, Figure 2.4). Scotland shows the smallest percentage habitat area exceeded of all countries (75.8%).

Table 2.4: Nitrogen-sensitive habitat area and percentage and total area of habitats where nutrient nitrogen critical loads are exceeded, by country and deposition dataset year.

Years	England	Wales	Scotland	NI	UK
	<i>Total area (km²) nutrient-N-sensitive habitats</i>				
	26038	9412	54333	3990	93774
2002-2004	100.0 [26,034]	99.9 [9,401]	89.3 [48,529]	100.0 [3,990]	93.8 [87,954]
2003-2005	100.0 [26,030]	99.9 [9,403]	89.2 [48,480]	100.0 [3,990]	93.7 [87,903]
2004-2006	100.0 [26,027]	99.9 [9,405]	97.0 [52,720]	100.0 [3,990]	98.3 [92,142]
2005-2007	100.0 [26,027]	100.0 [9,409]	97.7 [53,100]	100.0 [3,990]	98.7 [92,527]
2006-2008	99.9 [26,020]	100.0 [9,411]	97.5 [52,950]	100.0 [3,990]	98.5 [92,371]
2007-2009	99.9 [26,022]	99.9 [9,403]	89.9 [48,843]	100.0 [3,990]	94.1 [88,259]
2008-2010	100.0 [26,028]	99.8 [9,394]	87.8 [47,696]	100.0 [3,990]	92.9 [87,109]
2009-2011	100.0 [26,031]	99.9 [9,403]	89.6 [48,687]	100.0 [3,990]	94.0 [88,112]
2010-2012	100.0 [26,033]	99.9 [9,405]	87.0 [47,266]	100.0 [3,990]	92.5 [86,694]
2011-2013	100.0 [26,033]	100.0 [9,410]	91.4 [49,681]	100.0 [3,990]	95.0 [89,114]
2012-2014	100.0 [26,030]	99.9 [9,406]	89.3 [48,506]	100.0 [3,990]	93.8 [87,932]
2013-2015	100.0 [26,030]	99.9 [9,406]	91.6 [49,763]	100.0 [3,990]	95.1 [89,188]
2014-2016	100.0 [26,027]	100.0 [9,408]	88.8 [48,268]	100.0 [3,990]	93.5 [87,693]
2015-2017	99.9 [26,024]	99.9 [9,403]	86.0 [46,726]	100.0 [3,990]	91.9 [86,144]
2016-2018	100.0 [26,028]	99.8 [9,396]	81.2 [44,140]	100.0 [3,990]	89.1 [83,554]
2017-2019	100.0 [26,029]	99.6 [9,378]	80.8 [43,898]	100.0 [3,990]	88.8 [83,295]
2018-2020	100.0 [26,032]	99.3 [9,348]	80.0 [43,488]	100.0 [3,990]	88.4 [82,858]
2019-2021	99.9 [26,011]	99.1 [9,327]	75.8 [41,200]	100.0 [3,990]	85.9 [80,528]
Reduction 2003-2020	0.1 [22]	0.8 [74]	13.5 [7,330]	0.0 [0]	7.9 [7,427]

The magnitude of the nutrient-nitrogen exceedance (expressed as Excess Nitrogen or AAE) across the UK has decreased, from 10.6 kg N ha⁻¹ year⁻¹ in 2003 to 7.6 kg N ha⁻¹ year⁻¹ in 2020 (Table 2.5, Figure 2.3). Excess Nitrogen varied among regions, with lowest values in Scotland and highest in Northern Ireland (Table 2.5, Figure 2.4).

Table 2.5: Nutrient nitrogen: Excess Nitrogen (Average Accumulated Exceedance in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	England	Wales	Scotland	NI	UK
2002-2004	19.2	14.2	5.7	14.0	10.6
2003-2005	18.9	13.9	5.5	13.4	10.4
2004-2006	18.8	15.2	7.4	14.1	11.7
2005-2007	18.6	15.7	7.4	14.9	11.7
2006-2008	18.1	15.7	7.1	15.0	11.3
2007-2009	17.4	13.6	5.1	14.8	9.8
2008-2010	17.1	12.9	5.2	15.2	9.7
2009-2011	17.9	13.1	5.8	16.3	10.3
2010-2012	17.6	13.1	5.6	15.8	10.1
2011-2013	17.8	14.0	6.1	16.4	10.6
2012-2014	17.1	13.5	5.8	15.6	10.1
2013-2015	17.3	14.3	6.5	15.0	10.7
2014-2016	17.0	14.6	6.0	14.0	10.3
2015-2017	16.0	13.4	5.1	13.4	9.3
2016-2018	16.4	12.5	4.3	14.9	8.9
2017-2019	16.9	12.0	4.3	16.2	9.1
2018-2020	16.8	12.1	4.2	16.1	9.0
2019-2021	14.5	10.7	3.3	13.8	7.6
Reduction 2003-2020	4.7	3.5	2.4	0.2	3

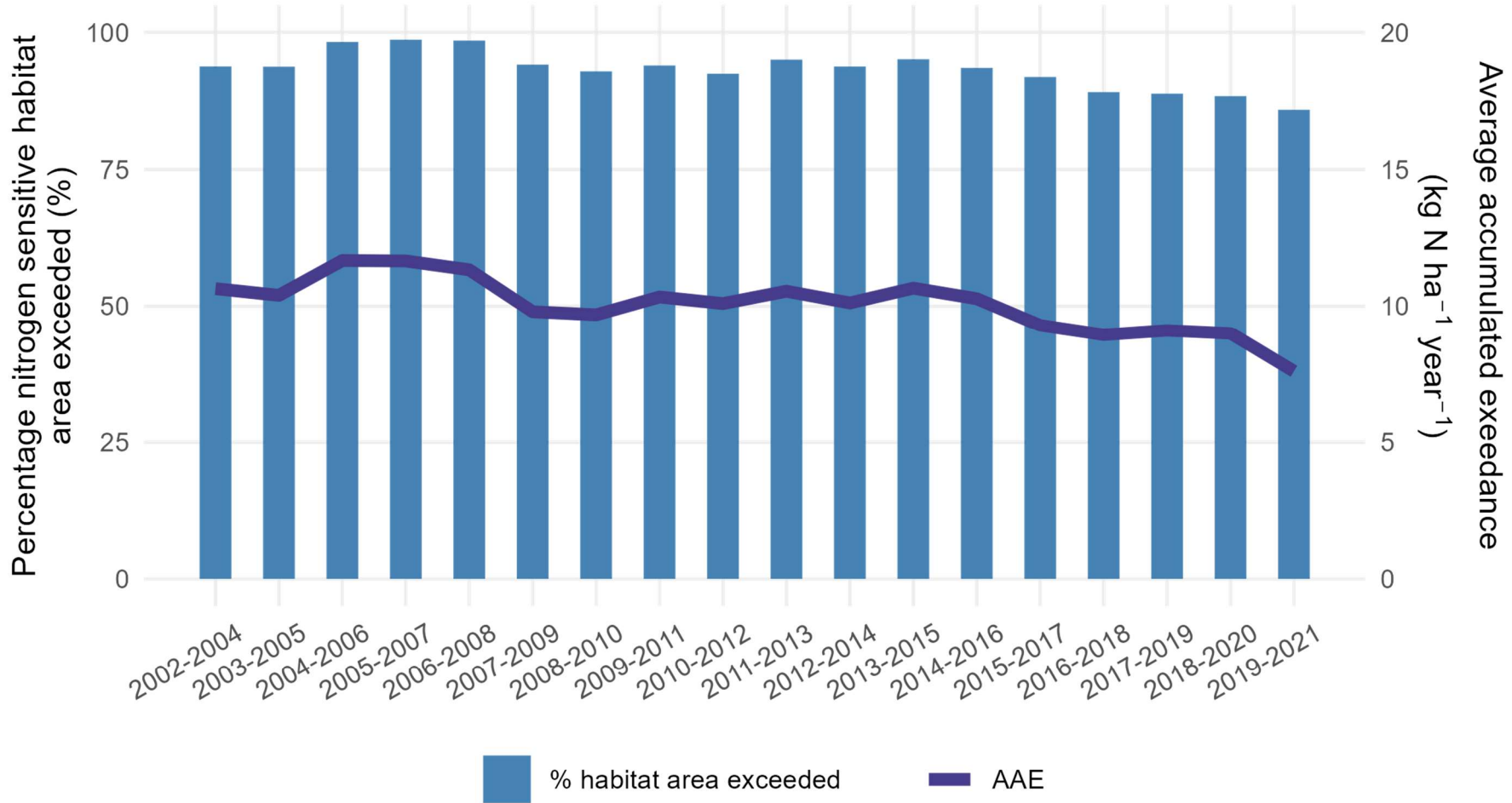


Figure 2.3: Nutrient nitrogen: Percentage area of nitrogen-sensitive habitats with exceedance of nitrogen critical loads in the UK by year, and Excess Nitrogen (Average Accumulated Exceedance in kg N ha⁻¹ year⁻¹).

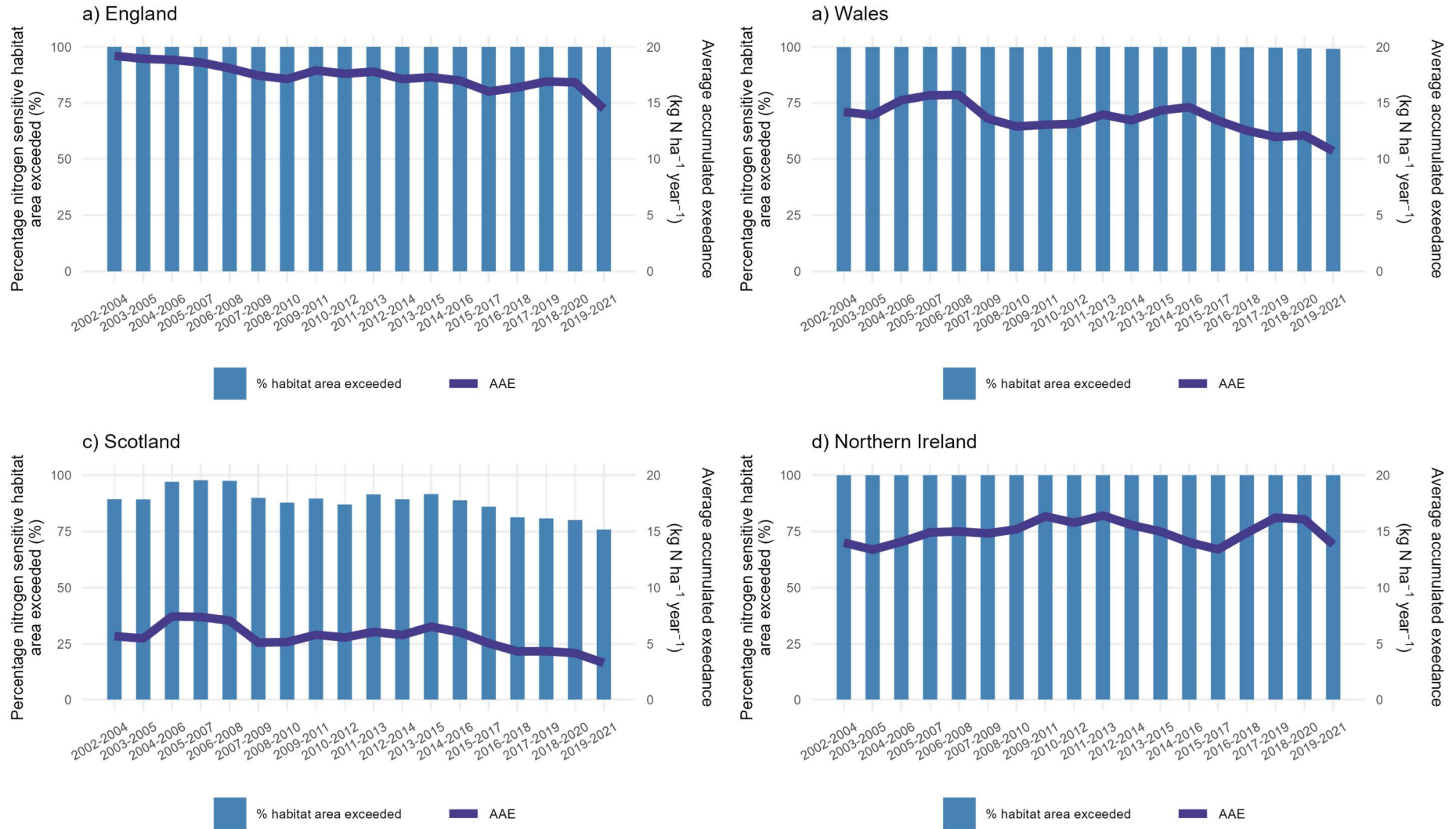


Figure 2.4: Nutrient nitrogen: Percentage area of nitrogen-sensitive habitats with exceedance of nitrogen critical loads, by country and year, and Excess Nitrogen (Average Accumulated Exceedance in kg N ha⁻¹ year⁻¹).

2.2 Trends by habitat

This section focuses on the results by habitat for the UK. Habitat results for individual countries are also calculated and tabulated in the Annexe to this report.

2.2.1 Acidity results

The habitats with the highest percentage area where acidity critical load was exceeded are acid grassland and coniferous woodlands (Table 2.6, Figure 2.5); these habitats also have some of the highest Excess Acidity values (Table 2.7, Figure 2.5). Acidity critical load was exceeded over relatively small areas of calcareous grassland, declining from 14.9% in 2003 to 2.1% in 2020 (Table 2.6), with 0.0 keq ha⁻¹ year⁻¹ Excess Acidity value (Table 2.7). Montane habitat shows the largest decrease in the area exceeded, from 92.5% (4,547 km²) in 2003 to 45.8% (2,249 km²) in 2020. The largest reductions in Excess Acidity over the same timescale are for acid grassland and montane habitats (Table 2.7).

2.2.2 Nutrient nitrogen results

With exception of bog, all habitats had more than 80% of their area exceeded for nutrient N in all years from 2003 to 2020 (Table 2.8, Figure 2.6). The largest reduction in the percentage area exceeded is for bog from 93.1% in 2003 to 67.8% in 2020. Excess Nitrogen is generally highest for the woodland habitats (Table 2.9, Figure 2.6), apart from managed coniferous woodlands for which critical load is set on a different basis (see Section 1.1.3) and scots pine woodland which is only found natively in Scotland where deposition is generally lower.

Table 2.6: Acid-sensitive habitat areas in the UK and percentage area of habitats where acidity critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
<i>Habitat area (km²)</i>	<i>20365</i>	<i>1012</i>	<i>21846</i>	<i>9118</i>	<i>4915</i>	<i>15935</i>	<i>17354</i>	<i>7856</i>	<i>90543</i>
2002-2004	85.6	14.9	34.0	54.9	92.5	83.9	66.0	24.6	65.6
2003-2005	85.0	13.8	31.8	52.5	92.0	82.0	64.6	24.9	64.0
2004-2006	88.1	14.5	40.8	68.4	98.0	85.7	66.1	34.5	69.8
2005-2007	87.2	13.1	37.7	74.6	96.8	85.3	66.0	26.1	69.3
2006-2008	85.7	12.7	31.7	70.5	96.3	84.8	64.7	25.3	66.7
2007-2009	77.9	11.9	23.1	55.7	79.4	78.4	59.0	23.0	58.2
2008-2010	77.9	10.1	24.1	52.6	80.1	77.8	57.6	22.2	57.8
2009-2011	81.1	11.3	27.0	52.6	83.1	79.8	59.7	21.5	60.1
2010-2012	79.1	11.3	26.8	52.0	80.3	79.1	58.8	21.3	59.2
2011-2013	82.4	12.4	26.9	50.8	87.2	81.6	60.7	23.4	61.0
2012-2014	81.5	11.3	27.1	51.0	85.2	79.8	58.2	23.3	60.0
2013-2015	83.8	11.3	28.6	51.9	92.3	81.3	58.4	23.9	61.6
2014-2016	82.6	10.0	26.5	48.7	89.5	78.0	56.3	23.6	59.4
2015-2017	77.9	7.1	19.9	44.6	80.5	72.0	52.5	21.5	54.0
2016-2018	72.1	5.5	18.5	43.1	56.7	67.0	51.3	16.9	49.8
2017-2019	70.5	6.1	18.3	42.9	51.3	67.4	52.1	16.4	49.3
2018-2020	70.5	4.6	17.7	40.9	48.9	67.3	52.0	16.0	48.8
2019-2021	66.0	2.1	13.9	35.7	45.8	62.3	47.9	15.2	44.5
Reduction 2003-2020	19.6	12.8	20.1	19.3	46.7	21.6	18.1	9.5	21.2

Table 2.7: Excess Acidity (Average Accumulated Exceedance in keq ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
2002-2004	0.73	0.09	0.18	0.45	0.69	0.87	0.73	0.25	0.58
2003-2005	0.70	0.07	0.15	0.43	0.63	0.83	0.69	0.25	0.55
2004-2006	0.82	0.07	0.19	0.49	0.85	0.90	0.68	0.29	0.61
2005-2007	0.75	0.06	0.17	0.47	0.70	0.86	0.65	0.27	0.57
2006-2008	0.69	0.05	0.14	0.44	0.62	0.82	0.60	0.25	0.53
2007-2009	0.52	0.04	0.11	0.36	0.40	0.67	0.53	0.20	0.42
2008-2010	0.51	0.04	0.11	0.34	0.43	0.66	0.51	0.18	0.41
2009-2011	0.56	0.04	0.13	0.37	0.50	0.73	0.54	0.19	0.45
2010-2012	0.57	0.04	0.13	0.38	0.47	0.72	0.53	0.20	0.45
2011-2013	0.63	0.05	0.13	0.40	0.51	0.77	0.54	0.22	0.48
2012-2014	0.60	0.04	0.13	0.38	0.50	0.72	0.50	0.21	0.45
2013-2015	0.65	0.04	0.12	0.38	0.60	0.73	0.48	0.21	0.47
2014-2016	0.61	0.04	0.11	0.35	0.55	0.66	0.45	0.20	0.43
2015-2017	0.48	0.02	0.08	0.28	0.37	0.55	0.40	0.17	0.35
2016-2018	0.40	0.01	0.08	0.26	0.23	0.53	0.43	0.14	0.32
2017-2019	0.37	0.01	0.08	0.26	0.20	0.54	0.46	0.13	0.32
2018-2020	0.36	0.01	0.07	0.24	0.18	0.54	0.45	0.12	0.31
2019-2021	0.28	0.00	0.05	0.18	0.14	0.43	0.36	0.09	0.25
Reduction 2003-2020	0.45	0.09	0.13	0.26	0.55	0.44	0.37	0.16	0.34

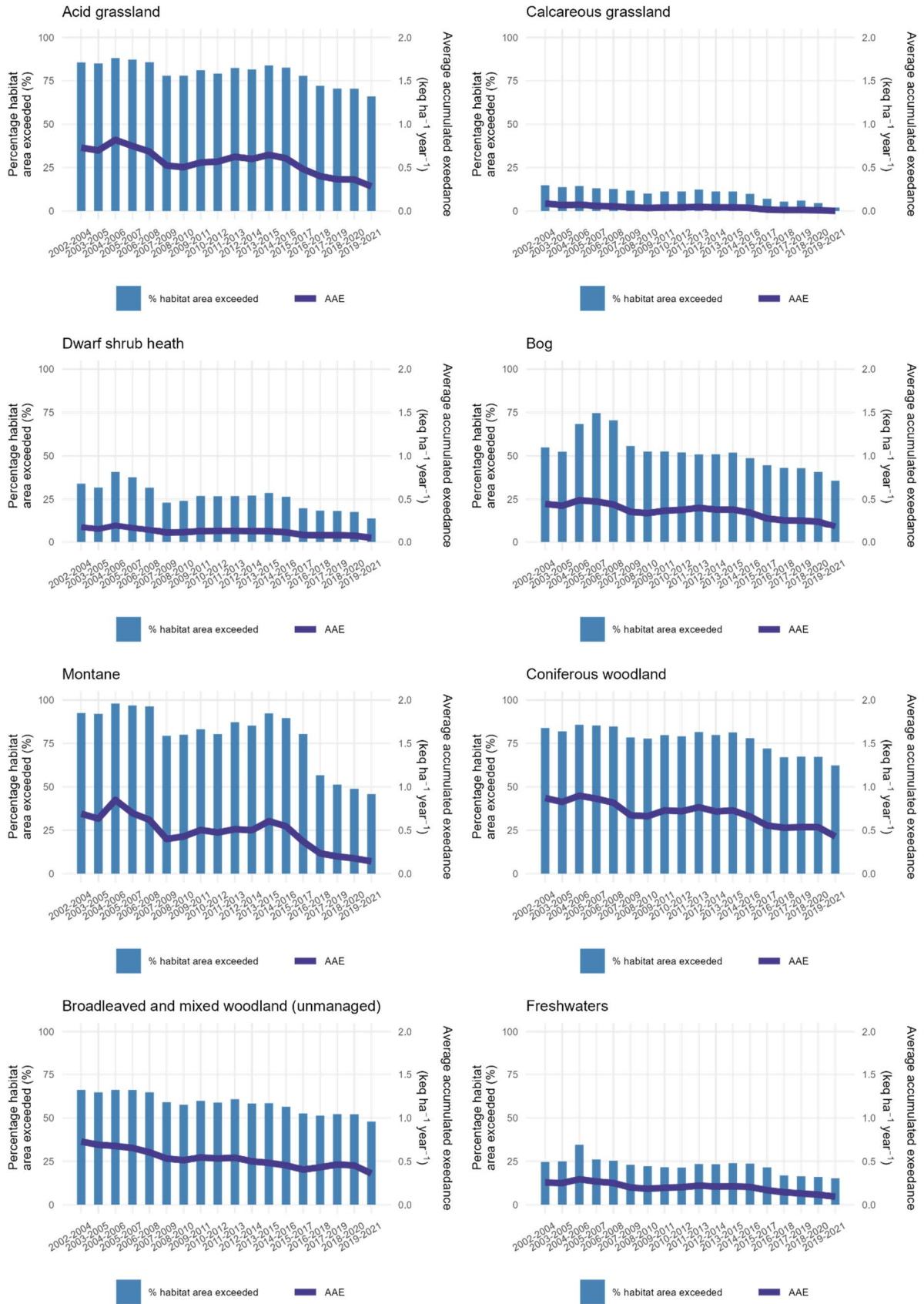


Figure 2.5: Acidity: Percentage area of habitats where acidity critical loads are exceeded, and Excess Acidity (AAE in keq ha⁻¹ year⁻¹) for the UK by deposition dataset year.

Table 2.8: Nutrient-sensitive habitat area in the UK and percentage area of habitats where nutrient-nitrogen critical load is exceeded.

Years	Habitat area (km ²)													
	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	
	20365	1012	21846	9118	4915	14450	8706	2059	6958	1485	1422	631	808	
2002-2004	91.7	99.9	92.7	93.1	100.0	91.4	97.4	100.0	96.3	100.0	97.9	91.8	90.5	
2003-2005	92.1	99.9	93.2	91.7	100.0	90.8	97.3	100.0	95.9	100.0	97.6	91.2	89.6	
2004-2006	97.2	99.9	99.1	99.3	100.0	96.9	98.5	100.0	98.6	100.0	99.0	94.7	91.3	
2005-2007	97.2	99.9	99.3	99.4	100.0	98.7	99.1	100.0	98.9	100.0	99.3	94.8	91.4	
2006-2008	96.1	99.9	98.7	99.3	100.0	99.5	99.5	100.0	99.5	100.0	99.6	94.8	90.0	
2007-2009	89.2	99.6	93.9	93.2	100.0	94.9	97.7	100.0	96.4	100.0	98.2	91.6	87.9	
2008-2010	88.8	99.6	91.2	91.5	100.0	93.0	97.4	100.0	96.3	100.0	97.9	91.2	88.4	
2009-2011	90.0	99.7	92.8	93.3	100.0	93.8	97.7	100.0	96.9	100.0	98.3	92.8	90.6	
2010-2012	87.2	99.9	90.8	92.0	100.0	92.6	97.3	100.0	96.2	100.0	97.6	92.2	90.6	
2011-2013	90.6	99.9	94.1	93.6	100.0	96.6	98.4	100.0	98.3	100.0	99.0	93.0	91.2	
2012-2014	90.1	99.9	92.7	93.0	100.0	93.1	97.7	100.0	96.8	100.0	98.3	92.7	90.0	
2013-2015	93.6	99.9	93.6	92.1	100.0	94.7	98.3	100.0	98.2	100.0	98.9	92.9	90.8	
2014-2016	93.1	99.7	91.1	89.8	100.0	91.8	97.5	100.0	96.4	100.0	97.9	91.6	90.6	
2015-2017	91.1	99.7	89.9	84.5	100.0	90.4	97.1	100.0	94.9	100.0	96.9	89.4	88.9	
2016-2018	86.3	99.7	86.6	80.7	98.9	88.4	96.6	100.0	92.8	100.0	95.0	91.6	88.2	
2017-2019	84.8	99.8	86.4	81.4	96.6	89.0	96.8	100.0	93.2	100.0	95.1	93.5	87.1	
2018-2020	85.4	99.9	86.4	75.9	96.1	88.9	96.9	100.0	93.1	100.0	95.0	93.4	86.4	
2019-2021	81.4	99.3	83.7	67.8	97.8	88.0	96.7	100.0	92.0	100.0	94.5	89.8	82.8	
Reduction 2003-2020	10.3	0.6	8.9	25.2	2.2	3.3	0.7	0	4.3	0	3.4	2	7.7	

Table 2.9: Nutrient nitrogen: Excess Nitrogen (AAE in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh
2002-2004	8.3	10.7	6.1	7.3	9.3	13.5	20.1	20.9	16.4	9.4	17.6	5.5	4.9
2003-2005	8.2	10.0	5.9	7.2	9.0	13.2	19.7	20.6	16.0	9.1	17.1	5.3	4.8
2004-2006	10.0	10.3	7.7	8.7	11.9	14.2	19.3	19.7	16.3	11.4	17.4	5.6	4.8
2005-2007	9.6	10.0	7.7	9.1	10.8	14.3	19.4	19.9	16.5	11.1	17.8	5.9	4.8
2006-2008	9.4	9.9	7.3	8.9	10.6	14.2	18.8	19.4	16.1	11.3	17.6	5.8	4.7
2007-2009	7.5	9.5	5.7	7.5	7.9	12.5	18.1	18.8	14.8	9.0	16.7	5.4	4.2
2008-2010	7.5	9.5	5.7	7.2	8.2	12.3	18.0	18.4	14.5	9.0	16.7	5.5	4.3
2009-2011	8.2	9.9	6.1	7.5	8.9	13.3	18.8	19.0	15.3	9.6	17.7	5.9	4.7
2010-2012	8.2	9.7	5.8	7.5	8.4	13.0	18.3	18.4	14.8	9.1	17.2	5.9	4.8
2011-2013	8.8	10.0	6.2	7.9	9.0	13.7	18.2	18.5	15.3	10.1	17.6	5.9	4.8
2012-2014	8.6	9.9	6.0	7.5	8.9	13.0	17.4	17.8	14.4	9.6	16.6	5.5	4.4
2013-2015	9.5	10.0	6.5	7.9	10.5	13.5	17.2	17.7	14.8	10.5	16.6	5.5	4.3
2014-2016	9.3	9.9	6.2	7.4	10.0	12.8	16.9	17.5	14.4	9.5	16.0	5.4	4.4
2015-2017	8.0	9.2	5.5	6.6	8.2	11.6	16.3	16.8	13.5	8.2	15.2	5.2	4.1
2016-2018	7.0	9.4	5.1	6.2	6.1	11.4	17.4	17.8	13.6	6.7	16.1	5.4	4.4
2017-2019	6.7	10.0	5.2	6.3	5.4	11.6	18.7	18.4	14.1	6.8	17.3	5.8	4.8
2018-2020	6.7	9.4	5.2	6.1	5.1	11.6	18.5	18.2	13.9	6.5	17.2	5.7	4.6
2019-2021	5.5	7.6	4.3	5.1	4.6	9.7	15.9	15.6	12.0	6.0	14.9	4.8	3.6
Reduction 2003-2020	2.9	3.1	1.8	2.1	4.7	3.8	4.2	5.3	4.4	3.4	2.7	0.7	1.4

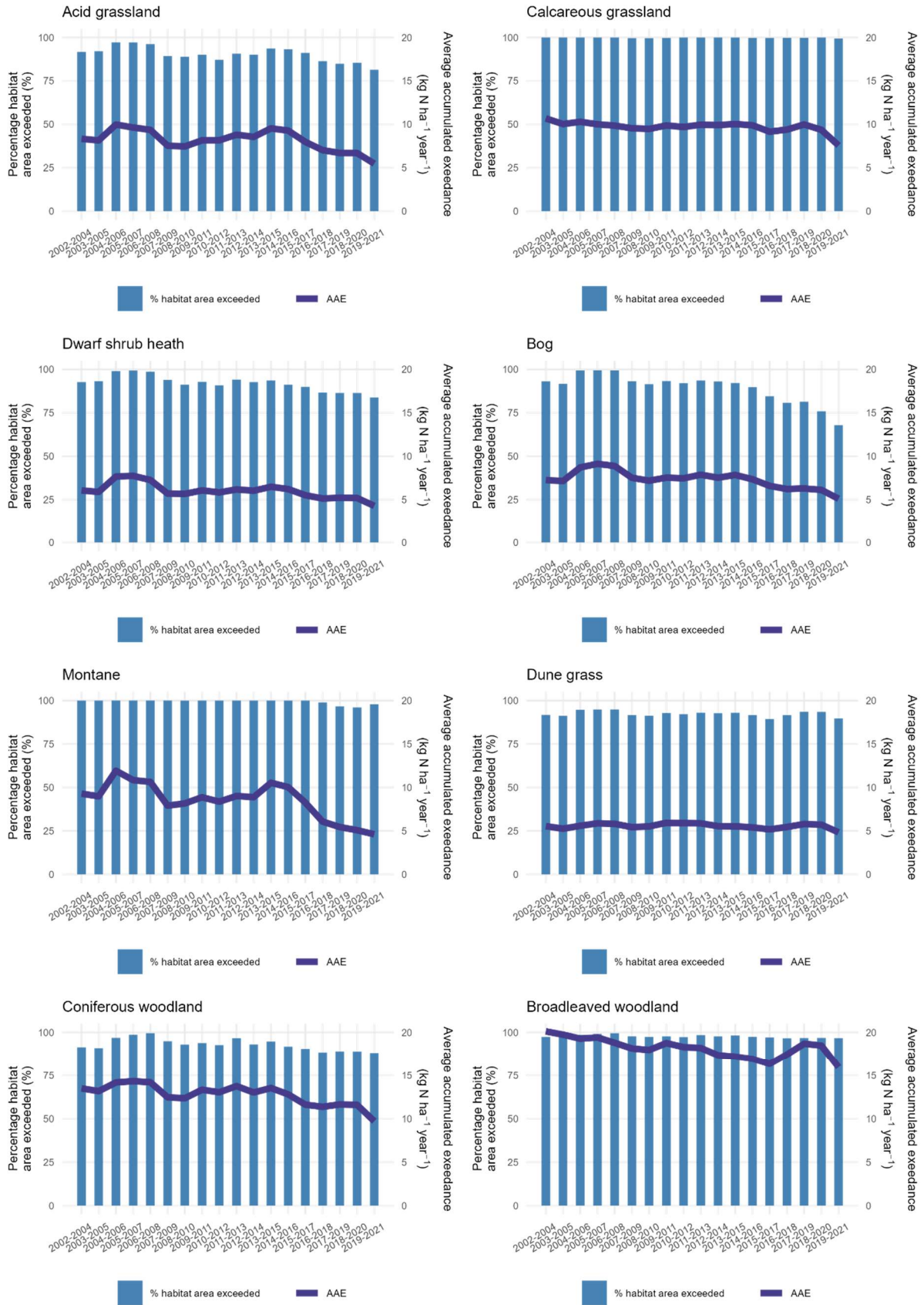


Figure 2.6: Nutrient nitrogen: Percentage area of habitats where nutrient nitrogen critical loads is exceeded, and Excess Nitrogen (AAE in kg N ha⁻¹ year⁻¹) in the UK by deposition dataset year.

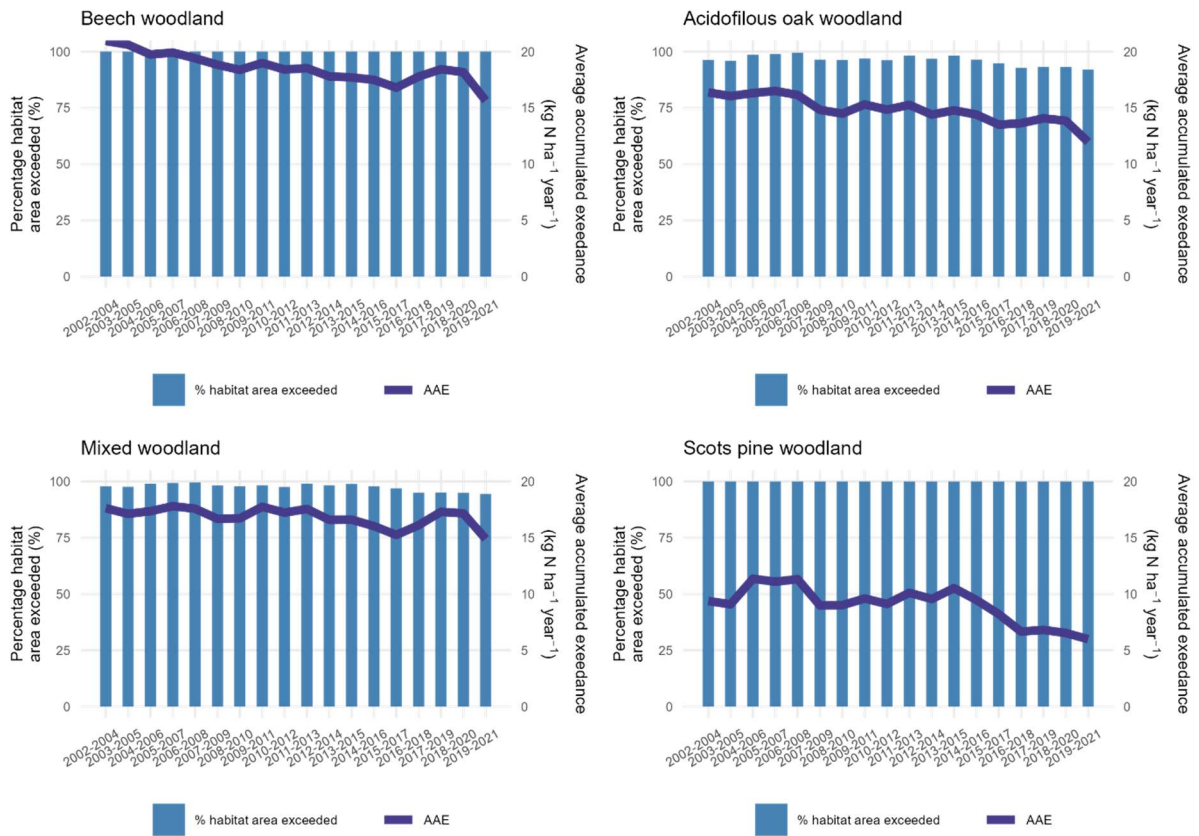


Figure 2.6 (continued): Nutrient nitrogen: Percentage area of habitats where nutrient nitrogen critical load is exceeded, and Excess Nitrogen (Average Accumulated Exceedance in kg N ha⁻¹ year⁻¹) in the UK by deposition dataset year.

Section 3: Site-relevant critical loads and their exceedances

This section of the report focuses on the application of critical loads to sites designated for their nature conservation importance, hereafter referred to as site-relevant critical loads (SRCL), and their exceedances. The critical loads are based on the same methods applied to UK acid- and N-sensitive habitats described in Section 1 of this report, and are applied to acid- and N-sensitive features within the designated sites. Exceedances of critical loads are also calculated in the same way as the habitats (Section 1) and also based on UK 5x5 km CBED deposition, however, some different metrics are used to describe the exceedance results for SRCL and are explained below.

3.1 Overview of site-relevant critical loads

Site relevant critical loads (SRCL) have been applied to three types of statutory protected sites:

1. Special Areas of Conservation (SACs) are protected sites designated using the criteria set out in Annex III of the EC Habitats Directive. Annexes I and II of the Directive identify the habitats and species (excluding birds) to be protected; 78 Annex I habitat types and 41 species are believed to occur in, or be native to the UK.
2. Special Protected Areas (SPAs) are sites classified using the criteria set out in Article 4 of the EC Birds Directive to protect rare and vulnerable birds (as listed in an Annex to the Directive) and regularly occurring migratory species.
3. Sites of Special Scientific Interest (SSSIs in England, Wales and Scotland) and Areas of Special Scientific Interest (ASSIs in Northern Ireland) provide statutory protection to the UK's flora and fauna. There are additional SSSIs designated for geological or physiographic features but these are not included in the SRCL assessments.

Digital boundaries for all sites in the UK have been collated by each of the country conservation agencies, together with tables identifying the designated feature habitats and species associated with each site, but no digital information is currently available on the spatial area of each feature within each site. Therefore it is assumed that all features recorded for a site occur across the entire site area. To avoid double counting the area exceeding critical loads for sites with more than one designated feature, the maximum area exceeded for any feature is used when summarising results to the site and country levels (Section 3.2).

To assign SRCLs, the first step is to consider whether the interest feature is potentially sensitive to acidification and/or eutrophication. Specialists within Natural England, NatureScot and CEH used expert judgement to determine this (SNIFFER, 2007). For SPAs where the features are bird species, the broad habitats that the birds depend upon for feeding, breeding and roosting are considered.

To assign critical loads to the habitat features of designated sites, it is necessary to cross-match the different habitat classifications used. Acidity critical loads are mapped by broad habitat, and empirical critical loads for nutrient-nitrogen are assigned based on the European Nature Information System (EUNIS) habitat classification. Look-up tables developed by Moss & Davies (2002) and available from

the JNCC website (<https://hub.jncc.gov.uk/assets/9e70531b-5467-4136-88f6-3b3dd905b56d>) enable linkages to be made between:

- Annex I habitats and EUNIS classes
- Annex I habitats and broad habitats
- EUNIS habitats and broad habitats

Using the look-up tables the most appropriate EUNIS class and broad habitat class were assigned to each interest feature. However, some sites contain features that may be sensitive to acidification and/or eutrophication, but for which no appropriate critical loads are available.

The critical loads assigned to protected sites are based on designated features. However, the national critical load maps are based on national-scale datasets, which may not include small areas of sensitive habitats or some coastal habitats. Some sites are designated for habitat areas that are not included in the national-scale habitat mapping for critical loads. To overcome this, for SRCLs a separate database of national critical loads for terrestrial habitats was created, that provides critical loads for every 1 km x 1 km square in the UK, whether the habitat is known to exist there or not. The appropriate SRCL can then be extracted for terrestrial habitat features of each designated site. The SRCLs assigned do not take into account acidity critical loads for freshwater habitats. For further information refer to the “Methods Report” (Hall et al., 2015).

For nutrient N, the empirical critical loads approach is applied to designated feature habitats sensitive to N. The critical load values applied to each habitat are the “Recommended” values agreed by habitat specialists for Article 17 reporting. For more information refer to Hall et al. (2015).

The tables later in this section give percentages of the number of **sensitive** sites (those with a SRCL for acidity or for nutrient-N) where the critical load is exceeded for at least one feature. Percentages of the total number of protected sites where a critical load is exceeded can be obtained using the formula below and the numbers in Table 3.1.

$$\% \text{ all sites exceeded} = \frac{\% \text{ sensitive sites exceeded} \times \text{number of sensitive sites}}{\text{total number of sites}}$$

Table 3.1. Total number (T) of protected sites in UK countries in 2023, and numbers of sites with at least one feature that is sensitive to acidity (A) or to nutrient-Nitrogen (N). All ASSIs and SSSIs are within a single country. SACs and SPAs that straddle a border are not included in this table.

	SAC			SPA			A/SSSI		
	T	A	N	T	A	N	T	A	N
England	256	180	197	88	63	72	4128	2924	2954
Wales	95	71	79	21	13	14	1079	676	686
Scotland	253	182	201	164	86	124	1422	905	938
Northern Ireland	58	47	50	16	10	12	394	178	188

3.2 Overview of SRCL exceedance metrics

Exceedances are calculated separately for SACs, SPAs and SSSIs, for all site features that critical loads and deposition data can be assigned to (Hall et al., 2015). Metrics are calculated by:

a) Feature (within each site)

- Exceedance
- Exceeded area[#]
- Accumulated Exceedance (AE) (i.e. exceedance * exceeded area)
- Average Accumulated Exceedance (AAE)(i.e. AE / total site area)

b) Site

- Total number of features with SRCL
- Number and percentage of features with exceedance of SRCL.
- Maximum area exceeded^{##} for any feature within a site
- Maximum AE for any feature within a site
- Maximum AAE for any feature within a site

c) Country

- Total number of sites
- Total number and percentage of sites with SRCL for one or more features
- Total number of features with SRCL
- Total number and percentage of sites with exceedance of SRCL for one or more features
- Total number and percentage of features with exceedance of SRCL
- Total area of all sites
- Total area of all sites with SRCL
- Maximum exceeded area^{###}
- Maximum AE calculated as the sum of the maximum AE for all sites
- Maximum AAE; calculated from the country maximum AE and total area of all sites (with SRCL) within a country.

[#] Feature exceeded area: if the critical load is exceeded and the deposition values are constant across the whole site, the exceeded area equals the site area; if the deposition values vary across the site (e.g. as a result of the site crossing the boundaries between grid squares with different deposition values), then the exceeded area will be the sum of the 1 km x 1 km portions of the site where the deposition exceeds the critical load.

^{##} Site maximum exceeded area: set to the maximum exceeded area for any feature within a site.

^{###} Country maximum exceeded area: calculated as the sum of the site maximum exceeded areas for all sites within a country.

The sections below summarise the key results by country, based on the CBED deposition (Section 1.2) for 1996 to 2017. Summary statistics and maps present the worst case, since they are based on exceedance of at least one feature. Other features within a site may have a smaller exceedance, or not be exceeded. The results for AAE are based on the maximum exceedance of any feature within a site.

3.2.1 Exceedance of acidity critical load

The trends in acidity critical load exceedances are summarised in Tables 3.1-3.3 and present the percentage of sites (with SRCL) by country, where the SRCL was exceeded for one or more features, together with the maximum Excess Acidity.

Between 2003 and 2020, there was an 8.3% decrease in the percentage of SACs with critical load exceedance in England and a 20.3 % decrease in Scotland, but less change in Wales or Northern Ireland, 7% and 4.3% accordingly. Excess Acidity for SACs fell by between 26% (NI) to 80% (Scotland). For SPAs, the reductions in four countries span a range from 10% in NI to 18.6% in Scotland, accompanied by a range of 50 - 82% reduction in their maximum Excess Acidity values. For SSSIs (ASSIs in NI) as well, all countries showed reductions in the percentage of exceeded sites (Scotland 26.2%; England 12.9%, Wales 14.3% and NI 2.8%), with a 30 - 76% reduction in maximum Excess Acidity values.

At the UK level, the trends results show:

- for SACs, the percentage of sites with acidity exceedance decreased from 79.9% in 2003 to 67.8% in 2020, and the maximum AAE fell by 59% from 0.88 keq ha⁻¹ year⁻¹ to 0.0.36 keq ha⁻¹ year⁻¹ over the same time period;
- for SPAs, the percentage of exceeded sites fell from 73.7% in 2003 to 57.1% in 2020, and the maximum AAE decreased by 58% from 0.67 keq ha⁻¹ year⁻¹ in 2003 to 0.28 keq ha⁻¹ year⁻¹ in 2020.
- for SSSIs, the percentage of exceeded sites decreased from 64.5% in 2003 to 49.2% in 2020, and the maximum AAE fell by 59% from 0.68 keq ha⁻¹ year⁻¹ in 1996 to 0.28 keq ha⁻¹ year⁻¹ in 2020.

Maps of the maximum AAE per site (Figure 3.1) based on the latest CBED deposition (2020) show the highest exceedances of acidity critical load mainly in northern England, parts of Wales and south-west England and southern Scotland. Some sites in the far north of Scotland have no exceedance for any site feature due to low levels of acid deposition. Some SACs and SPAs and many small SSSIs in southern and eastern England are also not exceeded, because more calcareous soils in this region mean that site features are less sensitive to acidity pollution.

Table 3.2: Trends in acidity exceedances for SACs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Acidity [maximum AAE keq ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	Eng/Wales*	Eng/Scot*	UK
Number of sites with SRCL	180	71	182	47	6	1	487
2002-2004	76.7 [1.46]	93.0 [1.04]	72.5 [0.30]	97.9 [0.79]	100.0 [NR]	100.0 [NR]	79.9 [0.88]
2003-2005	76.7 [1.41]	93.0 [1.01]	73.6 [0.28]	95.7 [0.71]	100.0 [NR]	100.0 [NR]	80.1 [0.84]
2004-2006	76.1 [1.40]	93.0 [1.10]	80.2 [0.41]	97.9 [0.77]	100.0 [NR]	100.0 [NR]	82.5 [0.91]
2005-2007	76.1 [1.31]	94.4 [1.10]	80.2 [0.39]	97.9 [0.79]	100.0 [NR]	100.0 [NR]	82.8 [0.86]
2006-2008	73.9 [1.23]	95.8 [1.04]	76.4 [0.33]	97.9 [0.77]	100.0 [NR]	100.0 [NR]	80.7 [0.79]
2007-2009	73.3 [1.11]	91.5 [0.85]	66.5 [0.20]	97.9 [0.74]	100.0 [NR]	100.0 [NR]	76.2 [0.66]
2008-2010	73.9 [1.06]	88.7 [0.79]	67.0 [0.21]	97.9 [0.76]	100.0 [NR]	100.0 [NR]	76.2 [0.64]
2009-2011	73.9 [1.11]	90.1 [0.75]	71.4 [0.24]	97.9 [0.80]	100.0 [NR]	100.0 [NR]	78.0 [0.67]
2010-2012	73.9 [1.13]	91.5 [0.79]	69.8 [0.23]	97.9 [0.79]	100.0 [NR]	100.0 [NR]	77.6 [0.68]
2011-2013	74.4 [1.18]	91.5 [0.86]	72.5 [0.24]	97.9 [0.82]	100.0 [NR]	100.0 [NR]	78.9 [0.71]
2012-2014	73.3 [1.12]	90.1 [0.84]	71.4 [0.23]	97.9 [0.78]	100.0 [NR]	100.0 [NR]	77.8 [0.68]
2013-2015	72.2 [1.11]	90.1 [0.88]	72.5 [0.27]	97.9 [0.68]	100.0 [NR]	100.0 [NR]	77.8 [0.69]
2014-2016	70.0 [1.03]	90.1 [0.88]	70.9 [0.23]	97.9 [0.61]	100.0 [NR]	100.0 [NR]	76.4 [0.64]
2015-2017	69.4 [0.88]	88.7 [0.77]	64.3 [0.15]	95.7 [0.56]	100.0 [NR]	100.0 [NR]	73.3 [0.53]
2016-2018	71.7 [0.84]	88.7 [0.67]	61.5 [0.09]	97.9 [0.66]	100.0 [NR]	100.0 [NR]	73.3 [0.48]
2017-2019	72.2 [0.84]	87.3 [0.58]	62.6 [0.08]	97.9 [0.73]	100.0 [NR]	100.0 [NR]	73.7 [0.47]
2018-2020	71.7 [0.80]	87.3 [0.53]	61.5 [0.07]	97.9 [0.73]	100.0 [NR]	100.0 [NR]	73.1 [0.44]
2019-2021	68.3 [0.64]	85.9 [0.44]	52.2 [0.06]	93.6 [0.58]	100.0 [NR]	100.0 [NR]	67.8 [0.36]
Reduction 2003-2020	8.3 [0.82]	7.0 [0.60]	20.3 [0.24]	4.3 [0.21]	0.0 [NR]	0.0 [NR]	12.1 [0.52]

* Some sites cross the England/Wales or England/Scotland border. For Excess Acidity calculations, each 1 x 1 km square (or part thereof) within each border site was assigned to a single country, so results are calculated for individual countries only.

Table 3.3: Trends in acidity exceedances for SPAs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Acidity [maximum AAE keq ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	Eng/Wales*	UK
Number of sites with SRCL	63	13	86	10	3	175
2002-2004	93.7 [1.21]	76.9 [0.94]	57.0 [0.17]	80.0 [0.28]	100.0 [NR]	73.7 [0.67]
2003-2005	92.1 [1.16]	76.9 [0.88]	58.1 [0.15]	70.0 [0.24]	100.0 [NR]	73.1 [0.63]
2004-2006	90.5 [1.15]	76.9 [0.96]	64.0 [0.23]	80.0 [0.29]	100.0 [NR]	76.0 [0.67]
2005-2007	88.9 [1.06]	76.9 [0.97]	66.3 [0.22]	70.0 [0.29]	100.0 [NR]	76.0 [0.63]
2006-2008	87.3 [1.01]	84.6 [0.85]	61.6 [0.18]	70.0 [0.29]	100.0 [NR]	73.7 [0.58]
2007-2009	88.9 [0.93]	76.9 [0.63]	54.7 [0.10]	70.0 [0.24]	100.0 [NR]	70.3 [0.49]
2008-2010	87.3 [0.90]	69.2 [0.55]	53.5 [0.11]	70.0 [0.26]	100.0 [NR]	68.6 [0.48]
2009-2011	87.3 [0.94]	76.9 [0.53]	60.5 [0.14]	70.0 [0.31]	100.0 [NR]	72.6 [0.51]
2010-2012	85.7 [0.96]	76.9 [0.55]	58.1 [0.13]	70.0 [0.31]	100.0 [NR]	70.9 [0.52]
2011-2013	87.3 [0.97]	76.9 [0.62]	55.8 [0.14]	80.0 [0.33]	100.0 [NR]	70.9 [0.53]
2012-2014	85.7 [0.94]	61.5 [0.59]	57.0 [0.13]	80.0 [0.29]	100.0 [NR]	69.7 [0.51]
2013-2015	85.7 [0.91]	53.8 [0.64]	55.8 [0.15]	70.0 [0.22]	100.0 [NR]	68.0 [0.51]
2014-2016	84.1 [0.87]	61.5 [0.64]	54.7 [0.13]	70.0 [0.18]	100.0 [NR]	67.4 [0.48]
2015-2017	82.5 [0.75]	61.5 [0.52]	48.8 [0.08]	70.0 [0.14]	100.0 [NR]	64.0 [0.40]
2016-2018	84.1 [0.74]	61.5 [0.44]	47.7 [0.06]	70.0 [0.22]	100.0 [NR]	64.0 [0.38]
2017-2019	85.7 [0.74]	61.5 [0.37]	51.2 [0.05]	80.0 [0.26]	100.0 [NR]	66.9 [0.37]
2018-2020	82.5 [0.71]	61.5 [0.34]	51.2 [0.05]	80.0 [0.24]	100.0 [NR]	65.7 [0.35]
2019-2021	77.8 [0.56]	61.5 [0.26]	38.4 [0.03]	70.0 [0.14]	100.0 [NR]	57.1 [0.28]
Reduction 2003-2020	15.9 [0.65]	15.4 [0.68]	18.6 [0.14]	10.0 [0.14]	0.0 [NR]	16.6 [0.39]

* Some sites cross the England/Wales border. No SPAs cross the England/Scotland border. For Excess Acidity calculations, each 1 x 1 km square (or part thereof) within each border site was assigned to a single country, so results are calculated for individual countries only.

Table 3.4: Trends in acidity exceedances for SSSIs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Acidity [maximum AAE keq ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	UK
Number of sites with SRCL	2924	676	905	178	4683
2002-2004	61.0 [1.01]	77.7 [0.93]	64.0 [0.29]	74.7 [0.63]	64.5 [0.68]
2003-2005	60.3 [0.97]	76.9 [0.89]	59.6 [0.27]	73.6 [0.56]	63.1 [0.65]
2004-2006	58.7 [0.97]	77.2 [0.98]	68.6 [0.39]	74.2 [0.62]	63.9 [0.71]
2005-2007	57.9 [0.90]	77.8 [0.98]	67.0 [0.37]	73.6 [0.64]	63.1 [0.67]
2006-2008	56.8 [0.84]	78.3 [0.93]	64.8 [0.31]	74.7 [0.62]	62.1 [0.62]
2007-2009	55.9 [0.76]	75.3 [0.75]	55.1 [0.19]	73.6 [0.59]	59.2 [0.50]
2008-2010	55.3 [0.72]	72.8 [0.69]	56.1 [0.19]	73.6 [0.61]	58.7 [0.48]
2009-2011	55.3 [0.76]	74.3 [0.66]	60.4 [0.22]	75.3 [0.65]	59.8 [0.51]
2010-2012	55.4 [0.77]	74.3 [0.70]	59.8 [0.22]	74.7 [0.64]	59.7 [0.52]
2011-2013	55.1 [0.80]	75.4 [0.76]	59.4 [0.24]	77.5 [0.67]	59.7 [0.55]
2012-2014	54.5 [0.76]	73.1 [0.75]	57.6 [0.23]	76.4 [0.63]	58.6 [0.53]
2013-2015	53.1 [0.75]	72.9 [0.78]	57.6 [0.26]	74.7 [0.55]	57.7 [0.54]
2014-2016	52.3 [0.70]	72.9 [0.79]	54.1 [0.23]	71.9 [0.49]	56.4 [0.50]
2015-2017	50.8 [0.60]	70.4 [0.68]	47.2 [0.16]	70.2 [0.44]	53.7 [0.41]
2016-2018	52.7 [0.57]	70.7 [0.59]	44.9 [0.12]	73.6 [0.52]	54.6 [0.37]
2017-2019	53.0 [0.57]	69.4 [0.52]	46.1 [0.11]	76.4 [0.58]	54.9 [0.36]
2018-2020	52.1 [0.55]	67.8 [0.50]	45.0 [0.10]	75.8 [0.58]	53.9 [0.35]
2019-2021	48.1 [0.44]	63.3 [0.42]	37.8 [0.07]	71.9 [0.45]	49.2 [0.28]
Reduction 2003-2020	12.9 [0.57]	14.3 [0.51]	26.2 [0.22]	2.8 [0.19]	15.3 [0.40]

(a) SAC

(b) SPA

(a) SSSI

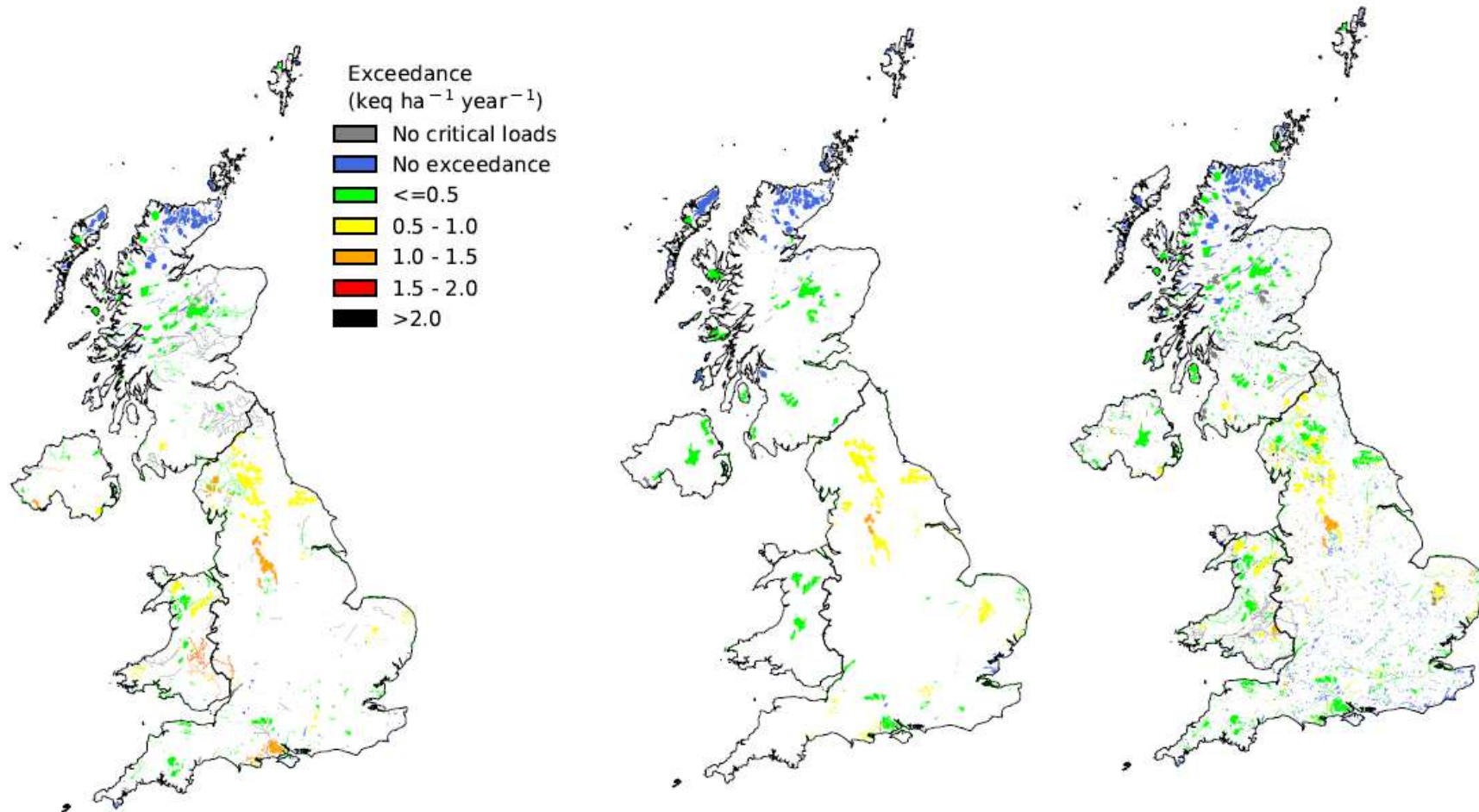


Figure 3.1: Excess Acidity. Average Accumulated Exceedance (AAE) of acidity critical loads by CBED deposition for 2019-21; maps show the maximum Excess Acidity for any feature within each site (other features may have lower or no exceedance).

3.2.2 Nutrient-nitrogen results

The trends in nutrient-N critical load exceedances from 2003 to 2020 are summarised in Tables 3.4-3.6. The decreases in the percentage of sites with exceedance of nutrient-N critical loads for one or more features, and decreases in Excess Nitrogen, were smaller than the decreases seen for acidity, reflecting the smaller decline in N deposition than in acidity deposition. Reductions varied by country for the different site types. In Northern Ireland and Wales, there was no decrease in the percentage of SACs and SPAs with critical load exceedance and in Excess Nitrogen. For Northern Ireland, the Excess Nitrogen was slightly higher in 2020 than in 2003 (difference of < 1%). The largest decreases in the percentage of sites with exceedance were in Scotland for SACs (5% reduction), SPAs (8.9% reduction) and SSSIs (7.8% reduction). There was a marginal increase (0.1%) in the percentage of SPAs and SSSIs exceeding nutrient-N critical load in Northern Ireland.

At the UK level:

- for SACs, the percentage of sites with nutrient-N exceedance decreased from 97.4% in 2003 to 95.1% in 2020, and the maximum Excess Nitrogen decreased a little from 12.2 kg N ha⁻¹ year⁻¹ to 8.7 kg N ha⁻¹ year⁻¹ over the same time period.
- for SPAs, the percentage of exceeded sites decreased from 91.6% in 2003 to 86.7% in 2020, and the maximum Excess Nitrogen decreased from 12 kg N ha⁻¹ year⁻¹ in 2003 to 9 kg N ha⁻¹ year⁻¹ in 2020.
- for SSSIs, the percentage of exceeded sites decreased from 98.2% in 2003 to 96.4% in 2020, and the maximum Excess Nitrogen decreased from 12.7 kg N ha⁻¹ year⁻¹ in 2003 to 9.3 kg N ha⁻¹ year⁻¹ in 2020.

Maps of the maximum Excess Nitrogen per site (Figure 3.2) based on the latest CBED deposition (2019-2020) show few sites with no exceedance for any feature. Exceedances are widespread across all countries, although lower proportions of protected sites are exceeded in Scotland. Except for Scotland, the maximum Excess Nitrogen is above 7 kg N ha⁻¹ year⁻¹ for the majority of sites, with many sites having maximum Excess Nitrogen up to 28 kg N ha⁻¹ year⁻¹.

Table 3.5: Trends in nutrient nitrogen exceedances for SACs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Nitrogen [maximum AAE kg N ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	Eng/Wales*	Eng/Scot*	UK
Number of sites with SRCL	197	79	201	50	7	2	536
2002-2004	99.0 [18.3]	100.0 [10.8]	94.0 [6.3]	100.0 [12.5]	100.0 [NR]	100.0 [NR]	97.4 [12.2]
2003-2005	99.0 [18.1]	100.0 [10.6]	93.5 [6.2]	100.0 [11.9]	100.0 [NR]	100.0 [NR]	97.2 [12.0]
2004-2006	99.0 [18.1]	100.0 [11.1]	96.0 [8.3]	100.0 [12.5]	100.0 [NR]	100.0 [NR]	98.1 [13.0]
2005-2007	99.0 [17.7]	100.0 [11.7]	97.0 [8.6]	100.0 [13.2]	100.0 [NR]	100.0 [NR]	98.5 [13.0]
2006-2008	99.0 [17.3]	100.0 [11.7]	97.0 [8.3]	100.0 [13.0]	100.0 [NR]	100.0 [NR]	98.5 [12.7]
2007-2009	99.0 [16.5]	100.0 [10.2]	92.0 [6.3]	100.0 [13.1]	100.0 [NR]	100.0 [NR]	96.6 [11.4]
2008-2010	99.0 [16.2]	100.0 [9.6]	91.5 [6.1]	100.0 [13.2]	100.0 [NR]	100.0 [NR]	96.5 [11.1]
2009-2011	99.0 [16.9]	100.0 [9.7]	93.0 [6.5]	100.0 [13.7]	100.0 [NR]	100.0 [NR]	97.0 [11.6]
2010-2012	99.0 [16.8]	100.0 [9.7]	92.0 [6.2]	100.0 [13.4]	100.0 [NR]	100.0 [NR]	96.6 [11.4]
2011-2013	99.0 [17.2]	100.0 [10.3]	94.5 [6.7]	100.0 [13.8]	100.0 [NR]	100.0 [NR]	97.6 [11.9]
2012-2014	99.0 [16.6]	100.0 [9.7]	93.0 [6.4]	100.0 [13.4]	100.0 [NR]	100.0 [NR]	97.0 [11.4]
2013-2015	99.0 [16.8]	100.0 [10.2]	94.0 [7.0]	100.0 [12.5]	100.0 [NR]	100.0 [NR]	97.4 [11.7]
2014-2016	99.0 [16.4]	100.0 [10.3]	93.0 [6.3]	100.0 [11.8]	100.0 [NR]	100.0 [NR]	97.0 [11.3]
2015-2017	99.0 [15.3]	100.0 [9.6]	92.0 [5.5]	100.0 [11.5]	100.0 [NR]	100.0 [NR]	96.6 [10.4]
2016-2018	99.0 [15.2]	100.0 [9.1]	90.0 [4.7]	100.0 [12.9]	100.0 [NR]	100.0 [NR]	95.9 [10.0]
2017-2019	99.0 [15.6]	100.0 [8.8]	91.0 [4.7]	100.0 [14.1]	100.0 [NR]	100.0 [NR]	96.3 [10.2]
2018-2020	99.0 [15.1]	100.0 [9.0]	90.5 [4.6]	100.0 [14.3]	100.0 [NR]	100.0 [NR]	96.1 [10.0]
2019-2021	98.0 [13.2]	100.0 [8.1]	89.1 [4.1]	100.0 [12.4]	100.0 [NR]	100.0 [NR]	95.1 [8.7]
Reduction in 2003-2020	1.0 [5.1]	0.0 [2.7]	5.0 [2.3]	0.0 [0.1]	0.0 [NR]	0.0 [NR]	2.2 [3.5]

* Some sites cross the England/Wales or England/Scotland border. For Excess Nitrogen calculations, each 1 x 1 km square (or part thereof) within each border site was assigned to a single country, so results are calculated for individual countries only.

Table 3.6: Trends in nutrient nitrogen exceedances for SPAs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Nitrogen [maximum AAE kg N ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	Eng/Wales*	UK
Number of sites with SRCL	72	14	124	12	3	225
2002-2004	98.6 [17.4]	100.0 [15.3]	85.5 [5.8]	100.0 [13.2]	100.0 [NR]	91.6 [12.0]
2003-2005	98.6 [17.1]	100.0 [15.0]	82.3 [5.6]	100.0 [12.5]	100.0 [NR]	89.8 [11.8]
2004-2006	98.6 [17.0]	100.0 [15.9]	86.3 [7.4]	100.0 [12.9]	100.0 [NR]	92.0 [12.6]
2005-2007	98.6 [16.4]	100.0 [16.4]	86.3 [7.9]	100.0 [13.8]	100.0 [NR]	92.0 [12.6]
2006-2008	98.6 [16.0]	100.0 [15.8]	86.3 [7.3]	100.0 [13.7]	100.0 [NR]	92.0 [12.2]
2007-2009	98.6 [15.6]	100.0 [13.8]	83.1 [5.7]	100.0 [13.6]	100.0 [NR]	90.2 [11.1]
2008-2010	98.6 [15.5]	100.0 [13.1]	83.9 [5.6]	100.0 [13.7]	100.0 [NR]	90.7 [10.9]
2009-2011	98.6 [16.1]	100.0 [13.1]	86.3 [6.0]	100.0 [14.7]	100.0 [NR]	92.0 [11.5]
2010-2012	98.6 [16.0]	100.0 [13.1]	85.5 [5.8]	100.0 [14.2]	100.0 [NR]	91.6 [11.3]
2011-2013	98.6 [16.1]	100.0 [13.8]	87.1 [6.0]	100.0 [14.4]	100.0 [NR]	92.4 [11.5]
2012-2014	98.6 [15.6]	100.0 [13.4]	86.3 [5.9]	100.0 [13.8]	100.0 [NR]	92.0 [11.2]
2013-2015	98.6 [15.6]	100.0 [14.2]	86.3 [6.2]	100.0 [13.3]	100.0 [NR]	92.0 [11.4]
2014-2016	98.6 [15.5]	100.0 [14.4]	84.7 [5.8]	100.0 [12.6]	100.0 [NR]	91.1 [11.1]
2015-2017	98.6 [14.6]	100.0 [13.3]	79.8 [5.1]	100.0 [12.4]	100.0 [NR]	88.4 [10.3]
2016-2018	98.6 [15.0]	100.0 [12.6]	81.5 [4.8]	100.0 [14.0]	100.0 [NR]	89.3 [10.4]
2017-2019	98.6 [15.5]	100.0 [12.1]	80.6 [4.9]	100.0 [15.6]	100.0 [NR]	88.9 [10.7]
2018-2020	98.6 [15.0]	100.0 [12.1]	80.6 [4.9]	100.0 [15.4]	100.0 [NR]	88.9 [10.5]
2019-2021	98.6 [12.9]	100.0 [10.5]	76.6 [4.2]	100.0 [13.2]	100.0 [NR]	86.7 [9.0]
Reduction 2003-2020	0.0 [4.5]	0.0 [4.8]	8.9 [1.6]	0.0 [-0.1]	0.0 [NR]	4.9 [3.0]

* Some sites cross the England/Wales border. No SPAs cross the England/Scotland border. For Excess Nitrogen calculations, each 1 x 1 km square (or part thereof) within each border site was assigned to a single country, so results are calculated for individual countries only.

Table 3.7: Trends in nutrient nitrogen exceedances for SSSIs: a) % of sites with SRCL with exceedance of Site-Relevant Critical Load for at least one feature; b) Excess Nitrogen [maximum AAE kg N ha⁻¹ year⁻¹]. NR = Not recorded.

Years	England	Wales	Scotland	NI	UK
Number of sites with SRCL	2954	686	938	188	4766
2002-2004	99.4 [18.2]	99.9 [15.9]	93.1 [6.0]	100.0 [14.6]	98.2 [12.7]
2003-2005	99.4 [17.9]	99.9 [15.6]	92.0 [5.9]	100.0 [13.8]	98.0 [12.5]
2004-2006	99.3 [17.9]	99.9 [16.7]	96.4 [8.1]	100.0 [14.3]	98.8 [13.6]
2005-2007	99.3 [17.6]	99.9 [17.3]	97.3 [8.4]	100.0 [15.3]	99.0 [13.6]
2006-2008	99.3 [17.2]	99.9 [17.4]	97.5 [8.0]	100.0 [15.2]	99.0 [13.3]
2007-2009	99.3 [16.5]	99.9 [15.3]	93.3 [5.8]	100.0 [15.1]	98.2 [11.9]
2008-2010	99.3 [16.1]	99.9 [14.5]	92.2 [5.7]	100.0 [15.3]	98.0 [11.6]
2009-2011	99.4 [16.9]	99.9 [14.5]	94.5 [6.1]	100.0 [16.4]	98.5 [12.1]
2010-2012	99.4 [16.7]	99.9 [14.6]	92.9 [5.9]	100.0 [15.8]	98.2 [12.0]
2011-2013	99.4 [17.1]	100.0 [15.4]	95.5 [6.4]	100.0 [16.1]	98.7 [12.4]
2012-2014	99.4 [16.4]	99.9 [14.9]	93.7 [6.1]	100.0 [15.3]	98.3 [11.9]
2013-2015	99.4 [16.6]	100.0 [15.7]	94.6 [6.8]	100.0 [14.8]	98.5 [12.4]
2014-2016	99.3 [16.3]	100.0 [15.8]	92.0 [6.2]	100.0 [14.0]	98.0 [12.0]
2015-2017	99.3 [15.3]	99.9 [14.7]	90.0 [5.3]	100.0 [13.6]	97.5 [11.0]
2016-2018	99.3 [15.4]	99.9 [13.8]	88.4 [4.5]	100.0 [15.2]	97.3 [10.7]
2017-2019	99.3 [15.8]	99.9 [13.1]	89.7 [4.4]	100.0 [16.9]	97.5 [10.8]
2018-2020	99.3 [15.4]	99.9 [13.3]	89.2 [4.3]	100.0 [16.9]	97.4 [10.6]
2019-2021	99.1 [13.4]	99.4 [12.1]	85.3 [3.7]	99.5 [14.7]	96.4 [9.3]
Reduction 2003-2020	0.3 [4.8]	0.4 [3.8]	7.8 [2.3]	0.5 [-0.1]	1.8 [3.5]

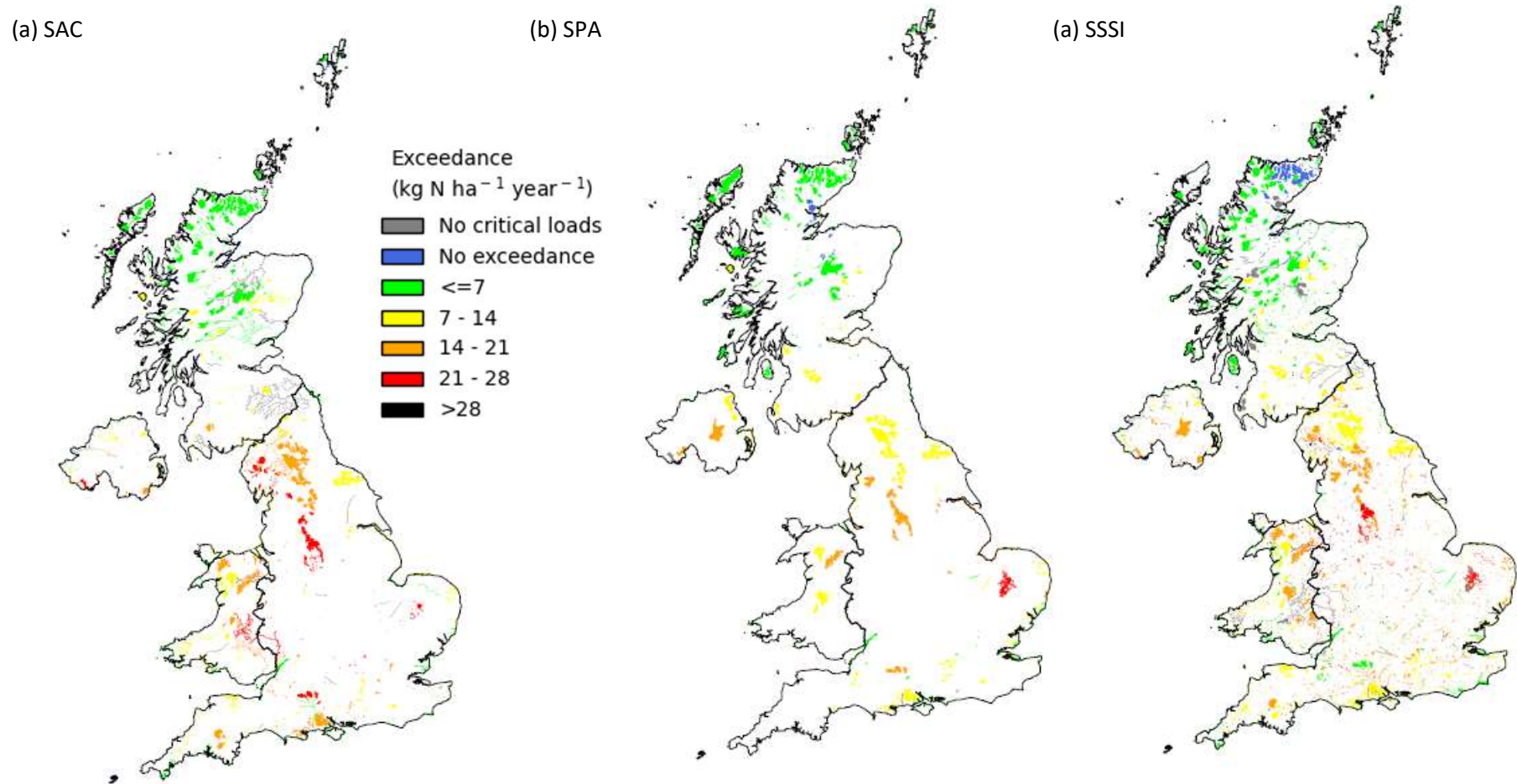


Figure 3.2: Excess Nitrogen: Average Accumulated Exceedance (AAE) of nutrient nitrogen critical loads by CBED deposition for 2019-21. Maps show the maximum AAE for any feature within each site. Other features may have lower or no exceedance. Note that small sites may not show at this resolution.

Section 4: Critical levels and their exceedances

This section of the report focuses on the trends in exceedance of critical levels for ammonia, that are set to protect: a) lichens and bryophytes; or b) higher (i.e. vascular) plants. The trends are based on rolling 3-year mean ammonia gaseous concentrations. For brevity, the three-year means are mainly referred to in this report using the middle year, for example “2010 to 2015” equates to “2009-11 to 2014-16”.

4.1 Critical levels of ammonia

The critical levels considered here are annual mean concentrations of ammonia above which direct adverse effects on sensitive vegetation may occur according to present knowledge (CLRTAP, 2017). Critical levels have also been defined for other pollutants: sulphur dioxide, some N oxides, and ozone. Critical levels based on mean concentrations over shorter time periods have also been defined. These other pollutants and timescales are not considered here; further information can be found in (CLRTAP, 2017).

The critical levels for ammonia were reviewed and updated at an international workshop held in 2006 (UNECE, 2007) and approved by the Task Forces of the International Cooperative Programmes (ICPs) of the CLRTAP. Critical levels are only defined for two taxonomic groups (Table 4.1), in contrast to the range of habitats for which critical loads have been defined. This means that critical levels of ammonia have not here been applied to individual habitats or to habitat features of designated sites in the UK. Critical level exceedance metrics used are described in Section 4.3.

Table 4.1: Critical levels of ammonia (CLRTAP, 2014)

Vegetation type	Critical level NH ₃ [µg m ⁻³]	Time period
Lichens and bryophytes (including ecosystems where lichens and bryophytes are a key part of the ecosystem integrity)	1	Annual mean concentration
Higher plants (including heathland, semi-natural grassland and forest ground flora)	3*	Annual mean concentration

*An explicit uncertainty range of 2-4 µg m⁻³ was set for higher plants; this was intended to be used when applying the critical level in different assessment contexts.

4.2 Concentrations of ammonia

The spatial distribution of ammonia is calculated using a process model, with results calibrated to observations. As a relatively short-lived primary pollutant, ammonia (NH₃) exhibits greater spatial variability than can reasonably be captured by national-scale networks, and a simple interpolation between existing monitoring sites would not give accurate concentration estimates. To account for this high spatial variability, we deploy an atmospheric chemistry and transport model (EMEP4UK) to simulate the spatial distribution of ammonia concentration, based on current understanding of emissions, chemical transformations and fate. EMEP4UK is a high-spatial-resolution UK implementation (EMEP4UK; Vieno et al., 2016) of the European Monitoring and Evaluation Programme (EMEP) model (EMEP MSC-W - <https://github.com/metno/emep-ctm>), version 4.36 (Simpson et al., 2012). Process modelling has associated uncertainties, so EMEP4UK outputs are

calibrated to measured concentrations from the UKEAP National Ammonia Monitoring Network (Stephens et al., 2021), using a proportional relationship to scale the modelled concentrations according to the measurements. Data from all monitoring stations that have a temporal coverage of measurements data for at least 70% of the year were used for calibration, with the exception of a few stations where the measurements were not representative of the surrounding area as modelled for the grid cell. Individual years were also removed from the calibration dataset where the annual mean deviated by more than approximately three standard deviations from the long-term mean for the site.

Whilst only the ammonia output is used in the calculation of ecosystem impacts, EMEP4UK requires annual total emissions for NO_x, SO_x, fine and coarse primary particulate matter (PM_{2.5} and PM₁₀), CO, and non-methane volatile organic carbon, as well as NH₃ emissions. These are provided as both diffuse area emissions (on a 1 km x 1 km grid) and as point source emissions. The EMEP4UK model uses meteorological data generated by the Weather Forecasting and Research model (WRF version 4.2.2; Skamarock, 2019), constrained with 6-hourly reanalysis data (NCEP, 2000).

Maps of diffuse-area emissions for the years 2002-2019 were obtained by rescaling the UK National Atmospheric Emissions Inventory (NAEI) 2019 emission map using UK total emissions for the year, at the sectoral level. Point-source emissions were obtained from the specific year's data and combined with re-scaled diffuse emissions, at the sectoral level. At the time of analysis, emissions data were not yet available for 2020 (publication delays due to the Covid-19 pandemic) and 2021 (due autumn 2023), so 2019 emissions were used. Although emissions data are not available for recent years, meteorological data are available up until close to the present day. This allows estimation of ammonia concentrations for recent years, since meteorology has a large influence on atmospheric processes. In Trends Reports up until 2021, ammonia concentrations were calculated using the FRAME model (Dore et al., 2007), which does not allow for the effects of meteorology. The EMEP4UK model does incorporate meteorology, and for this reason we have sufficient confidence in the outputs to present ammonia concentration data closer to the present day than in previous Trends Reports. There is more uncertainty associated with the results for 2020 and 2021, because the spatial patterns of emissions have not been updated, but calibration to observed ammonia concentrations in those years means that these results are reasonably reliable.

The EMEP4UK model has the potential to be run at 1 km x 1 km resolution, but is currently run at a 3 km x 3 km resolution, to reduce computation time and data storage requirements. The UK run is nested into a coarser resolution European run, using EMEP emissions provided for 2002-2019 at a resolution of 0.1 ° x 0.1 ° (approximately 11 km x 6 km, at UK latitudes)¹, together with higher-resolution data for the Republic of Ireland².

¹ <https://www.emep.int/>; <https://www.ceip.at/webdab-emission-database>

² https://www.epa.ie/publications/research/air/Research_Report_317.pdf

4.3 Calculation of critical levels exceedance

The critical level exceedance metrics calculated for this Trends Report are:

- The percentage land area in England, Wales, Scotland, Northern Ireland and UK where ammonia concentrations exceed the critical levels.
- The percentage area of N-sensitive habitats in England, Wales, Scotland, Northern Ireland and UK where ammonia concentrations exceed the critical levels. The habitat areas are based on the habitat distribution maps used for mapping nutrient N critical loads (Section 1.1).
- The percentage of designated sites (SAC, SPA, SSSI) in England, Wales, Scotland, Northern Ireland and the UK, where ammonia concentrations exceed the critical levels anywhere across a site. Ammonia-sensitivity has not yet been determined for the current set of designated sites, so critical level exceedance was calculated for all sites.

4.4 Trends in ammonia critical levels exceedance

4.4.1 UK land area with exceedance of ammonia critical levels

Ammonia concentrations exceeded the critical level of $1 \mu\text{g m}^{-3}$ (set to protect sensitive bryophytes and lichens) across 64.3% of the UK land area in 2020, compared to 56.7% in 2003 (Figure 4.1a; Table 4.2). Exceedance varies spatially, with minimum 15.8% of Scotland, but more than 90% of England and Northern Ireland having ammonia concentrations above $1 \mu\text{g m}^{-3}$.

The ammonia critical level of $3 \mu\text{g m}^{-3}$, set to protect sensitive vascular plants, was not exceeded in 2020 in Scotland and relatively small areas of the rest of the UK: 0.7 % for Wales, 3.0 % for England, and 16.7% for NI. (Figure 4.1b; Table 4.2).

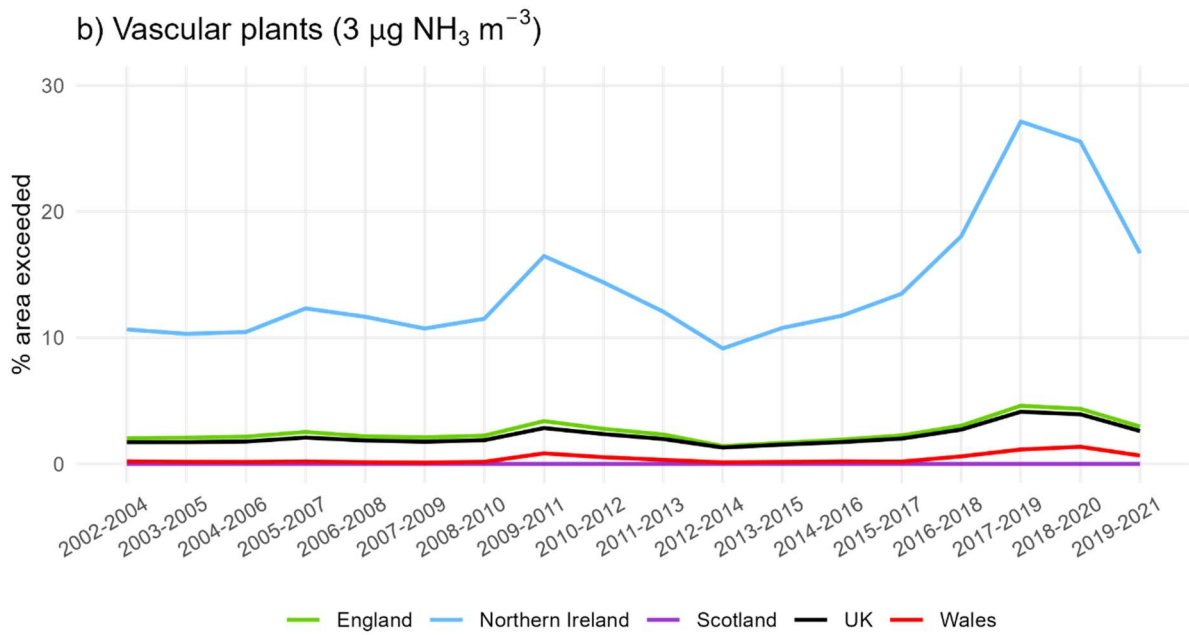
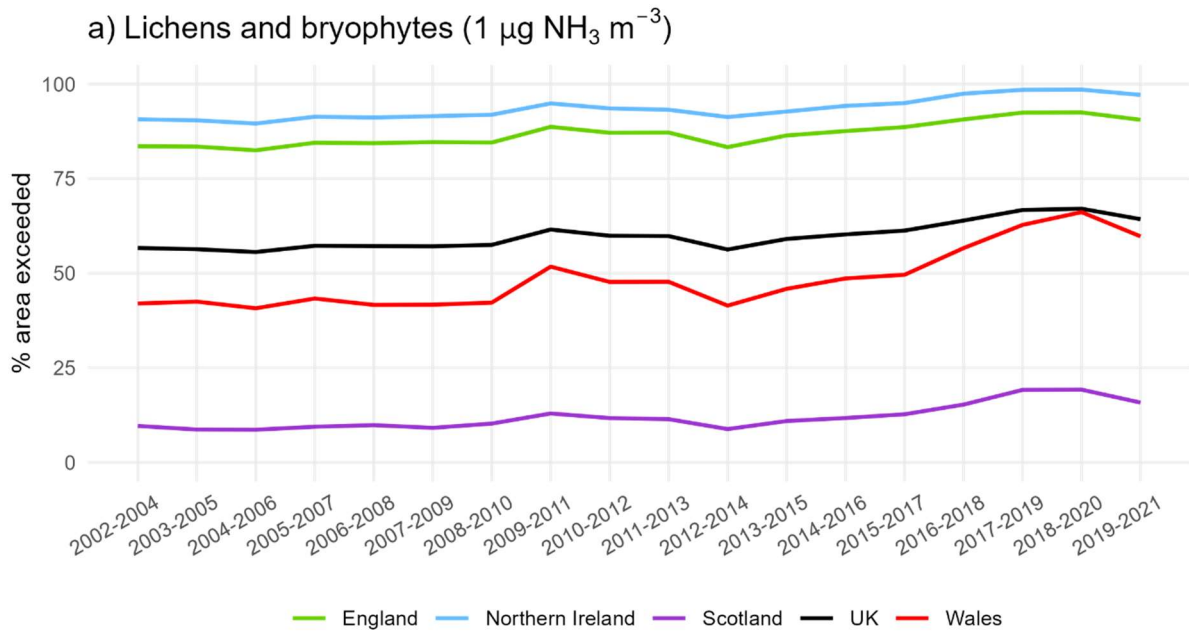


Figure 4.1: Percentages of the UK land area where ammonia concentrations exceeded critical levels: for a) lichens and bryophytes; b) vascular plants.

Table 4.2a: Percentages of the UK land area where ammonia concentrations exceed the critical level set to protect sensitive bryophytes and lichens, 1 µg m⁻³.

Years	England	Wales	Scotland	NI	UK
2002-2004	83.6	42.0	9.6	90.7	56.7
2003-2005	83.5	42.5	8.7	90.4	56.3
2004-2006	82.5	40.7	8.6	89.6	55.6
2005-2007	84.5	43.3	9.4	91.4	57.2
2006-2008	84.4	41.6	9.8	91.1	57.2
2007-2009	84.6	41.7	9.1	91.5	57.1
2008-2010	84.5	42.2	10.3	91.9	57.5
2009-2011	88.7	51.7	12.9	94.9	61.5
2010-2012	87.1	47.7	11.7	93.6	59.9
2011-2013	87.2	47.7	11.4	93.2	59.8
2012-2014	83.3	41.4	8.8	91.3	56.3
2013-2015	86.4	45.9	10.9	92.7	59.1
2014-2016	87.6	48.6	11.7	94.2	60.3
2015-2017	88.6	49.6	12.7	95.0	61.3
2016-2018	90.6	56.6	15.3	97.4	63.9
2017-2019	92.4	62.8	19.2	98.5	66.7
2018-2020	92.5	66.1	19.2	98.5	67.0
2019-2021	90.5	59.7	15.8	97.1	64.3
Change from 2003-2020	7	17.7	6.2	6.5	7.6

Table 4.2b: Percentages of the UK land area where ammonia concentrations exceed the critical level set to protect sensitive vascular plants, 3 $\mu\text{g m}^{-3}$.

Years	England	Wales	Scotland	NI	UK
2002-2004	2.0	0.2	0.0	10.7	1.7
2003-2005	2.1	0.2	0.0	10.3	1.7
2004-2006	2.2	0.2	0.0	10.5	1.8
2005-2007	2.5	0.2	0.0	12.3	2.1
2006-2008	2.2	0.1	0.0	11.7	1.9
2007-2009	2.1	0.1	0.0	10.7	1.8
2008-2010	2.2	0.2	0.0	11.5	1.9
2009-2011	3.4	0.8	0.0	16.5	2.8
2010-2012	2.8	0.5	0.0	14.4	2.4
2011-2013	2.3	0.3	0.0	12.1	2.0
2012-2014	1.4	0.1	0.0	9.1	1.3
2013-2015	1.7	0.2	0.0	10.8	1.5
2014-2016	1.9	0.2	0.0	11.8	1.7
2015-2017	2.3	0.2	0.0	13.5	2.0
2016-2018	3.0	0.6	0.0	18.0	2.7
2017-2019	4.6	1.1	0.0	27.1	4.1
2018-2020	4.4	1.4	0.0	25.5	3.9
2019-2021	3.0	0.7	0.0	16.7	2.6
Change from 2003-2020	0.9	0.5	0.0	6.0	0.9

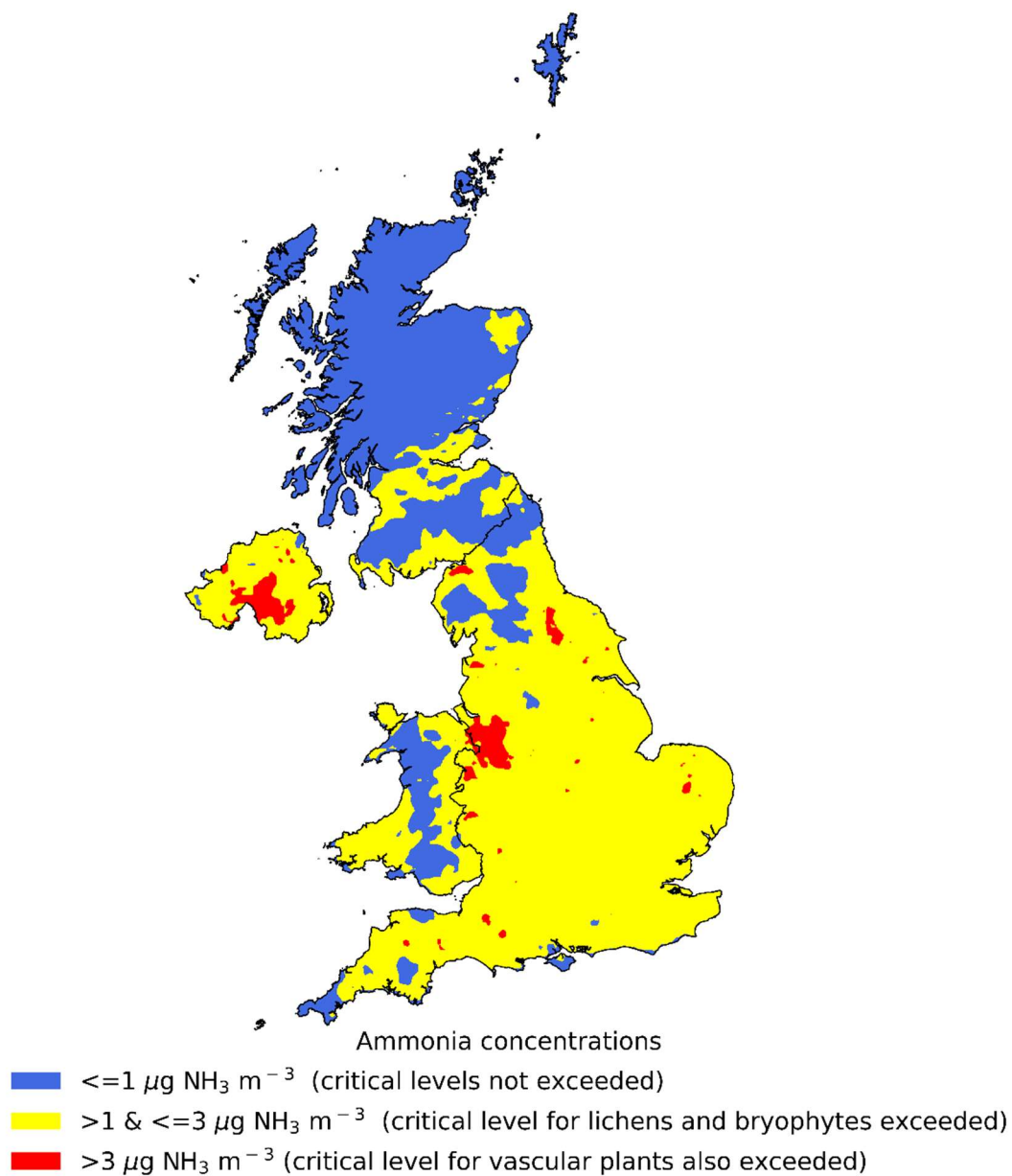


Figure 4.1: CBED 1 km x 1 km mean ammonia concentrations for 2019-21.

4.4.2 Nitrogen-sensitive habitats with exceedance of ammonia critical levels

Around a third of the mapped area of N-sensitive habitats in the UK receives ammonia concentrations above the critical level of $1 \mu\text{g m}^{-3}$ (Table 4.3). The results vary spatially across the UK depending on the variability in ammonia concentrations (see Figure 4.1), and the distributions of the different N-sensitive habitats. Although 58% of the total UK area of N-sensitive habitats is found in Scotland, because the ammonia concentrations are generally low in this part of the country (Figure 4.1) only 5.1% of the Scottish habitat area coincides with ammonia concentrations above $1 \mu\text{g m}^{-3}$ and there were no areas in Scotland with concentrations above $3 \mu\text{g m}^{-3}$ in 2020 (Table 4.3). The highest exceedances are seen in England and Northern Ireland, with ammonia concentrations above $1 \mu\text{g m}^{-3}$ for 70.7-93.2% of their N-sensitive habitat areas and above $3 \mu\text{g m}^{-3}$ for 1.0 - 5.7%. The percentage area of N-sensitive habitats in the UK with exceedance of the critical level of $1 \mu\text{g m}^{-3}$ increased from 22.4% in 2003 to 30.1% in 2020.

Table 4.3a: Percentages of the area of nitrogen sensitive habitats in the UK where ammonia concentrations exceed the critical level set to protect sensitive bryophytes and lichens, $1 \mu\text{g m}^{-3}$, by country.

Years	England	Wales	Scotland	NI	UK
N-sensitive habitat area (km ²)	25813	9347	53530	3898	92588
2002-2004	56.2	19.4	2.7	76.1	22.4
2003-2005	56.5	19.7	2.4	75.6	22.3
2004-2006	55.3	18.4	2.4	73.7	21.7
2005-2007	58.6	20.0	2.6	78.0	23.1
2006-2008	58.3	18.8	2.8	77.3	23.0
2007-2009	58.9	18.8	2.6	78.2	23.1
2008-2010	58.7	19.6	2.9	79.0	23.3
2009-2011	66.5	27.5	3.8	86.8	27.2
2010-2012	63.2	23.6	3.4	83.1	25.5
2011-2013	63.2	23.6	3.3	82.3	25.4
2012-2014	56.4	18.8	2.5	77.2	22.3
2013-2015	61.5	22.2	3.1	81.1	24.6
2014-2016	63.8	24.4	3.4	85.4	25.8
2015-2017	66.1	25.0	3.8	87.5	26.8
2016-2018	70.7	31.8	4.8	94.1	29.7
2017-2019	75.2	38.4	6.7	97.3	32.8
2018-2020	75.4	42.6	6.7	97.6	33.3
2019-2021	70.7	35.2	5.1	93.2	30.1
Change from 2003-2020	14.5	15.8	2.4	17	7.7

Table 4.3b: Percentages of the area of nitrogen sensitive habitats in the UK where ammonia concentrations exceed the critical level set to protect sensitive vascular plants, $3 \mu\text{g m}^{-3}$, by country.

Years	England	Wales	Scotland	NI	UK
N-sensitive habitat area (km ²)	25813	9347	53530	3898	92588
2002-2004	0.6	0.0	0.0	3.6	0.3
2003-2005	0.7	0.0	0.0	3.4	0.3
2004-2006	0.7	0.0	0.0	3.4	0.3
2005-2007	0.9	0.0	0.0	4.1	0.4
2006-2008	0.7	0.0	0.0	3.8	0.4
2007-2009	0.7	0.0	0.0	3.5	0.3
2008-2010	0.7	0.0	0.0	3.8	0.4
2009-2011	1.2	0.1	0.0	5.6	0.6
2010-2012	1.0	0.1	0.0	5.0	0.5
2011-2013	0.8	0.1	0.0	4.2	0.4
2012-2014	0.4	0.0	0.0	3.1	0.3
2013-2015	0.5	0.0	0.0	3.7	0.3
2014-2016	0.6	0.0	0.0	4.1	0.3
2015-2017	0.7	0.0	0.0	4.5	0.4
2016-2018	1.0	0.1	0.0	6.1	0.5
2017-2019	1.6	0.2	0.0	9.8	0.9
2018-2020	1.5	0.2	0.0	9.1	0.8
2019-2021	1.0	0.1	0.0	5.7	0.5
Change from 2003-2020	0.3	0.1	0	2.2	0.2

The N-sensitive habitats (Table 4.4) with the highest percentage area of exceedance of the ammonia critical level of $1 \mu\text{g m}^{-3}$ in 2020 were beech woodland (92.9%), calcareous grassland (84.7%), broadleaved woodland (84.2%), and mixed woodland (78.1%). Other habitats had smaller percentage areas where ammonia concentrations were above $1 \mu\text{g m}^{-3}$. However, although only 13.2% of Dwarf shrub heath is exceeded, this habitat occupies a large area across the UK so this equates to 2,822 km², which is more than 92.2% area of beech woodland exceeded (1,909 km²).

Differences between years were small, and reflect fluctuations in ammonia concentrations due to inter-annual variability in meteorology. However, there was no evidence of decreases in the area where critical levels are exceeded.

Table 4.4a: Percentages of the area of nitrogen-sensitive habitats where the 1 µg m⁻³ ammonia critical level set to protect sensitive bryophytes and lichens is exceeded, by habitat.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved woodland	Beech woodland	Acidophilous oak	Scots Pine	Mixed woodland	Dune grassland	Saltmarsh
Habitat area (km ²)	20339	994	21385	9070	4915	14401	8616	2055	6851	1484	1409	426	642
2002-2004	8.4	72.6	7.3	8.5	0.0	21.0	77.4	74.6	48.1	0.9	65.8	23.9	61.5
2003-2005	8.4	71.8	7.4	8.5	0.0	20.7	76.9	75.8	48.1	0.9	65.3	21.4	60.4
2004-2006	8.1	69.6	7.3	8.3	0.0	20.2	76.2	71.3	45.9	0.9	64.0	20.7	59.2
2005-2007	8.9	73.3	8.2	9.4	0.1	21.7	77.7	78.3	49.9	0.9	66.8	23.1	62.5
2006-2008	8.7	74.0	8.1	9.2	0.1	21.6	77.6	76.9	49.2	1.1	66.4	23.5	62.5
2007-2009	8.8	74.7	8.4	9.4	0.1	21.7	77.1	77.9	49.3	0.9	66.4	24.8	62.5
2008-2010	9.4	73.8	8.5	9.7	0.1	21.9	77.8	75.9	49.3	1.2	67.1	25.1	62.3
2009-2011	13.2	81.1	11.2	11.9	0.1	25.8	81.8	89.2	57.7	1.3	73.8	35.6	69.4
2010-2012	11.3	78.5	9.8	10.8	0.1	24.1	80.1	84.2	54.5	1.3	71.3	31.5	67.1
2011-2013	11.1	78.8	9.7	10.7	0.0	24.1	80.0	85.0	54.6	1.3	70.9	32.1	67.7
2012-2014	8.6	71.7	7.6	9.1	0.0	20.9	76.7	73.0	46.8	0.9	65.5	25.7	62.2
2013-2015	10.4	77.8	9.1	10.1	0.0	23.3	79.2	81.9	52.7	1.2	69.3	31.2	66.8
2014-2016	11.8	79.9	10.0	11.1	0.1	24.5	80.3	85.5	55.1	1.2	71.6	34.8	68.9
2015-2017	12.6	80.5	11.0	11.9	0.1	25.6	81.3	88.1	56.8	1.3	72.9	35.8	69.8
2016-2018	16.1	84.0	13.1	14.5	0.2	28.6	84.3	93.3	61.1	1.4	77.4	41.8	73.0
2017-2019	20.7	88.2	14.5	18.1	0.3	32.8	87.1	95.1	65.2	1.9	81.7	47.0	76.2
2018-2020	21.9	88.9	14.6	18.3	0.3	33.3	87.3	95.7	66.3	2.0	82.1	47.7	76.5
2019-2021	17.2	84.7	13.2	14.7	0.1	29.2	84.2	92.9	62.1	1.6	78.1	40.6	72.7
Change from 2003-2020	8.8	12	5.8	6.2	0.1	8.2	6.8	18.3	14	0.7	12.3	16.7	11.2

Table 4.4b: Percentages of the area of nitrogen-sensitive habitats where the 3 µg m⁻³ ammonia critical level set to protect sensitive vascular plants is exceeded, by habitat.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved woodland	Beech woodland	Acidophilous oak	Scots Pine	Mixed woodland	Dune grassland	Saltmarsh
Habitat area (km ²)	20339	994	21385	9070	4915	14401	8616	2055	6851	1484	1409	426	642
2002-2004	0.0	0.0	0.0	0.2	0.0	0.2	2.0	0.1	0.6	0.0	2.6	0.0	1.3
2003-2005	0.0	0.1	0.0	0.1	0.0	0.2	2.0	0.3	0.6	0.0	2.7	0.0	1.3
2004-2006	0.0	0.1	0.0	0.1	0.0	0.2	2.0	0.4	0.7	0.0	2.7	0.0	1.3
2005-2007	0.0	0.1	0.0	0.2	0.0	0.3	2.4	0.4	0.8	0.0	3.1	0.0	1.3
2006-2008	0.0	0.1	0.0	0.2	0.0	0.2	2.2	0.4	0.7	0.0	2.8	0.0	1.3
2007-2009	0.0	0.1	0.0	0.1	0.0	0.2	2.1	0.4	0.6	0.0	2.6	0.0	1.2
2008-2010	0.0	0.1	0.0	0.2	0.0	0.2	2.2	0.4	0.7	0.0	2.8	0.0	1.3
2009-2011	0.0	0.1	0.0	0.3	0.0	0.4	3.3	0.6	1.1	0.0	4.1	0.2	1.5
2010-2012	0.0	0.1	0.0	0.3	0.0	0.3	2.7	0.6	1.0	0.0	3.6	0.1	1.4
2011-2013	0.0	0.1	0.0	0.2	0.0	0.3	2.2	0.5	0.7	0.0	3.1	0.0	1.3
2012-2014	0.0	0.1	0.0	0.1	0.0	0.1	1.5	0.3	0.3	0.0	2.2	0.0	1.0
2013-2015	0.0	0.1	0.0	0.2	0.0	0.2	1.8	0.2	0.4	0.0	2.5	0.0	0.8
2014-2016	0.0	0.1	0.0	0.2	0.0	0.2	2.0	0.1	0.5	0.0	2.8	0.0	0.8
2015-2017	0.0	0.1	0.0	0.3	0.0	0.3	2.3	0.1	0.7	0.0	3.1	0.0	0.9
2016-2018	0.0	0.1	0.1	0.4	0.0	0.4	3.1	0.3	0.9	0.0	4.0	0.0	1.1
2017-2019	0.0	0.1	0.2	0.7	0.0	0.6	4.8	0.9	1.4	0.0	6.0	0.2	2.1
2018-2020	0.0	0.1	0.1	0.6	0.0	0.6	4.5	0.8	1.4	0.0	5.8	0.2	2.0
2019-2021	0.0	0.1	0.0	0.4	0.0	0.3	2.9	0.3	1.0	0.0	4.0	0.0	1.2
Change from 2003-2020	0	0.1	0	0.2	0	0.1	1	0.3	0.4	0	1.3	0	0

4.4.3 Designated sites with exceedance of ammonia critical levels

These results show the percentage of sites (SACs, SPAs, SSSIs) where ammonia critical levels are exceeded; a site is counted as exceeded if the ammonia concentration exceeds the respective critical level anywhere across a site. SACs may contain one or more SSSIs, and some SACs and SPAs may overlap one another (Figure 4.2), however in this analysis the sites are all assessed independently.

SACs

59.8% of SACs occur in areas of the UK where ammonia concentrations exceed $1 \mu\text{g m}^{-3}$ (Table 4.5); the lowest number of sites with exceedance is in Scotland, where ammonia concentrations are $< 1 \mu\text{g m}^{-3}$ at 14.4% of sites. The percentage of UK SACs with exceedance of the $1 \mu\text{g m}^{-3}$ critical level increased by 10.6% between 2003 and 2020, but this varies spatially across the UK with more increase seen in Wales, than in England and Scotland over this time period.

The percentage of SACs with ammonia concentrations above $3 \mu\text{g m}^{-3}$ is smaller, $<15\%$ in 2020 (Table 4.5). However, in Northern Ireland there has been a steady increase in the number of SACs where this higher critical level is exceeded, with $\sim 14.8\%$ of sites now exceeded.

Table 4.5a: Percentage of SACs where the $1 \mu\text{g m}^{-3}$ ammonia critical level set to protect sensitive bryophytes and lichens is exceeded, by country.

Years	England	Wales	Scotland	NI	England/ Wales	England/ Scotland	UK
Number of sites	231	84	236	54	7	3	615
2002-2004	80.1	52.4	5.9	94.4	100.0	66.7	49.3
2003-2005	78.8	52.4	5.9	94.4	100.0	66.7	48.8
2004-2006	76.6	50.0	5.9	92.6	100.0	66.7	47.5
2005-2007	81.8	53.6	6.8	96.3	100.0	66.7	50.6
2006-2008	82.3	52.4	6.8	96.3	100.0	66.7	50.6
2007-2009	82.3	52.4	6.4	96.3	100.0	66.7	50.4
2008-2010	81.8	54.8	6.8	96.3	100.0	66.7	50.7
2009-2011	88.3	63.1	11.4	100.0	100.0	100.0	56.6
2010-2012	87.4	58.3	9.7	100.0	100.0	100.0	55.0
2011-2013	87.4	60.7	9.3	100.0	100.0	100.0	55.1
2012-2014	79.7	54.8	6.4	96.3	100.0	66.7	49.8
2013-2015	86.6	56.0	8.1	96.3	100.0	100.0	53.3
2014-2016	87.0	59.5	9.7	100.0	100.0	100.0	55.0
2015-2017	88.3	61.9	11.0	100.0	100.0	100.0	56.3
2016-2018	90.9	69.0	15.3	100.0	100.0	100.0	59.8
2017-2019	92.2	75.0	16.9	100.0	100.0	100.0	61.8
2018-2020	92.2	78.6	16.9	100.0	100.0	100.0	62.3
2019-2021	89.6	75.0	14.4	100.0	100.0	100.0	59.8
Change from 2003-2020	9.5	22.6	8.5	5.6	0	33.3	10.6

*includes 7 sites on the England/Wales border and 3 on the England/Scotland border.

Table 4.5b: Percentage of SACs where the 3 µg m⁻³ ammonia critical level set to protect sensitive vascular plants is exceeded, by country.

Years	England	Wales	Scotland	NI	England/ Wales	England/ Scotland	UK
Number of sites	231	84	236	54	7	3	615
2002-2004	2.2	0.0	0.0	7.4	28.6	0.0	1.8
2003-2005	2.2	0.0	0.0	7.4	28.6	0.0	1.8
2004-2006	2.6	0.0	0.0	7.4	28.6	0.0	2.0
2005-2007	3.5	0.0	0.0	7.4	28.6	0.0	2.3
2006-2008	3.0	0.0	0.0	7.4	28.6	0.0	2.1
2007-2009	1.7	0.0	0.0	7.4	28.6	0.0	1.6
2008-2010	2.6	0.0	0.0	7.4	28.6	0.0	2.0
2009-2011	4.3	1.2	0.0	14.8	28.6	0.0	3.4
2010-2012	3.0	0.0	0.0	14.8	28.6	0.0	2.8
2011-2013	3.0	0.0	0.0	9.3	28.6	0.0	2.3
2012-2014	2.2	0.0	0.0	5.6	28.6	0.0	1.6
2013-2015	2.6	0.0	0.0	7.4	28.6	0.0	2.0
2014-2016	2.6	0.0	0.0	9.3	28.6	0.0	2.1
2015-2017	2.6	0.0	0.0	9.3	28.6	0.0	2.1
2016-2018	3.9	0.0	0.0	14.8	28.6	0.0	3.1
2017-2019	5.2	1.2	0.0	18.5	28.6	33.3	4.2
2018-2020	4.8	1.2	0.0	18.5	28.6	33.3	4.1
2019-2021	3.5	1.2	0.0	14.8	28.6	33.3	3.3
Change from 2003-2020	1.3	1.2	0	7.4	0	33.3	1.5

*includes 7 sites on the England/Wales border and 3 on the England/Scotland border.

SPAs

The results show a 40% increase from 2003 to 2020 in the number of sites in Wales exceeding the critical level of 1 µg m⁻³, although it should be noted there are only 15 SPAs in Wales. There was a small increase in the percentage of exceeded sites in England (1.3%), Scotland (3.6%) and Northern Ireland (7.1%) (Table 4.6). Over 93% of the SPAs in England and all 14 sites in Northern Ireland are exposed to ammonia concentrations above 1 µg m⁻³.

The critical level of 3 µg m⁻³ is not exceeded for any SPAs in Wales and Scotland; this is consistent with the fact that <1% of the land area has ammonia concentrations above 3 µg m⁻³ (Table 4.6). Overall, the percentage of sites with ammonia concentrations above 3 µg m⁻³ is considerably smaller than the number exceeding the critical level of 1 µg m⁻³. In Northern Ireland, two SPAs showed exceedance of the ammonia critical level of 3 µg m⁻³, representing roughly a 14% increase since 2003 of the percentage of sites exceeded, from 0% to 14.3%.

Table 4.6a: Percentage of SPAs where the 1 µg m⁻³ ammonia critical level set to protect sensitive vascular plants is exceeded, by country.

Years	England	Wales	Scotland	NI	England/ Wales	UK
Number of sites	77	15	139	14	3	248
2002-2004	92.2	20.0	8.6	92.9	100.0	41.1
2003-2005	90.9	20.0	6.5	92.9	100.0	39.5
2004-2006	88.3	20.0	6.5	92.9	100.0	38.7
2005-2007	92.2	26.7	9.4	92.9	100.0	41.9
2006-2008	92.2	26.7	9.4	92.9	100.0	41.9
2007-2009	93.5	26.7	8.6	92.9	100.0	41.9
2008-2010	90.9	26.7	9.4	92.9	100.0	41.5
2009-2011	93.5	33.3	11.5	100.0	100.0	44.4
2010-2012	93.5	33.3	10.8	100.0	100.0	44.0
2011-2013	93.5	33.3	10.1	100.0	100.0	43.5
2012-2014	89.6	26.7	8.6	92.9	100.0	40.7
2013-2015	93.5	33.3	10.1	92.9	100.0	43.1
2014-2016	93.5	33.3	10.8	100.0	100.0	44.0
2015-2017	93.5	33.3	10.8	100.0	100.0	44.0
2016-2018	93.5	53.3	13.7	100.0	100.0	46.8
2017-2019	96.1	60.0	16.5	100.0	100.0	49.6
2018-2020	94.8	60.0	16.5	100.0	100.0	49.2
2019-2021	93.5	60.0	12.2	100.0	100.0	46.4
Change from 2003-2020	1.3	40	3.6	7.1	0	5.2

*includes 3 sites on the England/Wales border.

Table 4.6b: Percentage of SPAs where the 3 µg m⁻³ ammonia critical level set to protect sensitive bryophytes and lichens is exceeded, by country.

Years	England	Wales	Scotland	NI	England/ Wales	UK
Number of sites	77	15	139	14	3	248
2002-2004	2.6	0.0	0.0	0.0	0.0	0.8
2003-2005	2.6	0.0	0.0	0.0	0.0	0.8
2004-2006	2.6	0.0	0.0	0.0	0.0	0.8
2005-2007	3.9	0.0	0.0	7.1	0.0	1.6
2006-2008	3.9	0.0	0.0	0.0	0.0	1.2
2007-2009	3.9	0.0	0.0	0.0	0.0	1.2
2008-2010	3.9	0.0	0.0	0.0	0.0	1.2
2009-2011	5.2	0.0	0.0	14.3	0.0	2.4
2010-2012	3.9	0.0	0.0	14.3	0.0	2.0
2011-2013	3.9	0.0	0.0	0.0	0.0	1.2
2012-2014	2.6	0.0	0.0	0.0	0.0	0.8
2013-2015	3.9	0.0	0.0	0.0	0.0	1.2
2014-2016	3.9	0.0	0.0	0.0	0.0	1.2
2015-2017	3.9	0.0	0.0	7.1	0.0	1.6
2016-2018	3.9	0.0	0.0	14.3	0.0	2.0
2017-2019	7.8	0.0	0.0	14.3	33.3	3.6
2018-2020	7.8	0.0	0.0	14.3	33.3	3.6
2019-2021	5.2	0.0	0.0	14.3	33.3	2.8
Change from 2003-2020	2.6	0	0	14.3	33.3	2

*includes 3 sites on the England/Wales border.

SSSIs

The percentage of SSSIs/ASSIs in the UK in areas where ammonia concentrations exceed the critical level of $1 \mu\text{g m}^{-3}$ increased by 9.2% between 2002 and 2020, to 71.9% (Table 4.7). Nearly 90% of the sites in England and over 98% sites in Northern Ireland are in locations where this critical level is currently exceeded, as well as 64.9% of sites in Wales and 20.5% of sites in Scotland.

There was a small increase in the number of UK SSSIs/ASSIs showing exceedance of the $3 \mu\text{g m}^{-3}$ Critical Load for ammonia, with the number of sites exceeded increasing in every country. In particular, it is worth noting that between 2003 and 2020 there has been an 4.5% increase in the percentage of ASSIs in Northern Ireland that received ammonia concentrations above the critical level of $3 \mu\text{g m}^{-3}$. England, Wales and Scotland showed marginal increases of $\leq 0.7\%$.

Table 4.7a: Percentage of SSSIs where the $1 \mu\text{g m}^{-3}$ ammonia critical level set to protect sensitive bryophytes and lichens is exceeded, by country.

Years	England	Wales	Scotland	NI	UK
Number of sites	4091	1006	1422	289	6808
2002-2004	82.8	44.9	12.2	88.6	62.7
2003-2005	82.8	45.9	10.9	87.9	62.6
2004-2006	81.4	44.2	10.5	85.8	61.3
2005-2007	83.8	46.9	12.0	90.3	63.6
2006-2008	83.5	45.5	12.3	90.3	63.3
2007-2009	83.7	45.7	11.3	91.3	63.3
2008-2010	83.5	46.0	12.4	91.3	63.4
2009-2011	87.7	56.4	16.0	96.2	68.5
2010-2012	86.3	52.4	14.7	94.8	66.7
2011-2013	86.5	52.0	14.0	93.8	66.6
2012-2014	82.3	45.4	11.0	90.0	62.3
2013-2015	85.7	49.8	13.6	92.4	65.6
2014-2016	86.7	53.4	14.5	94.8	67.0
2015-2017	87.5	54.5	15.8	95.5	68.0
2016-2018	89.8	60.6	19.8	98.6	71.2
2017-2019	91.7	66.7	25.5	99.7	74.5
2018-2020	91.7	71.0	25.2	99.7	75.1
2019-2021	89.6	64.9	20.5	98.6	71.9
Change from 2003-2020	6.8	20	8.3	10	9.2

Table 4.7b: Percentage of SSSIs where the 3 µg m⁻³ ammonia critical level set to protect sensitive vascular plants is exceeded, by country.

Years	England	Wales	Scotland	NI	UK
Number of sites	4091	1006	1422	289	6808
2002-2004	1.5	0.5	0.0	8.3	1.3
2003-2005	1.5	0.3	0.0	8.0	1.3
2004-2006	1.6	0.3	0.0	8.0	1.4
2005-2007	1.9	0.5	0.0	9.7	1.6
2006-2008	1.7	0.3	0.0	9.0	1.4
2007-2009	1.5	0.3	0.0	8.3	1.3
2008-2010	1.8	0.3	0.0	8.7	1.5
2009-2011	2.6	1.0	0.0	13.8	2.3
2010-2012	2.2	0.9	0.0	12.5	2.0
2011-2013	1.8	0.6	0.0	9.7	1.6
2012-2014	1.1	0.3	0.0	7.3	1.0
2013-2015	1.3	0.3	0.0	8.3	1.2
2014-2016	1.5	0.5	0.0	9.3	1.4
2015-2017	1.7	0.3	0.0	11.1	1.5
2016-2018	2.3	0.8	0.0	14.2	2.1
2017-2019	3.6	1.4	0.1	18.7	3.2
2018-2020	3.4	1.6	0.1	17.6	3.0
2019-2021	2.2	0.9	0.1	12.8	2.0
Change from 2003-2020	0.7	0.4	0.1	4.5	0.7

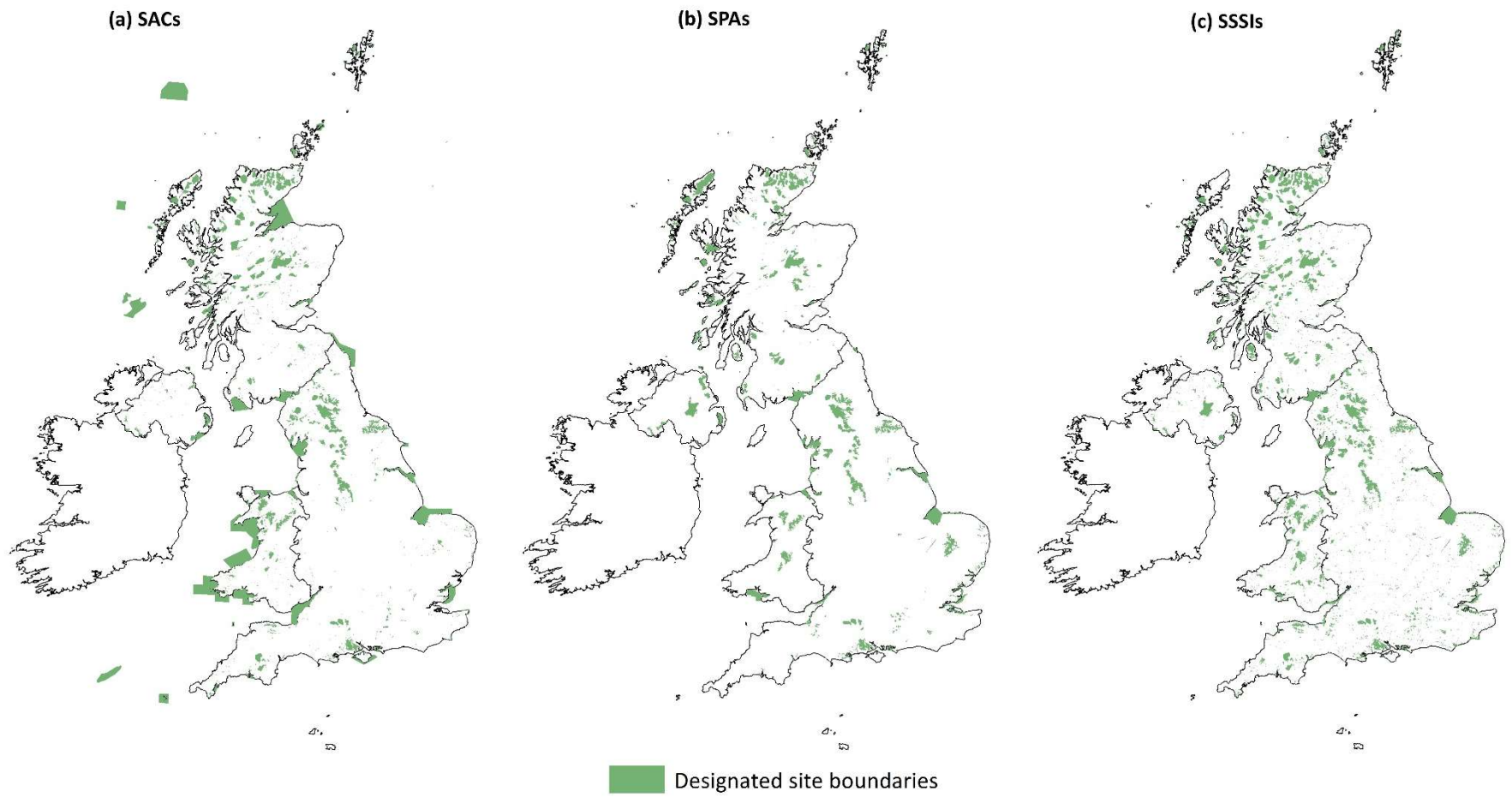


Figure 4.2: Distribution of SACs, SPAs and SSSIs / ASSIs in the UK.

Section 5: Nitrogen deposition onto protected sensitive habitats

An indicator of air pollution pressure, “total deposition of reactive N onto nutrient-N sensitive, protected, priority habitat” (abbreviated as N_{sens}) is included (Table 5.1) to illustrate progress towards the target for England in the UK Government’s Clean Air Strategy (Defra, 2019). See sections 1.1 and 1.4. This indicator has only been calculated for the most recent 5-year periods for which deposition data are available, i.e. periods centred on 2016, 2017, 2018, 2019 and 2020. Progress towards the target is assessed relative to a baseline year of 2016. As noted in Section 1.4, revisions of the underlying habitat maps are unlikely to affect the reported trend. Currently, this indicator is showing a 9.4 % decrease in England during the period 2016 to 2020 corresponding with the reduction in nitrogen deposition in 2020 (Figure 1.3).

The CAS target applies only to England, but N_{sens} was also calculated for the Devolved Administrations, for comparison. The Devolved Administrations are considering atmospheric nitrogen pollution indicators and targets for their own countries, and these will be presented in future Trends Reports.

Table 5.1: Total deposition of reactive N onto nutrient-N sensitive priority habitat (N_{sens} in $kg\ ha^{-1}\ year^{-1}$), by country.

Years	England	Wales	Scotland	NI	UK
2015-2017	22.7	20.8	10.4	16.9	14.6
2016-2018	22.6	19.9	9.6	18.2	14.0
2017-2019	22.8	19.3	9.5	19.1	14.0
2018-2020	22.6	19.5	9.3	18.9	13.9
2019-2021	20.6	18.2	8.5	17.0	12.7
Percent change from 2016-2020	-9.4	-12.1	-18.3	0.4	-13.4

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Annex: Critical load exceedances by habitat and country

This Annex contains summary critical load exceedance statistics for acidity and for nutrient nitrogen by habitat, with separate tables for each UK country.

Acidity results: Tables A1-A9

Nutrient nitrogen results: Tables N1-N9

Table A1: Acid-sensitive habitat areas in England and percentage area of habitats where acidity critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
<i>Habitat area (km²)</i>	4545	1000	2608	1782	366	3540	9588	1109	23429
2002-2004	90.8	15.0	88.8	97.9	99.9	84.6	66.9	49.3	77.3
2003-2005	90.5	13.9	88.2	97.6	99.9	84.1	66.6	49.5	76.9
2004-2006	90.8	14.5	88.0	97.8	100.0	83.8	65.7	49.3	76.6
2005-2007	90.3	13.1	87.0	97.5	99.9	83.7	65.5	49.3	76.2
2006-2008	89.7	12.8	86.4	97.5	99.9	82.7	63.9	49.1	75.2
2007-2009	88.8	11.9	82.9	97.2	99.9	80.6	61.3	48.9	73.2
2008-2010	88.7	10.1	82.9	97.2	99.9	79.9	60.4	47.8	72.6
2009-2011	89.0	11.3	83.3	97.3	99.9	81.0	61.1	48.8	73.2
2010-2012	89.1	11.3	84.8	97.3	99.9	81.2	60.8	47.7	73.3
2011-2013	89.7	12.4	84.8	97.5	99.9	81.5	61.5	48.8	73.8
2012-2014	89.3	11.3	83.5	97.4	99.9	80.3	59.7	48.8	72.7
2013-2015	89.4	11.3	82.1	97.5	99.9	80.1	58.9	48.6	72.1
2014-2016	89.1	10.0	80.0	97.3	99.9	78.9	57.4	48.4	71.0
2015-2017	87.9	7.2	74.4	97.2	99.9	76.8	55.1	46.3	68.8
2016-2018	87.1	5.5	75.2	97.1	99.9	76.9	56.6	45.8	69.3
2017-2019	87.0	6.1	74.9	97.1	99.9	77.1	57.4	45.4	69.6
2018-2020	87.1	4.6	78.1	97.1	99.9	78.4	57.0	45.6	69.9
2019-2021	86.8	2.1	69.6	97.1	99.9	74.5	52.7	44.3	66.5
Reduction 2003-2020	4	12.9	19.3	0.8	0.1	10.1	14.2	5	10.9

Table A2: Excess Acidity in England (Average Accumulated Exceedance in $\text{keq ha}^{-1} \text{ year}^{-1}$) by habitat and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
2002-2004	1.15	0.09	0.81	1.54	1.76	1.37	0.90	0.79	1.04
2003-2005	1.11	0.07	0.76	1.47	1.75	1.32	0.86	0.74	1.00
2004-2006	1.18	0.08	0.77	1.53	1.87	1.27	0.79	0.76	0.98
2005-2007	1.10	0.06	0.68	1.39	1.71	1.22	0.76	0.69	0.92
2006-2008	1.03	0.05	0.64	1.33	1.60	1.15	0.69	0.66	0.86
2007-2009	0.92	0.04	0.57	1.19	1.44	1.07	0.65	0.60	0.78
2008-2010	0.86	0.04	0.56	1.14	1.34	1.03	0.61	0.58	0.75
2009-2011	0.93	0.04	0.59	1.22	1.49	1.08	0.64	0.59	0.79
2010-2012	0.95	0.04	0.62	1.27	1.47	1.07	0.62	0.63	0.79
2011-2013	1.02	0.05	0.62	1.33	1.64	1.07	0.61	0.64	0.81
2012-2014	0.96	0.04	0.60	1.27	1.57	1.01	0.57	0.63	0.76
2013-2015	0.99	0.04	0.56	1.28	1.70	0.98	0.53	0.58	0.75
2014-2016	0.92	0.04	0.53	1.19	1.54	0.94	0.51	0.56	0.70
2015-2017	0.77	0.02	0.43	0.98	1.22	0.86	0.47	0.49	0.61
2016-2018	0.68	0.01	0.42	0.87	0.94	0.90	0.53	0.49	0.61
2017-2019	0.63	0.01	0.42	0.85	0.85	0.93	0.57	0.47	0.62
2018-2020	0.66	0.01	0.40	0.82	0.85	0.98	0.55	0.42	0.62
2019-2021	0.55	0.00	0.28	0.65	0.73	0.81	0.43	0.33	0.50
Reduction 2003-2020	0.6	0.09	0.54	0.89	1.03	0.56	0.47	0.46	0.54

Table A3: Acid-sensitive habitat areas in Wales and percentage area of habitats where acidity critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
<i>Habitat area (km²)</i>	<i>4444</i>	<i>1</i>	<i>677</i>	<i>193</i>	<i>17</i>	<i>1590</i>	<i>2152</i>	<i>1225</i>	<i>9075</i>
2002-2004	93.5	41.0	87.0	99.1	100.0	79.5	64.3	48.4	83.8
2003-2005	93.1	28.3	86.5	99.0	100.0	79.2	64.1	50.3	83.4
2004-2006	93.5	39.6	87.3	99.1	100.0	80.4	66.9	53.9	84.5
2005-2007	93.5	39.6	87.4	99.1	100.0	80.7	69.5	53.9	85.2
2006-2008	93.0	28.3	86.5	99.1	100.0	80.1	67.7	52.0	84.4
2007-2009	92.6	24.3	84.9	99.0	100.0	72.6	54.0	45.5	79.5
2008-2010	92.5	24.3	83.0	99.0	100.0	68.9	47.0	40.9	77.0
2009-2011	92.5	24.3	82.0	99.0	100.0	69.3	50.5	34.2	77.8
2010-2012	92.5	24.3	83.7	99.0	100.0	69.5	49.8	34.2	77.8
2011-2013	92.6	24.3	84.6	99.0	100.0	72.4	54.2	45.6	79.5
2012-2014	92.6	25.9	84.5	99.0	100.0	69.2	46.8	44.9	77.2
2013-2015	92.6	25.9	83.9	99.0	100.0	70.3	47.4	45.9	77.5
2014-2016	92.6	25.9	84.1	99.0	100.0	70.5	46.4	45.8	77.3
2015-2017	92.5	3.6	82.3	99.0	100.0	66.6	41.0	41.8	75.1
2016-2018	92.5	24.3	79.4	99.0	100.0	63.3	41.4	29.2	74.4
2017-2019	91.8	24.3	76.4	98.7	100.0	62.5	42.8	26.2	74.1
2018-2020	90.0	41.8	74.3	97.4	100.0	62.7	45.3	23.7	73.6
2019-2021	88.3	34.1	67.5	84.0	100.0	59.7	39.7	22.9	70.1
Reduction 2003-2020	5.2	7	19.5	15.1	0	19.8	24.6	25.6	13.6

Table A4: Excess Acidity in Wales (Average Accumulated Exceedance in keq ha⁻¹ year⁻¹) by habitat and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
2002-2004	0.91	0.09	0.63	0.89	1.59	1.03	0.54	0.46	0.82
2003-2005	0.86	0.05	0.59	0.88	1.53	0.99	0.52	0.45	0.78
2004-2006	0.98	0.07	0.70	1.02	1.74	1.05	0.53	0.54	0.87
2005-2007	0.97	0.06	0.69	1.01	1.69	1.06	0.56	0.54	0.87
2006-2008	0.93	0.04	0.64	0.94	1.80	1.01	0.51	0.52	0.82
2007-2009	0.74	0.02	0.47	0.72	1.55	0.80	0.40	0.36	0.65
2008-2010	0.67	0.01	0.41	0.65	1.47	0.73	0.36	0.31	0.59
2009-2011	0.66	0.00	0.37	0.63	1.36	0.73	0.38	0.30	0.58
2010-2012	0.70	0.01	0.41	0.68	1.41	0.74	0.38	0.33	0.61
2011-2013	0.78	0.03	0.49	0.75	1.54	0.81	0.41	0.37	0.67
2012-2014	0.77	0.04	0.49	0.72	1.52	0.76	0.37	0.34	0.65
2013-2015	0.80	0.03	0.52	0.78	1.59	0.79	0.37	0.38	0.67
2014-2016	0.81	0.02	0.52	0.79	1.60	0.78	0.36	0.40	0.68
2015-2017	0.69	0.00	0.41	0.67	1.39	0.68	0.31	0.33	0.58
2016-2018	0.59	0.01	0.34	0.52	1.13	0.62	0.32	0.23	0.51
2017-2019	0.53	0.04	0.29	0.41	0.93	0.57	0.32	0.16	0.46
2018-2020	0.53	0.10	0.28	0.37	0.90	0.56	0.33	0.14	0.47
2019-2021	0.45	0.05	0.19	0.30	0.85	0.46	0.26	0.13	0.38
Reduction 2003-2020	0.46	0.04	0.45	0.59	0.74	0.57	0.28	0.33	0.44

Table A5: Acid-sensitive habitat areas in Scotland and percentage area of habitats where acidity critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
<i>Habitat area (km²)</i>	<i>10928</i>	<i>0</i>	<i>17604</i>	<i>6295</i>	<i>4526</i>	<i>10314</i>	<i>4847</i>	<i>5337</i>	<i>54512</i>
2002-2004	80.1	0.0	22.4	36.9	91.9	83.9	63.2	14.2	56.7
2003-2005	79.3	0.0	19.9	33.5	91.4	81.1	59.0	14.2	54.4
2004-2006	84.7	0.0	30.9	56.4	97.9	86.8	64.5	27.4	63.8
2005-2007	83.2	0.0	27.2	65.5	96.6	86.2	63.6	15.1	63.0
2006-2008	80.9	0.0	19.9	59.5	96.0	85.8	62.7	14.4	59.3
2007-2009	67.0	0.0	9.9	38.3	77.6	77.8	53.6	12.5	47.0
2008-2010	67.1	0.0	11.1	33.7	78.4	77.7	53.5	12.6	46.9
2009-2011	72.7	0.0	14.5	33.5	81.7	80.3	57.9	13.0	50.2
2010-2012	69.1	0.0	14.0	32.8	78.7	79.2	55.6	12.9	48.6
2011-2013	74.9	0.0	14.0	30.8	86.1	82.5	59.2	13.1	51.1
2012-2014	73.3	0.0	14.5	31.2	84.0	80.7	57.0	13.1	50.3
2013-2015	77.7	0.0	16.9	32.5	91.6	82.9	59.3	13.9	53.3
2014-2016	75.6	0.0	14.8	28.2	88.7	78.1	55.4	13.6	50.3
2015-2017	67.5	0.0	7.8	22.6	78.8	70.2	48.5	11.8	42.8
2016-2018	57.0	0.0	5.7	20.2	53.0	63.0	40.8	8.1	35.6
2017-2019	54.3	0.0	5.5	20.0	47.2	63.6	41.6	8.1	34.6
2018-2020	55.0	0.0	4.5	17.1	44.5	63.0	40.7	8.1	33.7
2019-2021	47.6	0.0	1.8	10.8	41.1	57.1	37.3	7.3	28.9
Reduction 2003-2020	32.5	0	20.6	26.2	50.8	26.8	25.9	6.9	27.8

Table A6: Excess Acidity in Scotland (Average Accumulated Exceedance in keq ha⁻¹ year⁻¹) by habitat and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
2002-2004	0.49	0.00	0.06	0.11	0.60	0.66	0.39	0.10	0.34
2003-2005	0.47	0.00	0.04	0.11	0.54	0.62	0.36	0.10	0.31
2004-2006	0.62	0.00	0.09	0.17	0.76	0.73	0.45	0.14	0.41
2005-2007	0.52	0.00	0.06	0.19	0.61	0.69	0.41	0.12	0.37
2006-2008	0.45	0.00	0.04	0.16	0.53	0.65	0.38	0.10	0.32
2007-2009	0.28	0.00	0.02	0.09	0.31	0.49	0.26	0.08	0.22
2008-2010	0.29	0.00	0.03	0.08	0.35	0.50	0.27	0.07	0.22
2009-2011	0.36	0.00	0.04	0.09	0.42	0.58	0.33	0.09	0.27
2010-2012	0.36	0.00	0.04	0.09	0.39	0.57	0.31	0.08	0.26
2011-2013	0.40	0.00	0.03	0.10	0.42	0.63	0.34	0.10	0.29
2012-2014	0.38	0.00	0.03	0.09	0.41	0.58	0.31	0.09	0.27
2013-2015	0.45	0.00	0.04	0.10	0.51	0.61	0.34	0.10	0.30
2014-2016	0.40	0.00	0.03	0.08	0.46	0.53	0.29	0.09	0.26
2015-2017	0.28	0.00	0.01	0.06	0.30	0.41	0.22	0.06	0.19
2016-2018	0.21	0.00	0.01	0.05	0.17	0.36	0.19	0.05	0.15
2017-2019	0.18	0.00	0.01	0.05	0.14	0.37	0.20	0.05	0.14
2018-2020	0.17	0.00	0.01	0.04	0.12	0.35	0.18	0.05	0.13
2019-2021	0.10	0.00	0.00	0.03	0.09	0.27	0.14	0.03	0.10
Reduction 2003-2020	0.38	0	0.06	0.09	0.51	0.39	0.25	0.07	0.24

Table A7: Acid-sensitive habitat areas in Northern Ireland and percentage area of habitats where acidity critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
<i>Habitat area (km²)</i>	448	10	957	848	6	490	768	186	3527
2002-2004	89.5	6.7	61.2	88.3	93.3	93.3	77.9	18.7	79.3
2003-2005	88.5	6.6	57.2	87.6	93.3	93.1	77.4	16.2	77.8
2004-2006	89.5	6.7	61.6	88.4	93.3	93.2	77.5	20.3	79.3
2005-2007	89.6	6.6	61.6	88.7	93.3	93.4	78.3	18.7	79.6
2006-2008	89.3	6.6	62.1	88.8	93.3	93.4	78.2	20.0	79.7
2007-2009	88.6	6.3	59.6	87.7	93.3	93.2	77.5	18.7	78.5
2008-2010	89.2	6.6	61.3	88.1	93.3	93.2	78.0	20.3	79.3
2009-2011	90.4	8.0	63.7	89.5	93.3	93.5	78.8	20.3	80.6
2010-2012	90.1	7.1	63.2	89.1	93.3	93.5	78.6	20.3	80.3
2011-2013	91.1	7.9	66.1	90.2	93.3	93.6	78.8	20.0	81.5
2012-2014	90.2	6.7	64.2	89.4	93.3	93.3	78.2	20.0	80.5
2013-2015	89.3	6.6	60.0	89.0	93.3	93.2	77.8	18.4	79.1
2014-2016	87.9	2.6	54.9	87.9	93.3	93.0	77.1	18.4	77.1
2015-2017	85.0	2.2	49.6	85.7	93.3	92.8	76.8	16.0	74.6
2016-2018	87.1	2.9	55.6	87.0	93.3	93.1	77.8	16.0	77.1
2017-2019	87.5	3.1	58.4	86.6	92.5	93.5	78.7	16.0	78.1
2018-2020	87.2	3.1	57.0	85.8	90.7	93.5	78.8	16.3	77.5
2019-2021	82.5	2.9	47.9	80.5	87.3	93.0	77.5	16.0	72.8
Reduction 2003-2020	7	3.8	13.4	7.8	6	0.3	0.3	2.7	6.5

Table A8: Excess Acidity in Northern Ireland (Average Accumulated Exceedance in keq ha⁻¹ year⁻¹) by habitat and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland	Broadleaved and mixed woodland (unmanaged)	Freshwaters	All habitats
2002-2004	0.55	0.01	0.28	0.53	1.06	1.18	1.15	0.18	0.69
2003-2005	0.49	0.01	0.24	0.48	0.96	1.10	1.08	0.16	0.63
2004-2006	0.56	0.01	0.28	0.53	1.03	1.16	1.12	0.17	0.68
2005-2007	0.57	0.01	0.29	0.54	1.08	1.19	1.17	0.18	0.70
2006-2008	0.54	0.01	0.26	0.53	0.98	1.22	1.17	0.16	0.69
2007-2009	0.50	0.01	0.25	0.49	1.02	1.14	1.12	0.17	0.65
2008-2010	0.53	0.01	0.27	0.51	1.02	1.17	1.14	0.17	0.68
2009-2011	0.60	0.01	0.30	0.59	1.02	1.27	1.24	0.18	0.75
2010-2012	0.58	0.01	0.29	0.57	1.00	1.23	1.20	0.17	0.72
2011-2013	0.62	0.01	0.29	0.62	0.97	1.31	1.23	0.16	0.76
2012-2014	0.58	0.01	0.27	0.57	0.95	1.23	1.14	0.16	0.71
2013-2015	0.51	0.01	0.20	0.52	0.76	1.17	1.11	0.13	0.65
2014-2016	0.43	0.01	0.17	0.46	0.69	1.07	1.04	0.11	0.59
2015-2017	0.36	0.00	0.15	0.39	0.67	0.97	1.00	0.11	0.53
2016-2018	0.45	0.01	0.22	0.45	0.83	1.06	1.10	0.13	0.61
2017-2019	0.49	0.01	0.27	0.49	0.96	1.16	1.22	0.16	0.68
2018-2020	0.49	0.01	0.27	0.48	1.01	1.14	1.21	0.17	0.68
2019-2021	0.35	0.01	0.18	0.36	0.79	0.98	1.06	0.13	0.55
Reduction 2003-2020	0.2	0.01	0.11	0.17	0.27	0.2	0.09	0.05	0.14

Table N1: Nutrient-sensitive habitat area in England and percentage area of habitats where nutrient-nitrogen critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
<i>Habitat area (km²)</i>	4545	1000	2608	1782	366	3540	5950	1753	3222	0	538	199	535	26038
2002-2004	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.4	100.0
2003-2005	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.6	100.0
2004-2006	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.1	100.0
2005-2007	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.2	100.0
2006-2008	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	96.8	99.9
2007-2009	100.0	99.6	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	97.9	99.9
2008-2010	100.0	99.6	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.9	100.0
2009-2011	100.0	99.7	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.3	100.0
2010-2012	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.3	100.0
2011-2013	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.3	100.0
2012-2014	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.6	100.0
2013-2015	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.6	100.0
2014-2016	100.0	99.7	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.5	100.0
2015-2017	100.0	99.7	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.0	99.9
2016-2018	100.0	99.7	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.6	100.0
2017-2019	100.0	99.8	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.9	100.0
2018-2020	100.0	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.0	100.0
2019-2021	100.0	99.3	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	96.3	99.9
Reduction 2003-2020	0	0.6	0	0	0	0	0	0	0	NA	0	0	3	0.1

Table N2: Excess Nitrogen in England: Average Accumulated Exceedance for nutrient-Nitrogen (in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
2002-2004	13.52	10.73	15.94	19.54	21.79	21.94	23.54	21.47	21.99	NA	22.88	8.35	6.11	19.20
2003-2005	13.41	10.09	15.63	19.28	22.12	21.64	23.30	21.13	21.60	NA	22.58	8.08	5.88	18.94
2004-2006	14.72	10.34	16.11	20.45	24.22	21.23	22.27	20.05	20.80	NA	21.78	7.95	5.73	18.84
2005-2007	14.25	10.04	15.53	19.45	23.33	21.11	22.24	20.14	20.87	NA	21.81	8.08	5.62	18.61
2006-2008	13.97	9.88	15.44	19.39	23.04	20.48	21.31	19.51	20.23	NA	20.89	8.04	5.41	18.09
2007-2009	12.85	9.58	14.96	18.33	21.56	19.73	20.94	19.14	19.66	NA	20.37	7.72	5.16	17.44
2008-2010	12.47	9.50	14.98	17.97	20.67	19.46	20.53	18.72	19.28	NA	20.05	7.79	5.25	17.13
2009-2011	13.51	9.90	15.49	18.88	22.40	20.35	21.18	19.32	20.08	NA	20.84	8.34	5.75	17.90
2010-2012	13.49	9.72	15.52	19.04	21.98	19.92	20.63	18.71	19.52	NA	20.26	8.38	5.84	17.60
2011-2013	14.32	9.99	15.65	19.81	23.89	19.93	20.39	18.73	19.54	NA	20.15	8.31	5.71	17.79
2012-2014	13.72	9.94	15.45	19.34	23.11	19.04	19.61	18.02	18.63	NA	19.37	7.78	5.25	17.13
2013-2015	14.56	10.06	15.64	20.15	25.13	19.17	19.22	17.83	18.61	NA	19.23	7.91	5.07	17.29
2014-2016	14.05	9.93	15.46	19.44	23.86	18.78	19.16	17.58	18.32	NA	19.11	7.93	5.12	16.99
2015-2017	12.47	9.19	14.45	17.40	20.64	17.94	18.63	16.98	17.62	NA	18.44	7.78	4.90	16.02
2016-2018	11.35	9.43	14.51	16.40	17.70	18.72	20.07	18.11	18.43	NA	19.64	7.99	5.40	16.38
2017-2019	11.01	10.02	14.64	16.40	16.91	19.42	21.38	18.84	19.17	NA	20.75	8.22	5.91	16.91
2018-2020	11.38	9.40	14.41	15.92	16.83	19.98	21.10	18.50	18.85	NA	20.57	8.16	5.62	16.83
2019-2021	9.73	7.59	12.43	13.74	15.39	17.14	18.20	15.87	16.35	NA	17.79	7.14	4.38	14.48
Reduction 2003-2020	3.79	3.14	3.51	5.79	6.39	4.8	5.33	5.6	5.65	NA	5.09	1.21	1.73	4.72

Table N3: Nutrient-sensitive habitat area in Wales and percentage area of habitats where nutrient-nitrogen critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
<i>Habitat area (km²)</i>	4444	1	677	193	17	1590	281	279	1567	0	125	106	131	9412
2002-2004	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	92.4	99.9
2003-2005	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	93.2	99.9
2004-2006	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	95.2	99.9
2005-2007	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	97.7	100.0
2006-2008	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	99.2	100.0
2007-2009	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	93.6	99.9
2008-2010	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	86.8	99.8
2009-2011	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	93.6	99.9
2010-2012	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	94.9	99.9
2011-2013	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	98.6	100.0
2012-2014	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	95.4	99.9
2013-2015	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	95.5	99.9
2014-2016	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	97.4	100.0
2015-2017	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	93.5	99.9
2016-2018	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	88.0	99.8
2017-2019	99.9	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	77.3	99.6
2018-2020	99.3	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	74.9	99.3
2019-2021	99.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	100.0	68.6	99.1
Reduction 2003-2020	1	0	0	0	0	0	0	0	0	NA	0	0	23.8	0.8

Table N4: Excess Nitrogen in Wales: Average Accumulated Exceedance for nutrient-Nitrogen (in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
2002-2004	10.83	8.04	14.38	14.48	18.42	19.87	17.46	18.30	17.73	NA	18.27	5.76	3.64	14.21
2003-2005	10.41	7.18	14.11	14.58	17.78	19.66	17.03	18.01	17.66	NA	17.99	5.63	3.51	13.92
2004-2006	12.19	7.92	15.86	16.60	20.97	20.79	16.87	18.33	18.28	NA	18.43	5.99	3.96	15.25
2005-2007	12.37	8.06	16.03	16.89	20.88	21.51	17.60	18.90	19.18	NA	19.17	6.46	4.37	15.68
2006-2008	12.50	7.89	16.14	16.82	23.54	21.46	17.10	19.03	19.10	NA	18.99	6.72	4.36	15.72
2007-2009	10.46	7.02	14.16	14.57	20.91	18.74	15.62	17.31	16.98	NA	17.08	5.88	3.41	13.61
2008-2010	9.93	6.83	13.69	13.94	20.36	17.65	15.02	16.55	15.99	NA	16.33	5.54	3.09	12.90
2009-2011	9.86	6.78	13.36	13.81	19.16	17.99	15.72	17.17	16.61	NA	16.85	5.78	3.43	13.06
2010-2012	10.15	6.82	13.65	14.21	19.58	17.92	15.39	16.74	16.31	NA	16.53	5.95	3.50	13.14
2011-2013	10.96	7.25	14.51	14.93	20.93	18.98	15.88	17.45	17.10	NA	17.20	6.31	3.76	13.96
2012-2014	10.85	7.65	14.55	14.61	20.74	17.96	14.95	16.56	15.93	NA	16.26	5.84	3.24	13.46
2013-2015	11.78	7.80	15.46	15.82	22.25	19.02	15.32	17.07	16.67	NA	16.89	5.99	3.48	14.33
2014-2016	12.18	7.86	15.81	16.29	22.70	19.25	15.35	17.11	16.77	NA	17.00	6.11	3.71	14.62
2015-2017	11.00	6.86	14.68	15.22	20.80	17.75	14.38	15.88	15.67	NA	15.86	5.75	3.38	13.43
2016-2018	9.91	7.21	13.80	13.53	17.78	16.57	14.78	15.96	15.19	NA	15.68	5.59	3.32	12.54
2017-2019	9.24	8.46	13.10	12.42	15.65	15.61	15.18	16.15	14.99	NA	15.69	5.41	3.40	11.97
2018-2020	9.52	11.05	13.05	12.09	15.50	15.41	15.68	16.45	15.11	NA	15.98	5.55	3.52	12.11
2019-2021	8.36	9.51	11.63	11.15	14.93	13.65	13.61	14.36	13.53	NA	14.10	5.06	2.84	10.72
Reduction 2003-2020	2.47	-1.47	2.74	3.33	3.49	6.21	3.84	3.94	4.2	NA	4.18	0.7	0.8	3.49

Table N5: Nutrient-sensitive habitat area in Scotland and percentage area of habitats where nutrient-nitrogen critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
<i>Habitat area (km²)</i>	10928	0	17604	6295	4526	8829	1867	28	2038	1485	336	258	141	54333
2002-2004	84.5	0.0	90.9	89.9	100.0	85.9	87.9	100.0	87.4	100.0	91.1	79.8	55.2	89.3
2003-2005	85.2	0.0	91.6	88.0	100.0	85.0	87.5	100.0	86.0	100.0	90.0	78.5	52.1	89.2
2004-2006	94.8	0.0	98.8	99.0	100.0	94.9	93.2	100.0	95.3	100.0	95.9	86.9	61.8	97.0
2005-2007	94.7	0.0	99.1	99.2	100.0	97.9	96.0	100.0	96.2	100.0	97.1	87.4	59.7	97.7
2006-2008	92.8	0.0	98.4	99.0	100.0	99.2	97.8	100.0	98.2	100.0	98.2	87.4	55.4	97.5
2007-2009	79.9	0.0	92.4	90.1	100.0	91.7	89.5	100.0	87.7	100.0	92.6	79.6	45.0	89.9
2008-2010	79.2	0.0	89.0	87.6	100.0	88.6	88.0	100.0	87.3	100.0	91.2	78.6	49.9	87.8
2009-2011	81.5	0.0	91.1	90.3	100.0	89.9	89.3	100.0	89.4	100.0	92.9	82.4	54.8	89.6
2010-2012	76.1	0.0	88.6	88.5	100.0	88.0	87.4	100.0	87.0	100.0	89.9	80.9	53.7	87.0
2011-2013	82.6	0.0	92.7	90.7	100.0	94.4	92.7	100.0	94.0	100.0	96.0	82.8	54.0	91.4
2012-2014	81.5	0.0	90.9	89.9	100.0	88.6	89.3	100.0	89.2	100.0	92.8	82.1	52.4	89.3
2013-2015	88.1	0.0	92.1	88.6	100.0	91.3	91.9	100.0	94.0	100.0	95.5	82.7	56.8	91.6
2014-2016	87.2	0.0	89.0	85.3	100.0	86.5	88.2	100.0	87.8	100.0	91.2	79.5	54.7	88.8
2015-2017	83.4	0.0	87.5	77.6	100.0	84.3	86.4	100.0	82.4	100.0	87.1	74.0	50.0	86.0
2016-2018	74.5	0.0	83.4	72.1	98.8	80.9	83.9	100.0	75.3	100.0	79.0	79.5	49.0	81.2
2017-2019	71.8	0.0	83.1	73.0	96.3	82.1	85.1	100.0	76.6	100.0	79.2	84.1	51.8	80.8
2018-2020	73.0	0.0	83.1	65.1	95.7	81.9	85.4	100.0	76.5	100.0	78.8	83.9	49.7	80.0
2019-2021	65.7	0.0	79.8	53.4	97.6	80.4	84.4	100.0	72.7	100.0	76.8	74.9	44.6	75.8
Reduction 2003-2020	18.8	0	11.1	36.5	2.4	5.5	3.4	0	14.8	0	14.2	4.9	10.6	13.5

Table N6: Excess Nitrogen in Scotland: Average Accumulated Exceedance for nutrient-Nitrogen (in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
2002-2004	5.20	0.00	3.90	3.00	8.20	8.80	9.50	14.50	6.50	9.40	8.70	2.70	1.80	5.70
2003-2005	5.10	0.00	3.80	3.00	7.90	8.50	8.80	14.20	6.10	9.10	8.10	2.40	1.60	5.50
2004-2006	7.20	0.00	5.80	4.70	10.90	10.00	9.80	14.80	7.70	11.40	9.40	3.00	2.30	7.40
2005-2007	6.60	0.00	6.00	5.50	9.80	10.10	9.90	15.50	7.50	11.10	9.40	3.30	2.30	7.40
2006-2008	6.20	0.00	5.40	5.20	9.60	10.20	10.10	16.20	7.30	11.30	9.50	3.10	2.20	7.10
2007-2009	4.10	0.00	3.60	3.60	6.70	8.20	8.50	15.00	5.40	9.00	7.80	2.80	1.60	5.10
2008-2010	4.30	0.00	3.60	3.20	7.10	8.30	9.00	14.80	5.60	9.00	8.20	3.00	1.60	5.20
2009-2011	5.20	0.00	4.00	3.30	7.70	9.40	10.00	16.10	6.50	9.60	9.10	3.40	2.10	5.80
2010-2012	5.10	0.00	3.70	3.20	7.20	9.10	9.90	16.00	6.10	9.10	8.90	3.30	2.10	5.60
2011-2013	5.60	0.00	4.00	3.40	7.80	10.00	10.00	16.90	6.90	10.10	9.30	3.10	2.40	6.10
2012-2014	5.40	0.00	3.90	3.20	7.70	9.40	9.20	15.80	6.30	9.60	8.50	3.00	2.10	5.80
2013-2015	6.60	0.00	4.40	3.40	9.30	10.00	9.70	16.10	7.10	10.50	9.10	2.90	2.40	6.50
2014-2016	6.20	0.00	4.20	3.00	8.90	9.00	8.90	15.60	6.30	9.50	8.30	2.60	2.30	6.00
2015-2017	4.90	0.00	3.50	2.60	7.20	7.80	8.00	14.70	5.20	8.20	7.30	2.40	1.90	5.10
2016-2018	4.00	0.00	3.00	2.30	5.10	7.20	8.00	14.80	4.60	6.70	7.00	2.60	1.70	4.30
2017-2019	3.80	0.00	3.00	2.30	4.50	7.40	8.70	15.80	5.00	6.80	7.70	3.10	1.90	4.30
2018-2020	3.50	0.00	3.10	2.30	4.10	7.20	8.60	15.40	4.70	6.50	7.40	2.90	1.70	4.20
2019-2021	2.50	0.00	2.40	1.70	3.70	5.70	7.20	13.90	3.60	6.00	6.30	2.10	1.20	3.30
Reduction 2003-2020	2.71	0	1.55	1.29	4.54	3.12	2.28	0.68	2.93	3.39	2.42	0.58	0.57	2.38

Table N7: Nutrient-sensitive habitat area in Northern Ireland and percentage area of habitats where nutrient-nitrogen critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
<i>Habitat area (km²)</i>	448	10	957	848	6	490	608	0	131	0	424	68	0	3990
2002-2004	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2003-2005	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2004-2006	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2005-2007	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2006-2008	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2007-2009	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2008-2010	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2009-2011	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2010-2012	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2011-2013	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2012-2014	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2013-2015	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2014-2016	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2015-2017	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
2016-2018	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2017-2019	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2018-2020	100.0	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	100.0	100.0
2019-2021	99.9	100.0	100.0	100.0	100.0	100.0	100.0	NA	100.0	NA	100.0	100.0	61.7	100.0
Reduction 2003-2020	0.1	0	0	0	0	0	0	NA	0	NA	0	0	0	0

Table N8: Excess Nitrogen in Northern Ireland: Average Accumulated Exceedance for nutrient-Nitrogen (in kg N ha⁻¹ year⁻¹) by country and deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh	All
2002-2004	7.78	6.04	12.50	11.60	16.99	15.71	20.64	NA	15.91	NA	17.84	7.72	0.22	14.00
2003-2005	7.26	5.38	11.92	11.10	16.17	14.95	19.85	NA	15.21	NA	17.10	7.27	0.08	13.37
2004-2006	8.02	6.33	12.61	11.89	17.05	15.76	20.34	NA	15.81	NA	17.76	7.69	0.26	14.07
2005-2007	8.58	6.65	13.19	12.37	17.91	16.82	21.83	NA	17.04	NA	19.01	8.20	0.55	14.91
2006-2008	8.37	7.01	12.98	12.51	16.62	17.49	21.85	NA	17.39	NA	19.36	7.96	0.59	14.99
2007-2009	8.44	6.93	13.06	12.38	17.47	17.00	21.45	NA	16.99	NA	18.94	8.10	0.53	14.82
2008-2010	8.77	7.27	13.35	12.77	17.56	17.42	21.86	NA	17.45	NA	19.40	8.16	0.73	15.19
2009-2011	9.74	7.93	14.09	13.78	17.65	18.82	23.53	NA	18.94	NA	20.93	8.61	1.31	16.33
2010-2012	9.26	7.14	13.69	13.31	17.32	18.10	22.70	NA	18.27	NA	20.19	8.30	1.19	15.76
2011-2013	10.01	8.08	14.07	14.15	17.30	19.39	22.89	NA	19.02	NA	20.93	8.48	2.00	16.41
2012-2014	9.56	7.48	13.65	13.53	17.17	18.18	21.38	NA	17.72	NA	19.56	8.33	1.75	15.56
2013-2015	8.72	7.30	12.83	12.98	15.14	17.59	21.22	NA	17.42	NA	19.17	7.78	1.41	14.99
2014-2016	7.74	6.35	12.08	12.10	14.22	16.13	20.34	NA	16.34	NA	18.01	7.37	0.88	14.03
2015-2017	6.99	5.63	11.68	11.27	14.29	15.05	20.06	NA	15.65	NA	17.30	7.24	0.46	13.40
2016-2018	8.40	6.37	13.13	12.35	16.40	16.52	22.04	NA	17.12	NA	18.97	8.50	0.98	14.87
2017-2019	9.18	6.99	14.18	12.96	18.18	18.08	24.64	NA	18.94	NA	21.00	9.54	1.34	16.22
2018-2020	9.04	6.80	14.06	12.76	18.79	17.87	24.51	NA	18.85	NA	20.89	9.58	1.34	16.07
2019-2021	6.88	5.36	11.97	10.99	15.80	15.46	21.56	NA	16.52	NA	18.32	7.94	0.41	13.83
Reduction 2003-2020	0.89	0.67	0.53	0.6	1.19	0.25	-0.93	NA	-0.6	NA	-0.48	-0.23	-0.19	0.17

Table N9: Nutrient-sensitive habitat area in the UK and total area of habitats where nutrient-nitrogen critical load is exceeded, by deposition dataset year.

Years	Acid grassland	Calcareous grassland	Dwarf shrub heath	Bog	Montane	Coniferous woodland (managed)	Other broadleaved woodland	Beech woodland	Acidophilous oak woodland	Scots Pine woodland	Mixed woodland	Dune grassland	Saltmarsh
<i>Habitat area (km²)</i>	<i>20365</i>	<i>1012</i>	<i>21846</i>	<i>9118</i>	<i>4915</i>	<i>14450</i>	<i>8706</i>	<i>2059</i>	<i>6958</i>	<i>1485</i>	<i>1422</i>	<i>631</i>	<i>808</i>
2002-2004	18,669	1,011	20,243	8,484	4,914	13,206	8,479	2,059	6,702	1,485	1,392	579	731
2003-2005	18,752	1,011	20,362	8,361	4,914	13,124	8,472	2,059	6,672	1,485	1,389	575	724
2004-2006	19,795	1,011	21,639	9,053	4,915	14,000	8,579	2,059	6,863	1,485	1,408	597	738
2005-2007	19,785	1,011	21,686	9,065	4,915	14,260	8,631	2,059	6,880	1,485	1,413	598	738
2006-2008	19,576	1,011	21,559	9,055	4,915	14,383	8,664	2,059	6,921	1,485	1,416	598	727
2007-2009	18,167	1,008	20,511	8,494	4,915	13,718	8,509	2,059	6,708	1,485	1,397	578	710
2008-2010	18,089	1,007	19,913	8,339	4,913	13,440	8,481	2,059	6,700	1,485	1,393	576	714
2009-2011	18,338	1,009	20,276	8,509	4,914	13,559	8,505	2,059	6,742	1,485	1,398	585	732
2010-2012	17,748	1,011	19,833	8,391	4,914	13,387	8,471	2,059	6,694	1,485	1,388	582	732
2011-2013	18,460	1,011	20,558	8,532	4,914	13,956	8,570	2,059	6,836	1,485	1,409	586	737
2012-2014	18,341	1,011	20,240	8,481	4,914	13,446	8,505	2,059	6,739	1,485	1,398	585	727
2013-2015	19,063	1,011	20,452	8,402	4,914	13,685	8,555	2,059	6,835	1,485	1,407	586	734
2014-2016	18,966	1,009	19,910	8,190	4,914	13,262	8,486	2,059	6,709	1,485	1,393	578	732
2015-2017	18,547	1,009	19,643	7,706	4,914	13,067	8,452	2,059	6,600	1,485	1,379	564	718
2016-2018	17,583	1,009	18,926	7,362	4,860	12,767	8,406	2,059	6,455	1,485	1,352	578	712
2017-2019	17,275	1,009	18,879	7,417	4,748	12,866	8,427	2,059	6,482	1,485	1,353	590	704
2018-2020	17,383	1,011	18,879	6,919	4,721	12,850	8,433	2,059	6,478	1,485	1,351	589	698
2019-2021	16,573	1,005	18,295	6,184	4,807	12,723	8,415	2,059	6,401	1,485	1,344	566	669
Reduction 2003-2020	2096	6	1948	2300	107	483	64	0	301	0	48	13	62

